EVALUATION OF OPERATIONAL PERFORMANCE AND ENVIRONMENTAL IMPACT OF A COMMERCIAL SCALE ANAEROBIC DIGESTER UTILIZING MULTIPLE FEEDSTOCKS

By

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ABSTRACT

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Food waste and livestock manure become some the of major sources that contribute to greenhouse gas (GHG) emissions in the U.S. Utilizing manure and food wastes as biogas feedstocks through the anaerobic digestion (AD) process can improve renewable energy production while reducing the impact of climate change due to GHG emission from untreated organic wastes. This study evaluated the operational performance of Michigan State University's commercial South Campus Anaerobic Digester (SCAD) as well as the environmental impact during its operation from 2014 to 2020. Evaluation of feedstock supplies quantity and output parameters of SCAD was conducted to understand the operational performance of the digester. A life cycle assessment (LCA) was done to know the environmental impact of SCAD by comparing it to the conventional waste management methods. Technoeconomic analysis was conducted to know the financial feasibility of SCAD as a commercial digester. The result shows that during its operation from 2014 to 2020, SCAD has processed 159,145 metric tons of feedstock from 18 different organic wastes to produce 15,165,156 kWh of electricity for the MSU community. LCA results show that the AD system possesses fewer environmental burdens in both global warming potential (GWP) and water eutrophication potential (WEP) compared to the conventional system. Technoeconomic analysis reveals that SCAD needs 21.5 years to accomplish its payback period, which is considered quite economically competitive. Economic sensitivity analysis shows that electricity becomes the most sensitive parameter to affect the payback period.

This thesis is dedicated to my mother, Mrs. Fatmah, and my father, Mr. Muhammad Ali Ismail. Thank you for always believing in me.

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KEY TO ABBREVIATIONS

AD	Anaerobic Digester
ANS	Animal Science
ADREC	Anaerobic Digestion Research and Education Center
CapEX	Capital Expenditure
CH ₄	Methane
CO ₂	Carbon Dioxide
COD	Chemical Oxygen Demand
CSTR	Continuous Stirred Tank Reactor
DQI	Data Quality Inventory
EC	Electrical Conductivity
EPA	Environmental Protection Agency
FOG	Fat, Oil, and Grease
FU	Functional Unit
GWP	Global Warming Potential
H_2S	Hydrogen Sulfide
LCA	Life Cycle Analysis
LCFS	Low Carbon Fuel Standard
LCI	Life Cycle Inventory
LCIA	Life Cycle Impact Assessment
MLR	Multi Linear Regression
MSU	Michigan State University

N ₂ O	Nitrous Oxide	
NH3	Ammonia	
OLR	Organic Loading Rate	
OpEX	Operational Expenditure	
REC	Renewable Energy Credit	
RIN	Renewable Identification Number	
RFS	Renewable Fuel Standard	
RPS	Renewable Portfolio Standards	
SCAD	South Campus Anaerobic Digester	
ТА	Total Alkalinity	
TN	Total Nitrogen	
TP	Total Phosphorus	
TS	Total Solids	
TSS	Total Suspended Solids	
VFA	Volatile Fatty Acids	
VS	Volatile Solids	
VSS	Volatile Suspended Solids	
WEP	Water Eutrophication Potential	

CHAPTER 1. INTRODUCTION

1.1 Problem Statement

To address climate change issues and contribute to consuming more renewable energy, Michigan State University launched South Campus Anaerobic Digester (SCAD), a commercial anaerobic digester that produces biogas and renewable electricity from livestock manure and food waste from student's hall to provide energy to south campus buildings. This facility operated starting in 2013 and is still running today.

Throughout the food system, there is a significant loss of food due to spoilage, processing, and damage. The largest source of food waste in American waste occurs at the household level, where approximately 0.5 lb./d/person of food waste is generated (U.S. Environmental Protection Agency, 2018). Manure management is a significant source of methane (CH4) emissions and a contributor to the carbon footprint of food production. For one of gallon milk, manure management accounts for 24% of the carbon dioxide equivalent (CO2-e) emissions, which are largely related to CH4 emissions from long-term manure storage (Thoma et al., 2013). Animal handling sectors such as farms contribute about 18% of GHG emissions when it is not responsibly managed (Esfandiari et al., 2011). Meanwhile in the United States itself, approximately 14% of ammonia emissions come from livestock manure management which is one component of acid rain (Eckert et al., 2018). Environmental impacts associated with food waste and manure management are important contributors to climate change. One of the solutions to address climate change is to produce and utilize more renewable energy sources, especially waste materials. Given the statement, the food system contributes to vast quantities of organic waste materials, such as livestock manure and food waste. These resources can be utilized to produce biogas and further to be renewable electricity (Hosseini & Wahid, 2014).

The SCAD system has not had a comprehensive evaluation of the operational performance or the environmental impact of its existence to date. These operational performances include feedstock supplies (livestock manure and food waste), digester performance, biogas quantity and quality, electricity production, and laboratory analysis of influent and effluent. A life cycle assessment (LCA) needs to be conducted to evaluate SCAD in reducing the negative impact on the environment, such as greenhouse gas (GHG) emissions. Evaluating these parameters will provide valuable information on whether SCAD has been operating in its best condition and provide further recommendations on improving SCAD performance to continue serving the MSU community.

1.2 Goal and objectives

The goal of this study is to evaluate the operational performance of Michigan State University's commercial South Campus Anaerobic Digester (SCAD) as well as the environmental impact during its operation from 2014 to 2020. Additionally, this study aims to achieve these objectives:

- 1. Evaluate the feedstocks supplies quantity that SCAD received during 2014-2020
- 2. Evaluate output parameters of SCAD during 2014-2020
- 3. Determine feedstocks that potentially have the most significant impact on biogas production
- 4. Compare a life-cycle assessment of SCAD to the conventional waste management method
- Conduct a technoeconomic analysis to know the financial feasibility of SCAD as a commercial digester
- 6. Summarize lessons learned and operational experiences
 - 2

CHAPTER 2. LITERATURE REVIEW

Chapter 2 provides a series of information related to this study. It starts with changes in the global community over several decades, which has eventually led to climate change. Some of these changes have come from the food system, such as food and manure waste from farming. This waste has disrupted the ecosystem balance. To overcome this situation, a series of policies have been implemented to reduce the rate of climate change. This chapter specifically discusses the efforts that the U.S. government has taken to mitigate the impact of climate change, narrowing down to the promotion and utilization of renewable energy sources.

There are various types of renewable energy, including biogas that is produced from the anaerobic digestion (AD) process. The next step of this chapter talks about the AD and several factors that affect its performance. The advantage of this process led Michigan State University (MSU) to launch the South Campus Anaerobic Digester (SCAD), a commercial digester that provides electricity to power buildings on the south campus area. Furthermore, a life cycle assessment (LCA) was conducted to evaluate SCAD's performance in mitigating environmental burdens. The revenue section concludes this chapter to showcase the profits gained by the digester.

2.1 General introduction – global drivers

2.1.1 Societal change

Societal change is an alteration of social structure where social relationships become involved in the process. Relationship in this context refers to interactions, patterns, and processes that involve mutual activities of the various parts of society (Greenwood & Guner, 2008; Sharma, 2007). For example, significant societal change has been experienced by many countries in recent decades as the result of economic restructuring, changes in societal value systems, the spread of media technology, and changes in educational systems or population composition (Weichold & Barber, 2009).

Societal changes are the result of many factors. One of the most significant factors in the recent era technological advancement. Technology exercises its vast influence by changing the environment that demands society to adapt by modifying social norms (Greenwood & Guner, 2008). Endless new technological discoveries have created more machines and methods of communication which alter social interactions. For example, society is experiencing change due to the development and invention of electric, steam, and petrol driven machines for food production. This change cannot be avoided even for institutions like family and marriage. It affects the lives of children, adolescents, and adults throughout the family dynamic, resulting in a risk of emotional development, or less social control around neighborhoods. The obvious effects of technological advancement are labor organizations, specialization, high speed of life, and increase in production (Sharma, 2007; Weichold & Barber, 2009).

Michigan State University Sustainability explains that the advancement of electronic devices such as the latest cell phones and televisions have altered the behavior of students at every corner of campus resulting in the consumption of more energy. During 1965, a student consumed 66 watts of electricity for lighting and playing vinyl records in a student's dorm room. By 1978, the number rose to 255 watts to turn on small fridge, television, light, and radio. Moving forward to 2013, the number significantly increased to 3,671 watts to power computers, television, audio speakers, microwaves, cell phones, and mini fridges. The increasing appetite for power from student in Residence Halls is causing the University to improve energy systems to support campus growth (Michigan State University Sustainability, 2014).

Societal changes in technology have affected change in other related fields, such as agriculture, food, and the environment. Technology advancement has also improved crop production. Moreover, technology helps with food processing, prolonging expiration dates. Nevertheless, agriculture and food processing demand more resources. Agriculture has been correlated with land use changes from forest to farmland, as well as an increase in water demand. The food industry has increased the amount of food waste in landfills, creating new sources of diseases caused by rotten food. Eventually, these issues contribute to climate change which will be discussed in section 2.1.2. Various social movements were conducted to address climate change issues, such as reducing plastic bag use, saving electricity, and providing more mass transportation. Michigan State University enforced several breakthroughs to accomplish this mission, one of them is the establishment of SCAD in 2013 to address the needs of renewable energy sources and food waste management. SCAD will be discussed in section 2.3.

2.1.2 Climate change

Traditional energy sources that mostly consist of fossil fuels such as petroleum, coal, and natural gas consequently contribute to environmental pollution (Zhou et al., 2016). A consequence of fossil fuel use is the significant increase in greenhouse gas (GHG) in the atmosphere which has been observed since the late nineteenth century. GHG effect is a natural phenomenon where sun heat radiation on earth surface gets absorbed by gases in the atmosphere and re-emitted to all directions, increasing earth surface temperature as a result (Quinto et al., 2016). GHG, which includes compounds such as carbon dioxide (CO₂), methane (CH₄), nitrous oxide (N₂O), and ozone (O₃), have contributed to raising global temperature and potentially became the cause of climate change and global warming (Hosseini & Wahid, 2014).

The United Nations, through The Sustainable Development Goals Reports in 2019 reported that temperature is increasing globally at an average of 0.85 °C from 1880 to 2012. This increase has affected major crop yield, contributed to the melting of snow and ice which has caused an increase in sea level. If this trend continues, global temperature might exceed 1.5 °C causing a sea level rise of 40-63 cm by 2100. Not just causing environmental pollution, but traditional energy sources are also non-renewable. The increase of human activities, particularly in industry and economic sector, requires a larger energy supply. Increasing energy is demanded in industrialized countries while significant populations in developing regions lack reliable energy. Currently, approximately 3 billion people have limited access to clean-cooking solutions and 840 million people have restricted access to electricity. Therefore, increasing clean and renewable energy use is essential to create more resilience communities to face climate change (United Nations, 2019).

2.1.3 Food waste diversion

Food waste has become a global environmental issue over the past decade due to its environmental impact. Additionally, there are approximately 800 million people who suffer from hunger. A third of the food produced for global human consumption is lost or wasted every year, approximately 1.6 billion tons. This loss costs \$2.6 trillion, of which \$1 trillion is incurred from GHG emissions, water scarcity, biodiversity loss, increased conflicts, and loss of livelihood. Contributing factors are soil erosion, nutrient loss, reduced yields, wind erosion, and pesticides exposures. Moreover, food waste is responsible for 4.4 gigatons of CO₂-e per year. That number represents 8% of the global anthropogenic GHG emissions (World Biogas Association, 2018).

The United States Environmental Protection Agency (U.S. EPA) reported in 2018 that over 35 million tons of food were sent to landfills, which is the equivalent to half a pound per person per day. This waste costs the commercial food service industry about \$100 billion annually. The

number significantly increased almost a decade later. The country spent \$218 billion (about \$670 per person in the U.S.) to grow, process, transport, and dispose uneaten food. Each year, the number of foods dumped to the landfill is approximately 52.4 million tons. Meanwhile, there is an additional 10.1 million tons of unharvested food remaining on farms, thus the total of wasted food is roughly 63 million tons annually. From this number, the U.S. restaurant sector itself contributes 11.4 million tons of food waste annually. This number consists of 7.3 million tons from full-service restaurants and 4.1 million tons from limited-services restaurants (ReFED, 2018; U.S. Environmental Protection Agency, 2018).

According to the Michigan Recycling Coalition, of all municipal solid waste disposed in Michigan landfills, 13.6% is food waste. Food waste prevention is the most essential process to reduce the number of organic wastes going through the landfills, which is the top priority from environmental, social, and economic perspectives. The challenge is that the prevention efforts are difficult to quantify considering the diverse waste composition from each household. Therefore, the State of Michigan puts their effort into food surplus diversion by feeding hungry people and then animals. The Food Bank Council of Michigan cooperates with multiple partners such as farmers, individuals, non-profit organizations, corporations, and government to support 7 regional food banks and advance the access to healthy food for people in need in all 83 Michigan counties (O'Brien, 2017).

The U.S EPA explained that food waste diversion should be considered seriously because of the environmental impact of food waste can be damaging. It starts with food waste rooting in landfills that create CH₄, a greenhouse gas (GHG) 25 times stronger than carbon dioxide. Among GHG emissions in the United States, 13% results from the growth, manufacture, sale, transport, and disposal of food. Not only that, wasted foods consumes more than a quarter of the total freshwater in the country (United States Environmental Protection Agency, 2014).

Shifting the management system of food waste from linear to circular can be an option to optimize food waste as a valuable resource. For example, utilizing food waste for biogas production through anaerobic digestion can provide many benefits, such as generating renewable energy, reducing GHG emissions, protection of water bodies, additional revenue streams from sale of electricity and compost, creating more self-sufficient and resilient communities and sustainable industrialization (World Biogas Association, 2018).

2.1.4 Manure management

The existing waste treatment solutions for manure and slurries systems still have several drawbacks regarding soil contamination and negative environmental impacts, such the contamination of water streams and food crops with pathological entities (Goldstein, 2018). Animal handling sectors such as farms contribute about 18% of GHG emissions when not responsibly managed (Esfandiari et al., 2011). Meanwhile in the United States, approximately 14% of the ammonia emissions come from livestock manure management which is one component of acid rain (Eckert et al., 2018).

The perspective of manure as a waste product should be changed, as manure has valuable benefits to save fertilizer cost with proper storage, handling, and application (Courneya, 2010). A critical aspect of sustainable manure management is by constructing housing and manure storage systems thus helping to conserve and maintain high concentrations of nutrients (International Atomic Energy Agency, 2008). To do so, animal manure should not be mixed with human waste to control disease and parasites (Lorimor & Powers, 2018). Moreover, improper manure management can lead to environmental issues such as water pollution and source of odor, flies, and parasites (Bradley, 2019). Lorimor and Powers (2018) classified manure into four different solid contents, as explained in Table 2.1 below.

Solid content	Treatment	
4% or below	Treated as a liquid with irrigation equipment, such as liquids which	
	majority of solids have been removed or diluted manure	
4 to 10%	• Handled as a slurry	
	• It may require special pumps, for instance swine pit manure and dairy	
	manure with milking parlor wash water added	
10 to 20%	• This is typical of many dairy operations which are too thick to pump	
	but too thin to scoop	
	• Adding water to handle it as liquid and special pumps will be used to	
	agitate and move the manure	
20% or more	Treated as a solid which can be stacked and picked up with a fork or	
	bucket loader	

 Table 2.1 Manure treatments based on solid content

Bradley (2019) emphasizes that manure management and utilization plan must include these components: 1) Quantity of manure and bedding generated annually from all livestock on the farm; 2) Manure handling, collection methods, and equipment used. This includes the handling from barns, stalls, paddocks, and pastures; 3) Size and location of storage and/or composting facilities; 4) Methods used to prevent drainage into storage areas, paddocks, and pastures; 5) Nutrient analysis of manure prior to application (if applying to land); 6) Soil analysis for lands in which raw or composted manure will be applied; and 7) Utilization records: land application, compost monitoring, or off-sites uses.

2.1.5 Renewable energy

Renewable energy is energy generated partially or entirely from natural resources which are available on a renewable basis and inexhaustible (naturally replenished), for instance hydropower, wind, solar, geothermal, and bioenergy (plant-based). Renewable energy can be utilized without transforming it into other forms, such as cooking gas. Otherwise, renewable energy would be processed for electricity generation (Gorjian, 2017; The U.S. Environmental Protection Agency (EPA) State and Local Energy and Environment Program, 2018). Renewable energy has received noticeable interest from many countries as a solution to limited energy resources and environmental issues related to fossil fuels consumption (Mohammadrezaei et al., 2018).

According to Abolhosseini et al. (2014), the development of renewable energy technologies was driven by three factors: energy security, economic impacts, and carbon dioxide emissions reduction. The Arab oil embargo in 1973 played a vital role in raising the awareness of energy supply security. This was supported by high oil prices, increasing dependency on oil imports, depletion of fossil fuels, increasing competition from emerging economies, political instability in major oil producers, a high impact due to any disruption in energy supply on developed, and rapidly developing countries. Moreover, energy security issues must address climate change concerns, which demands the diversification of energy sources. That means if energy security is fulfilled only by fossil fuels, it will significantly increase GHG emissions which will worsen climate change. Therefore, energy demand must be fulfilled by renewable energy to embody energy security while lowering GHG emissions. Shifting to economic impacts, renewable energy opens job creation opportunities, industrial innovation, and a balance of payment.

Additionally, renewable energy technologies should reduce CO₂ emissions by substituting fossil fuels in the power generation industry and transportation sector.

Biogas is one of renewable energy sources that has become an alternative fuel for transportation, industrial engines, and residential electricity and heat. Biogas as an alternative fuel has the capacity to reduce the impact of pollution. Additionally, biogas can be utilized further to produce electricity and/or heat through combustion by combining heat and power (CHP) generation systems. The efficiency could reach 34-40% by using large turbines and 25% using small generators (Hosseini and Wahid, 2014). In Chile, electricity generated from biogas could fulfill 50 megawatts (MW) of demand. Moreover, the slurry and residues could be used as fertilizer (Passos et al., 2017; Zhou et al., 2016).

Biogas composition primarily consists of CH₄ and CO₂ with a small amount of hydrogen sulfide (H₂S) and water vapor (H₂O). Biogas typically contains 55-70% of CH₄ that can be used as fuel in a variety of purity levels and efficiencies (Chynoweth et al., 2001; Somers et al., 2018). Biogas which contains high concentration of CO₂ will lower its quality as cooking gas since CO₂ contributes to carbon monoxide (CO) formation which could weaken biogas combustion. Additionally, H₂S reduction level (i.e., via desulphurization processes) must happen before biogas can be used as H₂S is a toxic gas which poses safety concerns for people and may result in additional equipment maintenance (Hosseini & Wahid, 2014). Biogas is produced through anaerobic digestion (AD) which will be discussed further in section 2.2.

2.1.6 Policy drivers – focus in the U.S.

2.1.6.1 Food waste diversion

According to O'Brien (2017), an amendment of the Solid Waste Management Act in 1990 attempted to minimize the hazardous landfill gases yield by banning the disposal of yard clippings. At the same time, this effort promoted the conversion of valuable organics into beneficial resources for municipalities, agriculture, and industry. The Michigan Legislature enacted Public Act 264 in 1990 which define "yard clippings" as "leaves, grass clippings, vegetable or other garden debris, shrubbery, or bush or tree trimmings less than four feet in length and two inches in diameter that possibly to convert into compost humus." Due to this amendment, the Natural Resources and Environmental Protection Act (NREPA) prohibited owners or operators of landfills and municipal solid waste incinerators from accepting solid waste if it was known that the solid waste included yard clippings generated or collected on land owned by a county, municipality, or a state facility.

Granholm and Chester (2007) use the three principles of sustainability – economic vitality, ecological integrity, and improved quality of life – to guide solid waste management decision-making. According to the Policy, Michigan's preference is to first avoid waste generation, then to utilize generated waste for beneficial purposes, and properly dispose of what remains.

According to EPA (2014), much of food waste that reaches landfills is still edible. Wasted food can be divided into three categories as explained in Table 2.2.

Category	Description	Example
Avoidable	Food that can be easily prevented from	an entire tray of lasagna is
	going to waste. Reasons for waste include	left over every day at a
	overpreparation, improper storage, or	buffet
	spoilage. Understanding the cause of this	
	waste is key to preventing it.	
Possibly avoidable	Food that may seem inedible but can be	beet tops can be cooked
	used or repurposed	similarly to collard greens or
		spinach instead of discarded.
		Also, slightly stale bread can
		be used for croutons or
		breadcrumbs
Unavoidable	Food that cannot be consumed by people	banana peels or peach pits
	and should be used for animal feed,	
	compost, or anaerobic digestion	

Table 2.2 Food waste classification

Moreover, EPA also created the Food Recovery Hierarchy to reduce wasted food. Based on this hierarchy, there are five stages to be considered when utilizing excess food:

- a. Source reduction: reduce the volume of food waste generated
- b. Feed people: donate extra food to food banks, soup kitchens and shelters
- c. Feed animals: provide food to farmers for animal feed
- d. Industrial uses: provide fat for rendering, biofuel, and food discards for animal feed production
- e. Composting/digesting: convert food scraps into a nutrient rich soil amendment

2.1.6.2 Renewable Portfolio Standard (RPS)

A renewable portfolio standard (RPS) is a policy that demands that electricity providers include a minimum percentage of their electricity supplies from renewable energy sources, such as wind, solar, geothermal, and various forms of biomass and ocean energy. Electricity suppliers have two options to comply with this requirement. First, suppliers can own a renewable energy facility and produce their own electricity. Second, they can purchase renewable electricity from a renewable facility. Furthermore, RPS enables the market to choose any renewable energy resources to fulfill the mandate since renewable energy availability varies depending on the regional climate and geographies (Cory & Swezey, 2007; Rader & Hempling, 2001).

From an environmental perspective, RPS promotes climate change mitigation by reducing air and carbon pollution, waste reduction, and conserving water and other valuable natural resources. Moreover, it can boost local economic development by creating more jobs, taxes, and revenues associated with renewable energy (Environmental Protection Agency, 2015).

According to Lawrence Berkeley National Laboratory and U.S. Department of Energy, 29 States have implemented RPS, including Michigan, California, Nevada, and Minnesota (Figure 2.1). Additionally, Washington DC and 3 territories (Northern Mariana Island, Puerto Rico, and US Virgin Islands) have also had RPS. Meanwhile, 8 states (including Kansas, Oklahoma, and Indiana) and 1 territory (Guam) have established their renewable portfolio goals (Barbose, 2019; U.S. Department of Energy, 2016).

RPS Policies Exist in 29 States and DC



Apply to 56% of Total U.S. Retail Electricity Sales

Figure 2.1 RPS Policies Established in the United States (Barbose, 2019) RPS implementation's most prominent mechanism is renewable energy certificates

(RECs). RECs are payments that electricity suppliers offer to renewable energy companies for the renewable electricity they provide to the grid. One REC typically represents 1 megawatt-hour (MWh) of renewable electricity. The implementation of REC provides an accurate, durable record of electricity produced, also a fungible commodity that can be traded among providers to fulfill RPS (Cory & Swezey, 2007).

2.1.6.3 Renewable fuel standard

The United States is the biggest consumer of crude oil in the world. Consequently, the country is facing two major concerns which are low energy security and high greenhouse-gas emissions. Furthermore, the United States imported between 52 to 60% of oil consumed from 2005 to 2009 with roughly 30% of CO_2 emissions resulting from transportation fuels (The National Academy of Sciences, 2011).

Biofuels emerge as an alternative to petroleum-based fuels due to its production from renewable domestic sources which has the potential to improve U.S. energy security. It also provides life-cycle greenhouse-gas benefits when compared to fossil fuel. As a result, the U.S. Congress enacted the Energy Independence and Security Act (EISA) in 2007 to lead the country toward energy independence and security while improving clean renewable fuels production. EISA then resulted in the Renewable Fuel Standard (RFS) program which was created by Congress to address greenhouse gas emissions, broaden renewable fuel sector in the country₂ and dwindle the demand for imported oil (America's Oil and Natural Gas Industry, 2017).

RFS demands that U.S. transportation fuel contains a minimum volume of renewable fuel. It sets the target for consuming 35 billion gallons of ethanol-equivalent biofuel and 1 billion gallons of biomass-based diesel by 2022. The Environmental Protection Agency (EPA) holds statutory authority on volume amounts after 2022. There are four categories of renewable fuels according to RFS: total renewable, advanced, biomass-based diesel (BBD), and cellulosic. These categories are determined based on the reductions in life-cycle emissions of GHG, relative to petroleum, feedstock, and fuel characteristics (Congressional Research Service, 2020; Stock, 2015).

Nevertheless, many reports and studies are doubtful that the implementation of RFS will be successful and meet the target. First, there is a limit placed on car use of gasoline at 10% ethanol (E10) which is the maximum acceptance under the manufacturer's warranty. This caused an E10 plateau in the US fuel supply in 2013. Second, cellulosic biofuels experience policy uncertainty and high production costs that might hinder investors from supporting this initiative. Eventually, it depends on how biofuels are produced. There are many factors that can affect it that might be unpredictable. These factors include technical expertise, weather conditions, market stability, tax incentives, and trade disputes. These factors could impact the entire industry (Congressional Research Service, 2020; Stock, 2015). To facilitate RFS compliance, Renewable Identification Number (RIN) systems were created by the U.S. EPA. A RIN is a 38-character numeric code that represents a volume of renewable fuel produced in or imported into the United States. One RIN is equivalent to one gallon of biofuel produced and reported to EPA. A RIN contains information about biofuel sources like, production year, biofuel producer, and the type of fuel (Table 2.3). At the end of the calendar year, fuel suppliers must fulfil minimum requirements of RINs to be compliant with RFS. Similar to REC, RIN also can be traded like other commodities. RINs are valid for use during its production year and the following year (Christensen et al., 2014; Cooper, 2018; McPhail et al., 2011; Yacobucci, 2013).

38-character code:		
KYYYYCCCCFFFFFBBBBBBRRDSSSSSSSSEEEEEEE		
К	RIN assignment code	
YYYY	Year batch is produced/imported	
CCCC	Company registration ID	
FFFFF	Facility registration ID	
BBBBB	Producer-assigned bath number	
RR	Equivalence value for renewable fuel	
D	Renewable type code ¹	
SSSSSSSS	RIN block starting number	
EEEEEEE	RIN block ending number	

Table 2.3 RIN code definitions (McPhail et al., 2011)

¹Five separate RIN categories: D=3 for cellulosic biofuel; D=4 for biomass-based diesel; D=5 for advanced biofuel; D=6 for other renewable fuel; D=7 for cellulosic diesel

2.1.6.4 California low carbon fuel standard

The California Low Carbon Fuel Standard (LCFS) is a program created to reduce greenhouse gas emissions (GHG) and other air pollutants from the state's transportation sector, diversify the State's fuel mix, and decrease the dependence on petroleum. It is one of the policies in California that comes from the implementation of the Global Warming Act of 2006. LCFS evaluates the full life cycle emissions of transportation fuels and includes all GHG emissions that are the result of production, distribution, and consumption expressed as grams of CO_2e per megajoule (Renewable Fuels Association, 2021; Townsend & Havercroft, 2019). Recently, the transportation sector has contributed 50% of GHG emissions, 80% of nitrogen oxide emissions, and 95% of particulate matter emissions in California (Center for Law, 2019).

LCFS reduces GHG emissions by establishing annual standards to be followed by fuel producers and distributors. It emphasizes the reduction of the average life-cycle carbon intensity (CI) of the fuels supplied to the market. Fuels that gain lower CI values regulated by California Air Resources Board (CARB) will get compliance credits. On the other hand, fuels with CI values higher than the standard will get compliance deficits (Renewable Fuels Association, 2021). Additionally, certain requirements might be needed to fulfill LCFS such as life cycle assessment (LCA) which accounts for the environmental impact of fuel production from feedstock production to end use stage (Congressional Research Service, 2021).

The implementation of LCFS began in 2011. It was amended in 2015 to address fuel CI reduction standard adjustment and extension. Currently, LCFS requires a 10% reduction by 2022 and a 20% reduction by 2030. Since its first implementation, LCFS has achieved more than 77 million credits with each credit referencing a metric ton of GHG emissions reduction contra the annual standard (Renewable Fuels Association, 2021). It has helped to incentivize production of low-carbon fuels and generate additional revenue to encourage investment in statewide low-carbon transportation fuel infrastructure (Center for Law, 2019).

Despite its merits LCFS has some implementation challenges. These challenges include determining the appropriate energy related to GHG target, developing a robust LCA, as well as constructing a transparent compliance system. Moreover, part of Congress opposed LCFS due to concerns about economic effects such as job loss, limited affordable lower carbon fuel options,
and increasing fuel price as an effect of fulfilling the certain standards required. Therefore, further improvement of the LCFS is essential (Center for Law, 2019; Congressional Research Service, 2021).

2.1.6.5 Carbon Intensity

Carbon intensity (CI) is the number of GHG emission generated throughout production and use of life cycle energy sources, such as transportation fuel. CI is represented in units of grams of CO₂ per megajoule of energy (gCO₂e/MJ). CI calculation considers extraction, refinement, distribution, storage, and combustion of energy. Therefore, the calculation can be included in life cycle assessments (Ingram, 2015; United Nations Environment Programme, 2019).

CI is calculated using the Greenhouse Gases, Regulated Emissions and Energy Use in Transportation (GREET) model which is developed by U.S. Department of Energy and Argonne National Laboratory. The current model is the CA-GREET 3.0 Model and Tier 1 Simplified Carbon Intensity calculators. The GREET model aims to evaluate and compare energy, environmental impacts of transportation fuels, and vehicle technologies on a life-cycle basis. Over 100 alternative fuel pathways and over 80 vehicle technologies have been evaluated using this model: including aviation fuel, aircraft operation, marine fuels, and vessel operation. Furthermore, GREET evaluates total energy consumption, GHG emissions (CO₂, N₂O, CH₄), air pollutants (SOx, NOx, VOC, CO, PM10, PM2.5), and water consumption (Argonne National Laboratory, 2014; M. Wang, 2007).

2.2 Anaerobic digestion

2.2.1 Introduction

The anaerobic digestion (AD) process is defined as a biological process in which a lack of O_2 causes a degrade in carbon source materials through the help of various microorganism consortiums (Chynoweth et al., 2001). It transforms organic matters into CH₄ gas, thus reducing odor and pathogen risks (R. Chen et al., 2016). AD does not only provide a renewable alternative energy source, but also an alternate pathway to process organic waste, reduce GHG emissions from landfills, and mitigate the demand of fossil fuel and chemical fertilizers (Li et al., 2018; Zou et al., 2018). In terms of efficiency and costs, AD energy could compete with other biomass energy sources, such as heat, synthesis gases, and ethanol (Chynoweth et al., 2001). Moreover, AD could provide electricity storage via upgrading biogas to high purity of methane which might be stored and used for other purposes (Jürgensen et al., 2018). One factor that makes AD attractive is the technology used in AD is scalable (Hosseini & Wahid, 2014). This means AD technology can be applied in various digester capacities; thus, the digesters can be less costly to build, operate, and maintain. This scalability also opens the chance to implement AD technology for small scale farms or communities. For example, the small-scale AD project might be carried out starting from 100 dairy cows or 200 cows or between 200 to 5,000 tons of organic waste per year which is expected to produce 80 kW electricity (Marjolaine, 2019).

AD consists of four main stages: hydrolysis, acidogenesis, acetogenesis, and methanogenesis. The first three stages are maintained by bacterial communities to yield acetate, hydrogen, and carbon dioxide, while the last stage is maintained by methanogenic archaea communities to produce methane from acetate, or from hydrogen and carbon dioxide as alternative sources (Z. Yu et al., 2018; Zhou et al., 2016).

In the hydrolysis process, organic materials such as cellulose and hemicellulose are hydrolyzed, becoming soluble sugars, alcohol, and other organic substances. The hydrolysis products process will be transformed by acid producing bacteria to yield volatile fatty acids that consist of formic, acetic, propionic, and butyric acid. Hydrolysis and acidification are continuous biochemical reactions. When these processes are performed ideally, high yields of organic acids are achieved, degradation or loss of the organic acids is avoided, fermentation inhibitors are minimized, and cellulose and hemicellulose degradation are improved (J. Yu et al., 2017). Methanogens utilize formic and acetic acid directly, whereas propionic and butyric acid need to be converted into acetic acid by acetic acid producing bacteria (Shen et al., 2018).

Temperature and pH play key roles in developing appropriate AD systems. In most studies, AD can be developed in mesophilic and thermophilic temperatures. Mesophilic temperature is preferable if hydrolysis and acidification are done within high solid content condition, while thermophilic temperature is recommended for improving AD efficiency under low solid content in 3 days (J. Yu et al., 2017). An ideal digestion should have a pH range of 6.5-8.0 to provide a convenient environment for anaerobic microbes, especially for archaea in degrading organic compounds and producing methane (Zhong et al., 2015). Appropriate pH for hydrolysis and acidogenesis is 5.5 and 6.5, respectively (Zhou et al., 2016).

2.2.2 Feedstock

2.2.2.1 *Manures*

Historically, manure has been utilized for many purposes, such as fertilizer, soil amendment, energy source, and even construction material. There are many recyclable components in manure, including solids, organic matter, nutrients, and fiber. The contents in manure can be affected by several key factors such as species, digestibility, protein and fiber content, diet, age, housing, environment, and production stage. Moreover, manure contains nitrogen, phosphorus, and potassium which are critical nutrient sources for crops. In general, manure with higher solids concentrations have higher nutrition content (Kostic et al., 2020; Lorimor & Powers, 2018; U.S. Environmental Protection Agency, 2015).

Nutrients in manure are available in soluble and insoluble forms. Soluble nutrients can be consumed by the crops right away, while insoluble nutrients take up to a year or more to be available. Each nutrient has a different characteristic. For example, phosphorus is typically 80% available in the settled solids of manure storage and insoluble. Meanwhile, potassium is typically 80% found in the liquid and is highly soluble. Nitrogen is split almost evenly between liquid and solids (Lorimor & Powers, 2018).

Among the various methods to utilize manure, anaerobic digestion is a technology that processes manure into biogas as source of heat or electricity. The energy produced can be used on the farm or sold to the local power grid (U.S. Environmental Protection Agency, 2015). Diverse types of manure contribute slightly different percentages of CH₄ in biogas, as shown in Table 2.4.

Table 2.4 Percentage	composition	of	CH ₄	from	anaerobic	digestion	of	various	manures
(Anukam et al., 2019).									

Manure	CH ₄ Composition (%)
Cattle	50-60
Pig	60
Poultry	68
Sheep	65
Horse	66

Manure's characteristics are critical factors for biogas production and process stability during anaerobic digestion. These characteristics are moisture content, total solids (TS), volatile solids (VS; organic compounds from plant or animal that lost when the dry solids are burnt at 550 °C), biodegradability, particle size, pH, biological oxygen demand (BOD), chemical oxygen demand (COD), also carbon and nitrogen contents (L. Chen & Neibling, 2014).

Additionally, manure contains variative nutrition contents as shown in Table 2.5. These contents are important for the digestate once it is land-applied. The anaerobic digestion process slightly changes the nutritional contents of manure. Most nitrogen is found in the form of organic N and NH₄ in liquid phase of digester sludge. A negligible amount of nitrogen might be emitted as NH₃. Therefore, digester sludge will have higher ammonium content than raw manure. Meanwhile, phosphorus content does not change significantly due to anaerobic digestion process. All the P present in the manure will still be present in the digester sludge. The difference is that the dissolved portion of P will be moved into bacteria bodies that perform anaerobic digestion. Additionally, C in the form of simple sugars, volatile fatty acids, and alcohol is converted to CO₂ and CH₄. As a result, effluent has less C or organic matter compared to raw manure (Manitoba, 2015; Natural Resource Conservation Services, 2007).

Manure	Percentage content							
i i i i i i i i i i i i i i i i i i i	Nitrogen (N)	Phosphoric acid (P ₂ O ₅)	Potash (K ₂ O)					
Cattle, fresh	0.4-0.5	0.3-0.4	0.3-0.4					
Horse, fresh	0.5	0.4-0.6	0.3-1.0					
Poultry, fresh	1.0-1.8	1.4-1.8	0.8-0.9					

 Table 2.5 Nutrition content of Manures (Chandra, 2005)

2.2.2.2 *Food waste*

In this case, material contents are essential as they provide nutrition for the microbial community in the digester. The carbon/nitrogen (C/N) ratio in slurry has a critical contribution in the AD process. If slurry contains high C/N ratio, microbe will consume nitrogen sources rapidly

thus decrease biogas productivity. On the other hand, methanogen bacteria reproduction and metabolism will be inhibited as carbon deficit, ammonia accumulation, and pH increase occur in low C/N ratio slurry. Typically, C/N ratio range for AD is 10-90, but 30 is the most common (H. Chen et al., 2016; X. Wang et al., 2012).

To achieve an ideal C/N ratio, previous research combined several resources as AD materials what is known as co-digestion. The co-digestion of various substrates enhances methane yield, improves buffering capacity, and prevents acidification (Li et al., 2018; X. Wang et al., 2018). Compared to animal manure only, applying the AD process to the co-digestion between animal manure and feedstock provides more stable operational performance and produces more methane. This happens because feedstock with a higher C/N ratio balance out with manure which has low C/N ratio achieving an ideal C/N ratio range (X. Wang et al., 2012).

The mixture of food waste and dairy manure also attracts attention as this combination potentially has an ideal C/N achieving productive and efficient AD processes (H. Chen et al., 2016; X. Wang et al., 2012). Li et al. (2018) revealed that methane production could be improved by codigesting dairy manure and corn stover with tomato residues at 20-40% volatile solid based. However, this process will happen once tomato residue is added beyond 60% since it may trigger pH drop and over produce VFA (Volatile Fatty Acids).

2.2.3 Testing and analysis – Operation data

2.2.3.1 Operational parameters

2.2.3.1.1 pH

pH is a measure of acidity or basicity of a solution. Specifically, pH measures the hydrogen concentration, [H⁺] which the value ranges from 0 to 14. Additionally, pH 7 is called "neutral"

because it is the center of the measurement scale, where the ratio of [H⁺] and [OH⁻] (hydroxide ion concentration) is equal. A solution is determined to be acidic if the pH is below 7 because [H⁺] is greater than [OH⁻], while a solution with pH above 7 is grouped as basic or alkaline which means [OH⁻] is greater than [H⁺] (Hach, 2018). An anaerobic digester typically has a pH range between 6.4 and 8.2 to maintain methanogens population. This can be achieved by maintaining the balance of acetogens and methanogens. Acetogens are needed to produce acids, while methanogens consume acids to yield methane gas and increase alkalinity. If acetogens surpass methanogens population, it will drop the pH which then inhibit methanogens performance, resulting in a "sour" digester (L. Chen & Neibling, 2014; MSU Anaerobic Digestion and Research Center, 2019).

2.2.3.1.2 VFA

Volatile Fatty Acids (VFAs) are intermediate products of anaerobic digestion process which emerge after polymer hydrolysis and acidogenesis prior to being degraded into acetic acid for methanogenesis stage. VFA belongs to carboxylate which has a low molecular weight consisting of 2 to 6 carbon atoms. VFAs are produced from various feedstocks, for example, agricultural waste, food waste, milk sewage, dairy whey effluent, municipal waste, and cellulose sewage (Mayer et al., 2010; Szacherska et al., 2021; Wainaina et al., 2019). VFA acts in the biopolymers of biofuels production such as methane and hydrogen. Other applications of VFA include carbon sources in biological denitrification, production of biodiesel, and electricity production through microbial fuel cells. Moreover, there are diverse types of microorganisms that contribute to VFA production as shown in Table 2.6 (Lukitawesa et al., 2020; Magdalena et al., 2019). Several factors are crucial for VFA productivity, which are hydraulic retention time (HRT), organic loading rate (OLR), temperature, pH, and pretreatment (Wainaina et al., 2019).

Table 2.6 Production	of	volatile	fatty	acids	by	microorganisms	(Szacherska	et	al.,	2021;
Wainaina et al., 2019)										

Volatile Fatty Acid	Bacteria	Substrate	
	Acetobacter aceti	Cheese whey	
	Clostridium acetium	Mixed gas (4% H ₂ :18%	
	Closiniuum acenum	Argon:78% CO)	
Acetic Acid	Clostridium lentocellum SG6	Paddy straw	
	Moorela thermoaceatica	Sugarcane straw hydrolysate	
	Saccharomyces cerevisiae +	Glucose	
	Acetobacter pasteurianus	Glucose	
	Propionibacterium acidipropionici	Lactate	
	(ATCC 4965)	Glycerol	
	(1100 4)03)	Sugarcane molasses	
	Propionibacterium acidipropionici	Glycerol	
	(CGMCC 1.223)	Gryceror	
Propionic Acid	Propionibacterium acidipropionici	Hemicellulose hydrolysate	
	(ATCC 4875)	Cheese whey	
	Propionibacterium freudenreichii	Glucose	
	CCTCC M207015	Olucose	
	Pripionibacterium freudenreichii spp.	Glycerol	
	shermanii		
	Clostridium butyricum S21	Sucrose	
	Clostridium butyricum ZJUCB	Glucose	
	Clostridium thermobutyricum	Glucose	
Butyric Acid	JW171K	Glucose	
		Corn husk hydrolysate	
	Clostridium tyrobutyricum	Sugarcane bagasse	
		hydrolysate	

VFA is expressed in equivalent milligrams of acetic acid per liter. VFA is accounted together with total alkalinity (TA), which is expressed in milligrams equivalent of calcium

carbonate per liter. VFA/TA ratio will determine the follow up action regarding feedstock supply into the digester, as explained in Table 2.7 (Lossie & Pütz, 2011; MSU Anaerobic Digestion and Research Center, 2019).

VFA/TA Ratio	Background	Corrective Action
>0.6	Highly excessive biomass input	Stop adding biomass
0.5-0.6	Excessive biomass input	Add less biomass
0.4-0.5	Plant is heavily loaded	Monitor plant more closely
0.3-0.4	Biogas production at a maximum	Keep biomass input constant
0.2-0.3	Biomass input is too low	Slowly increase biomass input
<0.2	Biomass input is far too low	Rapidly increase biomass input

Table 2.7 VFA/TA ratio and its correlation to feedstock supply into the digester

There is a correlation between VFA and pH. If VFA level increases, that will decline pH, alkalinity, and biogas production. In a normal operating system, hydrogen and acetic acid formed by acidogenic and acetogenic bacteria is immediately converted into methane by methanogens. However, if VFA is increased, an unbalanced condition between acidogenic and methanogenic activities will reduce methanogens performance. The acceptable range for VFA is between 50 to 300 mg/L as acetic, meanwhile the acceptable range for alkalinity is between 1,500 to 5,000 mg/L as CaCO₃ (Krakat et al., 2017; MSU Anaerobic Digestion and Research Center, 2019; Schnaars, 2012).

2.2.3.1.3 Ammonia

Anaerobic degradation of proteins or amino acids produces ammonia through the degradation of nitrogenous matter. Ammonia is present at elevated levels in certain feedstock, such as meat processing by-products, food waste, also swine and poultry manure. Free ammonia (NH₃) form is more toxic to methanogens than the ionized form (NH₄⁺). Moreover, anaerobes are more

sensitive to ammonia toxicity at higher temperatures. Lowering digester pH will relieve NH₃ toxicity. Additionally, dilution may be needed to measure samples with extreme NH₃ levels. The acceptable range for ammonia content is between 1,500 to 3,000 mg/L. Table 2.8 explains the effect of ammonia concentration on digester. In low concentration (between 50-200 mg), ammonia is beneficial for amino acids, proteins, and nucleic acids synthesis. Furthermore, ammonia maintains neutral pH conditions by neutralizing the organic acids yielded by fermentative bacteria. These conditions are essential for bacterial growth (Y. Jiang et al., 2019; MSU Anaerobic Digestion and Research Center, 2019; Walker et al., 2011).

Effects	Ammonia-N (mg/L)
Beneficial	50-100
No adverse effect	200-1,000
Inhibitory effect at higher pH values	1,500-3,000
Toxic	> 3,000

 Table 2.8 Effect of Ammonia-N concentrations on digester

Anaerobes, especially methanogens, are sensitive to ammonia toxicity due to ammonia can freely pass-through methanogens' cell membranes hence cause a proton imbalance. Consequently, the intracellular pH of methanogenic bacteria changes then inhibits specific enzymatic reactions. Moreover, acetate degradation is hindered by high ammonia concentration, leading to acetate accumulation, buffer capacity depletion, methane yield decrease, VFA concentration increase, and pH drop. Digesters with high concentrations of ammonia will experience methanogenesis inhibition and lead to complete failure (H. Chen et al., 2016; Morozova et al., 2020).

2.2.3.1.4 Organic Loading Rate

Organic loading rate (OLR) is defined as the quantity of organics fed to a continuous digester per day (Meegoda et al., 2018). OLR data talks about digester health and space. It is important because OLR indicates the quantity of volatile solids to be fed into the digester each day (Babayee and Shayegan, 2011). OLR can be counted with formula (1) as follows (MSU Anaerobic Digestion and Research Center, 2019):

OLR = mass of Volatile Solids (VS) / volume of reactor (1)

Material characteristics and operation conditions are key points to in determining OLR. In general, the biogas yield increase is in line with the OLR increase. If OLR is optimized, the processing efficiency of anaerobic digestion can be improved. Moreover, optimum OLR can reduce the plant capital cost, improve biogas yield, and ensure operation stability (J. Jiang et al., 2020). However, excess OLR can contribute to biogas production inhibition (Musa et al., 2018).

OLR is essential for AD operation since it is related to system stability, waste treatment ability, and biogas production. Ideal OLR is favorable for cell activity, thus contributing to increasing methane production and improving substrate degradation. Nevertheless, increasing OLR has a risk of creating excessive VFA production and lowering the pH that inhibits the entire process (Moguel-Castañeda et al., 2020).

2.2.3.2 Performance indicators

2.2.3.2.1 Mixing

There are many aspects to consider in choosing a reactor type and mixing condition. For example, a conventional AD reactor is working for 20-30 days (about four and a half weeks) of hydraulic retention time, getting minimum once feeding per day with proper mixing at 35 °C (Chynoweth et al., 2001). Mixing in biogas digester aims to achieve balanced nutrient and heat

circulation, to minimize precipitation of materials by shaping an adequate mixture of solid particle and liquid suspension and facilitate gas lifting from fermentation substrates. The quality of mixing in AD plant will affect in choosing the mixer type (Eshkaftaki & Ebrahemi, 2019; Kress et al., 2018; Naegele et al., 2014). Mixing also helps create homogenous temperature and bacterianutrient composition in the digester. Currently, this process is considered to become an important stage in biomass-methane conversion (Mohammadrezaei et al., 2018).

Nevertheless, mixing has also become a major challenge in AD field application since it consumes approximately 51% of total electricity needs in the complete process. It makes several research focused on analyzing whether mixing time and rate have a significant impact on methane production. A previous study demonstrated that reducing mixing time in a full-scale AD reactor fed with crops did not strongly affect the nutrient distribution that endangers biogas production (Kress et al., 2018). A moderate stirrer rate at 80 rpm is considered as the best condition of stirring process in AD since it provides appropriate mixing pattern. The low-level mixing rate causes high death species near reactor wall. On the other hand, a high mixing rate would affect microorganism structures in the reactor, thus declining biogas production (Mohammadrezaei et al., 2018). Moreover, appropriate mixing for digester with high solid content is beneficial but not to low solid content condition (J. Yu et al., 2017). Therefore, it is essential to determine the type of material that would be digested to choose an ideal reactor and mixing type.

2.2.3.2.2 Economic Analysis

Renewable energy production often faces challenges regarding its economic feasibility. There are several processes in AD that need attention in cost efficiency. First is pretreatment. The type of pretreatment method used, and materials condition are critical for economic feasibility analysis (Fu et al., 2018). According to Passos et al. (2017), the costs that should be anticipated come from extra energy and chemical agent used for pretreatment. These costs might fluctuate depending on the market conditions and negotiation terms between the companies. In addressing this issue, reuse or energy recovery technologies during pretreatment process might be considered to improve the economic performance (Fu et al., 2018). Second, transportation of starting materials to the digester location can be accounted as energy consumption. Therefore, if the digester is placed near the livestock farms, it can eliminate the initial energy used and the cost of biogas production (Mohammadrezaei et al., 2018).

Besides pretreatment and transportation, another method to reduce substantial costs is to improve the efficiency of process steps before feeding biogas, such as harvesting and collection. Simultaneous harvesting and mechanical pretreatment of biomass might become an efficient method to increase energy yield per hectare by selecting an appropriate harvesting machine (Tsapekos et al., 2017).

2.3 Michigan State University South Campus Anaerobic Digester (MSU SCAD)

Michigan State University's South Campus Anaerobic Digester (MSU SCAD) is part of the "Keeping it Green, Recycling Waste to Resource" campus-based projects that concentrate on reducing and reusing waste. It is a single-tank complete mixed anaerobic digester which is projected to utilize approximately 17,000 tons of organic waste per year from MSU and the greater Lansing area to yield biogas than can be converted into more than 2.8 million kWh of electricity per year (Stuever, 2013). Approximately 10% of the energy powers the facility, then the rest offsets energy production in 10 south campus buildings, which is enough energy to power about 250-300 homes (MSU Sustainability). The total cost of the project is about \$5 million, and it is expected to pay off itself in less than 15 years (Oswald, 2013).



Figure 2.2 MSU South Campus Anaerobic Digester (personal documentation) Feedstock materials include dairy manure from MSU Dairy Teaching and Research Center,

food waste from campus dining halls, food manufacturing waste from southern Michigan, also fats, oil, and grease (FOG) from local restaurants. The energy produced is used to power several buildings on South Campus (Michigan State University Sustainability, 2014). These feedstocks are received in two reception tanks, one is for manure, and the other one is for other materials. Relying on the delivery schedule and the target blend, feedstock is pumped from each reception tank into a central mix tank to be homogenized. Then, this blended material is pumped through a heat exchanger to raise up the temperature to 37.78 °C before entering the anaerobic digester. The digester is an aboveground steel tank with a liquid capacity of more than 1.7 million liters. Two hydraulically powered submersible mixers are used to keep the digester contents well blended for 25-day hydraulic retention time (Oswald, 2013).

The electricity is generated through powering a 450-kW combined heat and power (CHP) system using biogas yielded from the digester. Hot water produced by the CHP is used to maintain

the digester temperature to stay at 37.78 °C and to provide heat to the other buildings at the site. The digestate—remaining mixture of solids and liquid after digestion—is pumped to a solid-liquid separator. Separated solids will be composted while liquid will be transferred to digestate holding tank (DHT) which is an aboveground steel tank with a 7.6-million-liter capacity. An airtight membrane will allow the headspace—the space above the digestate—to be used as biogas storage and minimize odors from the systems. The digestate will be land-applied seasonally as carbon-rich fertilizer. This project provides many benefits, such as renewable energy, emissions reduction, landfill and wastewater diversion, and enhanced fertilizer with few weed seeds and first year-available plant nutrients (Stuever, 2013).

2.4 Life Cycle Assessment (LCA)

2.4.1 What is it?

Life Cycle Assessment (LCA) is a methodology to evaluate environmental loads of processes and products based on their whole life cycles. The assessment includes the extraction and processing of raw materials, manufacturing, transportation, distribution, use, reuse, maintenance, recycling, and final disposal of a product, process, or system. Therefore, LCA has been widely used due to its integrated way of managing the framework, impact assessment, and data quality. LCA is often called as a "cradle-to-grave" method which explains this process by starting with gathering of raw materials from the earth to create product and ends at a stage where all materials are returned to the earth (Khasreen et al., 2009; Odey et al., 2021; Ram & Sharma, 2017).

LCA has gained support from various institutions, including United Nations Environment Programme (UNEP) / Society of Environmental Toxicology and Chemistry (SETAC) Life Cycle Initiative, the Forum for Sustainability through Life Cycle Assessment (FLSCI), International reference Life Cycle Database System (ILCD), the European reference Life Cycle Database System (ELCD) (Odey et al., 2021). According to DEAT (Department of Environmental Affairs and Tourism) (2004), LCA is currently implemented by large groups of users, such as:

- Industry and other commercial enterprises
- National governments and local, national, and intergovernmental regulatory bodies
- NGOs (consumer organizations and environmental groups)
- Consumers (which includes governments as consumers)

Nevertheless, LCA has several drawbacks to address, which are the absence of a perceived need for LCA, scarcity of LCA expertise, access to high-quality data, and incorrect perception of the application of LCA in relation to other tools (Department of Environmental Affairs and Tourism of South Africa, 2004).

2.4.2 Process

LCA implicates a thorough assessment of environmental aspects of a product system, including all stages of its life cycle, by three major processes: 1) collecting an inventory of relevant inputs and outputs of a system; 2) formulating a thorough evaluation of the potential environmental impacts correlated with those inputs and outputs; and 3) interpreting the results in correlation to objectives of the study (Jensen et al., 1997). LCA is conducted based on ISO 14040 which consists of four analytical stages (Khasreen et al., 2009):

- Defining goal and scope
- Creating the life-cycle inventory
- Assessing the impact

• Interpreting the results

2.4.3 Parameters

2.4.3.1 Water

The amount of freshwater on Earth is only about 2% of all water on the planet. In terms of access to water, approximately one out of six people on Earth lacks access to drinking water. A product's life cycle, either directly or indirectly, has a strong correlation with water consumption. Based on LCA methodologies mentioned in ISO 14040:2006, water is a parameter assessed in water consumption potential (WCP) section, which is described as water that has been removed from the watershed and cannot be returned (Arosemena, 2021).

2.4.3.2 Nutrients

The primary concern about nutrients in the LCA study is the impact of excess nutrients that pollute the environment, such as land, water body, and air. This correlates nutrients with several impacts studied in LCA, which are Water Eutrophication Potential (WEP) and Air Acidification Potential (AAP).

Eutrophication is a situation where a water body contains excessive nutrients that affect the dense growth of plant life and the death of water animals due to a lack of oxygen. It is due to nutrients runoff from the land, such as nitrogen and phosphorus, which then accumulate in the water. As the consequence, it creates a "dead zone" which is an area with low oxygen content that suffocates marine life (Mueller and Helsel, 1996). Water eutrophication potential (WEP) is the impacts resulting from excessive nutrient supplies on terrestrial and aquatic environments, particularly the most important substances such as nitrogen (N) and phosphorus (P). WEP can be presented as either mass of nitrogen equivalents (kg N-eq.) or phosphate equivalents (PO4-eq.) (Guinee, 2002).

According to the EPA, atmospheric acidification can be defined as: "the result of the oxidation of sulfur, nitrogen, and organic compounds to form their corresponding acids" (Durham & Demerjian, 1985). When absorbed by the atmosphere, these acids can lead to conditions such as acid rain. Air acidification potential (AAP) is an impact category used to convert processes or materials that form acid rain into common units of sulfur dioxide equivalents (SO2-eq.).

2.4.3.3 Greenhouse Gases (GHG)

Greenhouse gases is a parameter in LCA that is assessed into Global Warming Potential (GWP) which is the amount of GHG released during the life cycle of a process. Carbon dioxide is commonly used as a reference gas to compare the impact of various greenhouse (Shine, 2009). Fossil fuel consumption contributes approximately 65% of GHG emissions (Environmental Protection Agency (EPA), 2019). Therefore, renewable energy is expected to address this concern.

2.5 Revenue Value

As an emerging renewable energy source in the United States, anaerobic digestion provides various forms of revenues from renewable electricity, digestate, energy or fuel credits, and feeding fees. Nevertheless, the prices vary depending on location, operating procedures, and state regulations. One example is tipping fees, which are rates that a company should pay for disposing of waste materials as digester input. Tipping fees depend on the water content of materials. Fats, Oils, and Grease (FOG) have a monetary value of \$0.10 per gallon due to high water content. However, the price can be about \$0.05 per gallon if it has a lower water concentration or as dry material. On the other hand, digestate can be traded to farms or composting facilities for approximately \$7.00 per ton (Dr. Dana Kirk, 2020).

Other potential revenues are RINs and RECs as mentioned previously. RIN price ranges between \$0.01 and \$3.50 depending on the biofuel source. RECs prices are various following changes in policy and the availability of RECs within the state. Currently, most RECs are fulfilled through sources such as wind and solar which creates a deviation in the market for prices.

In Michigan itself, RPS increased from 10% in 2015 to 15% in 2021, while it has compliance requirements of 12.5% in 2019 and 2020 (Michigan Public Service Commission, 2021). REC is regulated by The Michigan Renewable Energy Certification System (MIRECS) as a tracking and certification system. According to Consumers Energy, REC price is \$0.014 per kWh or \$14/MWh. Additionally, Michigan's electric providers in total retired 12,812,152 RECs in 2019 (Scripps et al., 2016).

CHAPTER 3. MATERIAL AND METHODS

3.1 Manure Feedstock

The South Campus Anaerobic Digester (SCAD) digester uses various kinds of feedstock, including dairy, beef, waste feed, poultry, and swine manure. Most feedstocks are from the Michigan State University dairy farm which is located adjacent to the SCAD.

3.2 Food Waste Feedstock

Food waste used in the digester included pineapple, pulp, FOG (fat, oil, and grease), and waste feed. A small portion of the food waste comes from Michigan State University dining halls, while the rest comes from off-campus food processors and manufacturers. Most of the food waste comes from southern Michigan, but a small portion of food waste was also gained from the neighboring states such as Indiana and Ohio.

3.3 Real-time (commercial system) sampling methods

The process flow diagram shown in Figure 3.1 provides a better view of the sampling locations.

3.3.1 Gas quantity and quality, and electricity production

Gas quantity was measured by using Endress Hausser Proline t-mass 65 flow meter in real time 24 hours as CHP (Combined Heat and Power) daily total in standard cubic foot (SCF) unit. Gas quality was measured by using AwiFLEX Cool+ gas analyzer every hour. Gas was taken from 2 distinct locations which were before and after the carbon activator scrubber. Three gases were measured which were CH₄ (%), O₂ (%), and H₂S (ppm). Electricity was measured as daily electrical power generated in kilowatts hour (kWh) unit by using SATEC PM172E. The measurement was recorded by the software in the morning until May 2015, then it changed to be in the midnight afterwards.



Figure 3.1. Sampling and measurement maps (Source: MSU Anaerobic Research and Education Center)

(Notes: black line: feedstock; yellow line: separated solids; blue line: filtrate; green dashed line: biogas phase; orange line: sample collection and measurement)

3.3.2 Digestate, filtrate, and solids sampling

Digestate or effluent from the digester consists of a mix of cow manure from MSU Dairy Farms and food waste. The effluent was collected from a piping line that connects the digester to the solid separator. The effluent was taken before the slurry went through a solid separator. The ratio between cow manure and food waste changes daily based upon material received. The filtrate was collected after solid liquid separation from a line that connects the solid separator to the digestate holding tank (DHT). Solids were taken at the solid separator station. These samples were regularly sent to ADREC for laboratory analysis. The frequency of sampling varied depending on the workforce availability in the laboratory. However, in general, the sampling was at least once a month.

3.3.3 Temperature sampling

The temperature of digester was measured continuously by using Endress Hausser T13 RTD probe that is located about 3 feet off the floor on the southwest area of the digester. The measurement results were recorded by the Allen Bradley/Rockwell compact logix PLC processor every 10 minutes and put into a CSV file. The measurement results were also recorded by the digester manager twice a day in the morning and at the end of working hours.

3.3.4 pH sampling

The pH of slurry in the digester was measured continuously by using Endress Hauser Liquiline C CM42, which is in the pipe that goes through a heat exchanger. The measurement results were recorded by the Allen Bradley/Rockwell compact logix PLC processor every 10 minutes and put into a CSV file. The measurement results were also recorded by the digester manager twice a day in the morning and at the end of working hours.

3.4 Laboratory Analysis

The laboratory analysis for SCAD samples was done at the MSU Anaerobic Digestion Research and Education Center (ADREC).

3.4.1 Total Solids and Volatile Solids

Total and volatile solids (TS and VS) are the fundamental feedstock and digestate measure for SCAD operational management. In addition to TS and VS, this test also yields information regarding the sample moisture content (MC) and fixed solids (FS). TS and VS analysis were performed following the EPA accepted Hach methods 8271 and 8276, respectively. The procedure for TS was modified from a 6-hour oven holding time to 24 hours to ensure complete drying. The procedure for VS was also adjusted by increasing the time from 1 hour to 6 hours to ensure complete sample combustion.

Materials and equipment needed for the test included the digester's samples (filtrate, effluent, and solids), 50 mL glass beakers (3 per sample), laboratory analytical balances (Scientech SA 120), oven (Precision Scientific, Catalog No. 31578-10), furnace (Lindberg, Model No. CBFM516C), stir plate (Cole-Parmer Instrument Company, Catalog No. 03406-10), magnetic stir bar, desiccator (Boekel 1342), desiccant (Drierite 22001), syringe, spoon, marker, and white board. *3.4.2 Chemical Oxygen Demand*

Chemical Oxygen Demand (COD) is used as a measure of pollutant in wastewater or effluent. The results indicate the concentration of pollutants in the sample. The higher COD content, the more polluted the sample. The test was performed following the EPA accepted Hach method 8000. Materials and equipment needed were diluted samples using DI water, HACH heated reactor DRB 200, HACH spectrophotometer DR 5000, HACH COD test vials (Catalog No. 2125915), blank test vial, stir bar and stir plate (Cole-Parmer Instrument Company, Catalog No. 03406-10), micropipette, microtips, and delicate wipes.

3.4.3 Total Suspended Solids / Volatile Suspended Solids

Total Suspended Solids / Volatile Suspended Solids (TSS/VSS) test is a method to determine the amount of total suspended solids and total volatile suspended solids found within a sample. TSS/VSS tests were performed following the EPA accepted Hach methods 8158 and 8164, respectively. Materials and equipment needed were laboratory analytical balances (Scientech SA 120), desiccator, oven (Precision Scientific, Catalog No. 31578-10), furnace (Lindberg, Model No. CBFM516C), stir plate (Cole-Parmer Instrument Company, Catalog No. 03406-10), magnetic stir

bar, vacuum filtration system, glass-microfibre discs filter 47 mm (HACH, Catalog No. 253000), tweezers, watch glass, aluminum crucible, DI water, 1 ml syringe, and tongs.

3.4.4 pH and Electrical Conductivity

pH and electrical conductivity (EC) are key measurements to monitor the biological health of a digester hence liquid samples of the digester's filtrate and effluent were evaluated weekly. The recommended range for pH is between 6.4 and 8.2 for a healthy digester (MSU ADREC Operator Training, 2019). Prior to use, the pH probe was calibrated using three calibration solutions at pH 4.01, 7.00, and 10.00. After calibration, pH, and electronic conductivity (EC) probes were rinsed with DI water and wiped using delicate wipes. Sample was stirred on the stir plate using magnetic stirrer; then, the probes were dipped into sample. Both probes were rinsed and wiped after use and the pH probe was stored in a storage solution.

Materials and equipment needed were pH/Conductivity meter (Orion Star A215), pH probe (Orion 8157BNUMD), conductivity probe (Orion 013005MD), double ionized (DI) water, calibration solutions for pH 4.01 (Millipore Sigma BX1634), 7.00 (Millipore Sigma 7BX1635), and 10.00 (Millipore Sigma BX1642), delicate wipes, stir bar, stir plate (Cole-Parmer Instrument Company, Catalog No. 03406-10), magnetic collection stick, and empty beakers.

3.4.5 Alkalinity and Volatile Fatty Acids

Alkalinity and Volatile Fatty Acids (VFA) tests were conducted using the titration method to understand the susceptibility of the digester towards the change within its internal environment. Alkalinity and VFA tests were conducted by referencing O'Brien and Donlan (1977) methods, as follows:

• Samples were centrifuged for 30-40 minutes

- After centrifugation, liquids were filtered using 23 µm then 11 µm filters to gain 50 mL of liquids only
- Filtered samples were then poured into 150 ml and stirred using stir bar on the stir plate
- pH was measured using pH probe prior to titration to know the initial pH
- Samples were gently stirred during titration using 1.0 N H₂SO₄ to a pH of 3.3, then the volume reading is noted. All processes up to this point were part of the alkalinity test
- After the first titration, sample beaker was covered with 65 mm watch glass and then heated on the heated stir plate to a gentle boiling point for 3 minutes
- The sample was then cooled down to room temperature. After that, the watch glass was rinsed into the beaker with DI water
- Sample was then titrated again using 0.05 N NaOH to pH 4.0, then volume reading was noted
- Lastly, sample was titrated again using 0.05 N NaOH to pH 5.1 without refilling the solution from the previous titration, then the volume reading is totaled with the second titration. This entire process was part of the VFA test

Materials and equipment needed were centrifuge (Hermle Labnet Z 206 A), centrifuge vials 50 mL, 150 ml sample beaker, empty beaker, 23 μm (Whatman, Catalog No. 1441-047) and 11 μm filters (Whatman, Catalog No. 1001-047), filter flask (brand), tweezers, pH/Conductivity meter (Orion Star A215), pH probe (Orion 8157BNUMD), delicate wipes, stir bar, heated stir plate (Cole-Parmer Catalog No. EW-03407-36), cone hood, stir plate (Cole-Parmer Catalog No. 03406-10), 65 mm watch glass, double ionized (DI) water, 1.0 N H₂SO₄, and 0.05 N NaOH.

3.4.6 Total Nitrogen

Nitrogen is an important nutrient to both digester performance and fertilizer use of digestate. Total Nitrogen (TN) is a test to determine the total nitrogen by the per sulfate digestion method. The test was performed following the instructions available for the HACH TNT 827 test kit. Materials and equipment used were dilution of samples for range used, HACH heated reactor DRB 200, HACH nitrogen test kit high range (TNT 827), HACH spectrophotometer DR 5000, 20 mm reaction tube, stir bar and stir plate (Cole-Parmer Instrument Company, Catalog No. 03406-10), glass beakers, double ionized (DI water), delicate wipes, micropipette, and microtips.

3.4.7 Ammonia

Ammonia can cause toxicity in digesters if levels exceed 3,000 mg/L (MSU ADREC Operator Training, 2019). An ammonia test was conducted to analyze the ammonia content in the samples. The test was performed following the instructions available for the HACH TNT 832 test kit. Materials and equipment needed were diluted samples using DI water, HACH ammonia test kit high range (TNT 832), HACH spectrophotometer (DR 5000), stir bar and stir plate (Cole-Parmer Instrument Company, Catalog No. 03406-10), micropipette, microtips, and delicate wipes. *3.4.8 Total Phosphorus*

Phosphorus is the key limiting nutrient for land application of digestate as a fertilizer. In organic wastes, phosphates are present in organic and condensed inorganic forms. Phosphorus can be obtained from treating the samples with acid and heat providing conditions for hydrolysis of the condensed inorganic forms. Total Phosphorus (TP) test was conducted to measure the phosphorus content in filtrate and effluent samples. The test was performed following the instructions available for the HACH TNT 844 test kit. Materials and equipment needed were diluted samples using DI water, HACH heated reactor DRB 200, HACH phosphorus test kit high

range (TNT 844), HACH spectrophotometer DR 5000, stir bar and stir plate (Cole-Parmer Instrument Company, Catalog No. 03406-10), micropipette, microtips, and delicate wipes.

3.4.9 Other laboratory tests

There were several tests conducted outside of ADREC laboratory due to facilities availability. Those tests were conducted by A&L Great Lakes Laboratory (algreatlakes.com) in Fort Wayne, Indiana. The tests included manure nutrition analysis which were moisture, solids, Total Kjeldahl Nitrogen (TKN), phosphorus, potassium, sulfur, calcium, magnesium, sodium, iron, aluminum, manganese, copper, and zinc. All test methods were referred to Recommended Methods of Manure Analysis, UW A3769, summarized in Table 3.1 below.

Parameter	Method
Moisture	UW A3769 III.2
Solids	Solids were calculated from
	moisture
Total Kjeldahl Nitrogen (TKN)	UW A3769 III.3.2
Phosphorus, Potassium, Sulfur, Magnesium, Calcium,	UW A3769 III.6.3. All minerals ran
Sodium, Aluminum, Copper, Iron, Manganese, Zinc	on Thermo iCAP 6500.

 Table 3.1 List of test methods done by A&L Great Lakes Laboratory

3.5 Statistical Analysis

Three analyses were conducted for this research: mass and energy balance, life cycle impact assessment, and economic analysis. The original data from SCAD manager were reorganized to ease data calculation and further analysis. Data was organized in an Excel spreadsheet and used to calculate the descriptive analysis such as min, max, mean, average, standard deviation, number of samples, and coefficient of variation. Other spreadsheets were created from the reorganized spreadsheet to provide data sources for statistical analysis by using the R Studio programming software.

Multi linear regression (MLR) in R Studio programming software was used to determine the model for gas production from available feedstock combinations. The codes utilized for MLR analysis are provided in Appendix A.

MLR formula is defined as equation below:

 $egin{aligned} y_i &= eta_0 + eta_1 x_{i1} + eta_2 x_{i2} + ... + eta_p x_{ip} + \epsilon \ \mathbf{where, for } i &= n \ \mathbf{observations:} \ y_i &= ext{dependent variable} \ x_i &= ext{explanatory variables} \ eta_0 &= ext{y-intercept (constant term)} \ eta_p &= ext{slope coefficients for each explanatory variable} \ \epsilon &= ext{the model's error term (also known as the residuals)} \end{aligned}$

Furthermore, R Studio programming software was also used to create radar and violin charts to check data distribution of each feedstock year to year. A radar chart, also known as a spider plot, is used to visualize the amount of feedstock received by the digester each month and year. Radar chart was created by using fmsb and ggradar packages in R software. The codes utilized for Radar chart are provided in Appendix B.

(2)

A violin chart is used to visualize the distribution of individual feedstock amount year to year, also output parameters of the digester. Violin chart is created by using ggplot2 and geom_violin packages in R software. ANOVA (analysis of variance) Tukey multiple comparison was used to determine statistically significant differences between the various operational parameters via the R function TukeyHSD. The codes utilized for Violin chart and ANOVA Tukey multiple comparison are provided in Appendix C.

4.1 Feedstock Amount

During the operation period January 1, 2014, and ending December 31, 2020, the SCAD utilized 18 different feedstocks. Feedstocks were managed in two reception tanks, one for manure (low energy materials) and another for food waste (high energy materials). Table 4.1 summarizes the feedstocks and reception tank used to manage the inflow. The composition of feedstock received by the digester is variative each year.

No	Manure Pit	Food Pit
1	Digestate (recycle)	Filtrate (recycle)
2	Filtrate (recycle)	Cart Food
3	ANS Other	Fat, Oil, and Grease (FOG)
4	Beef Manure	Other
5	Dairy Gutter Manure (Dairy G.)	Pineapple (P.A.)
6	Dairy Freestall Manure (Parlor)	Pulp
7	Poultry Manure	SLS Solids
8	SLS Solids	Waste Feed
9	Swine Manure	
10	Waste Feed	

Table 4.1 Feedstocks in SCAD Manure Pit and Food Pit

4.1.1 Yearly total

The yearly total of feedstock processed in SCAD is shown in Table 4.2 for manure pit and Table 4.3 for food pit. Digestate is recycled directly from the digester effluent without solid-liquid separation. Filtrate is the liquid generated after coarse solids are separated from digestate. Digestate and filtrate are recycled to thin out pits and to dilute high TS feedstocks. SLS Solids are

	Feedstock (metric ton)									
Year	Digestate	Filtrate	SLS Solids	Dairy G.	Parlor	Beef	Waste Feed	Poultry	Swine	ANS Other
2014		5,643		4,104	4,273	697	17		25	4
2015	311	3,706	192	4,542	4,754	121	3		144	30
2016	1,001	412	56	5,091	4,200	243	34	5		17
2017	656	171	229	4,878	4,787	222	45	13	78	29
2018	296	179	42	4,855	5,110	220			151	5
2019		94	50	4,885	5,877	391	33	20		3
2020	23	45	59	4,664	5,388	7	56	15		75
Max	1,001	5,643	229	5,091	5,877	697	56	20	151	75
Min	23	45	42	4,104	4,200	7	3	5	25	3
Mean	268	380	82	4,707	4,882	160	23	12	81	13
Average	457	1,465	105	4,717	4,913	272	31	13	100	23
St. Dev	378	2,266	83	322	599	221	19	6	59	26
Coefficient of Variation	141%	597%	101%	7%	12%	139%	83%	54%	73%	191%

Table 4.2 Total Feedstock in Manure Pit Year 2014-2020

the coarse fiber separated from digestate via solid-liquid separation. SLS Solids are added solids back into the digester to thicken thin, low TS, feedstocks. ANS stands for animal science, meaning that there were research materials added into the digester in low quantities, for example eggs from research chickens.

There were several years where the digester did not receive certain types of feedstocks. SLS Solids were not used in 2014. Poultry manure did not come to the digester in 2014, 2015, and 2018. Furthermore, there was no swine manure in 2016, 2019, and 2020. Pineapple was only available until 2016. Waste feed was absent in 2018.

Several feedstocks, Dairy Gutter, and Parlor as well as FOG, produced the largest annual mass; meanwhile, the others had significant changes in a certain period. Moreover, each feedstock has a different peak of receiving in the digester. According to manure pit data (Table 4.2), the digester received the highest amount of SLS Solids in 2017; Dairy G in 2016; Parlor in 2019; Beef in 2014; Waste Feed in 2020; Poultry in 2019; and Swine in 2018. The average of each feedstock received by SCAD during 2014-2020 is 105 metric tons for SLS Solids, 4,717 metric tons for Dairy Gutter, 4,913 metric tons for Parlor, 272 metric tons for Beef, 31 metric tons for Waste Feed, 13 metric tons for Poultry, 100 metric tons for Swine, and 23 metric tons for ANS Other.

Regarding the coefficient variation of feedstock, it showed that Dairy Gutter and Parlor had the most consistent amount that the digester received year by year. The coefficient variation for Dairy G and Parlor was 7% and 12% respectively, showing that the number did not significantly change year over year (Figure 4.1). These two feedstocks were also the major feedstocks in manure pit. Based on calculation of Dairy Gutter and Parlor in the total of feedstocks received in manure pit, the percentage of these feedstocks increased every year. Dairy G and Parlor contributed 57% of total feedstock in manure pit in 2014, 67% in 2015, 84% in 2016, 87% in 2017,

92% in 2018, 95% in 2019, and 97% in 2020. It shows that the digester improves the feedstock composition which affects gas production. On the other hand, Filtrate significantly decreased since 2016 due to the conclusion of pineapples waste reception. The primary reason for the addition of filtrate and SLS Solids to the reception pits is to create suitable feedstock for pumping and mixing, TS less than 8%. In 2016 and prior, the digester received a vast amount of pineapples waste that required excessive amount of filtrate to dilute and breakdown the waste before pumped to the mix tank.

Meanwhile, on food pit data (Table 4.3), the digester received the highest amount of SLS Solids in 2019, Pineapple in 2014, Pulp in 2014, FOG in 2020, Waste Feed in 2015, Other Feedstock in 2015, and Cart Food in 2015. The average amount of feedstocks received by the digester in 2014-2020 was 152 metric tons for SLS Solids, 2,098 metric tons for Pineapple, 78 metric tons for Pulp, 8,380 metric tons for FOG, 18 metric tons for Waste Feed, 388 metric tons for Other Feedstocks, and 145 metric tons for Cart Food.

Regarding coefficients of variation, there was no feedstock in the food pit that was as consistent as Dairy G or Parlor in manure pit. FOG was the most consistent followed by Pulp and Pineapple, with coefficients of variation of 31%, 38% and 44%, respectively (Figure 4.2). Pulp feedstock which comes from Brody dining hall experienced a great reduction in 2020 due to COVID-19 pandemic that limited students' activities in the residence halls after March 15, 2020. Moreover, the Pineapple contract with the digester was only until 2016. The supplier proceeded to compost the feedstock afterwards, hence there were no more supplies starting in 2017. Food carts also experienced a significant reduction starting in 2017 due to contamination with "debris." Overall, FOG dominated feedstock supplies in the food pit.

	Feedstock (metric ton)							
Year	Filtrate	SLS Solids	Pineapple	Pulp	FOG	Waste Feed	Other	Cart Food
2014	1,321		2,681	102	4,005	24	117	283
2015	1,804	53	2,493	94	7,792	28	171	366
2016	1,312	271	1,118	80	8,470		154	320
2017	577	85		77	8,330		45	15
2018	201	196		80	8,068	2	44	14
2019	194	275		92	9,875		92	12
2020	256	35		20	12,122		2,094	3
Max	1,804	275	2,681	102	12,122	28	2,094	366
Min	194	35	1,118	20	4,005	2	44	3
Mean	564	115	1,955	71	8,016	10	142	43
Average	809	152	2,098	78	8,380	18	388	145
St. Dev	660	109	853	27	2,444	14	754	169
Coefficient of Variation	117%	94%	44%	38%	30%	138%	530%	396%

Table 4.3 Total Feedstock in Food Pit Year 2014-2020



Figure 4.1 Comparison of feedstock received by SCAD in manure pit



Figure 4.2 Comparison of feedstock received by SCAD in food pit A ratio of total feedstock between manure pit and food pit was calculated. Based on

Table 4.4, the ratio of feedstock from food pit ranged from 37% to 58% with the highest percentage being in 2020. On average, SCAD has received 11,897 metric tons of feedstock in manure pit each

year and 10,838 metric ton of feedstock in food pit, for a total of 22,375 metric ton/year with the average ratio of feedstock in food pit was 47%.

	Manure Pit	Food Pit	Total	Food						
Year	(Metric	(Metric	(Metric	Pit						
	ton/year)	ton/year)	ton/year)	(%)						
2014	14,763	8,533	23,297	37%						
2015	13,805	12,800	26,605	48%						
2016	11,059	11,726	22,785	51%						
2017	11,109	9,129	20,238	45%						
2018	10,859	8,605	19,464	44%						
2019	11,353	10,539	21,893	48%						
2020	10,332	14,531	24,863	58%						
Max	14,763	14,531	26,605	58%						
Min	10,332	8,533	19,464	37%						
Mean	11,802	10,637	22,618	47%						
Average	11,897	10,838	22,735	47%						
St. Dev	1,683	2,294	2,499	7%						

 Table 4.4 Ratio of Manure Pit and Food Pit

4.1.2 Monthly Average

The monthly average of feedstock processed in SCAD is shown in Table 4.5 and 4.6. FOG, Parlor, and Dairy Gutter are three major feedstocks that the digester received each month during the years 2014-2020. Radar charts in Figure 4.3 to 4.9 also give a better picture of feedstock peaks for each month. In 2014 to 2016, Parlor, FOG, and Filtrate consistently had higher peaks compared to the rest of feedstocks. However, from 2017 to 2020, Parlor and FOG remained as the only feedstocks that have consistent peaks of supply. There were several outliers that happened during 2014-2020, which were Digestate (July 2016), Swine (December 2017), Filtrate in manure pit (August 2017 and May 2018), Filtrate in food pit (July 2018), and FOG (October 2019).

	Feedstock (kg)									
Year	Digestate	Filtrate	SLS Solids	Dairy G.	Parlor	Beef	Waste Feed	Poultry	Swine	ANS Other
2014		25,840		11,463	25,466	6,354	4,069		12,645	179
2015	20,592	21,259	2,916	12,647	27,034	4,784	1,069		4,537	741
2016	18,979	12,123	3,263	13,844	21,659	3,914	3,426	568		441
2017	15,338	14,884	3,187	13,307	20,612	2,394	1,451	1,071	25,682	919
2018	16,210	15,258	2,749	13,205	23,907	2,878			17,275	688
2019		10,184	3,159	13,287	21,993	3,432	1,847	1,935		1,562
2020	11,259	12,507	3,637	12,678	21,269	3,053	3,159	1,950		3,022
Max	20,592	25,840	3,637	13,844	27,034	6,354	4,069	1,950	25,682	3,022
Min	11,259	10,184	2,749	11,463	20,612	2,394	1,069	568	4,537	179
Mean	16,136	15,259	3,139	12,899	23,031	3,648	2,239	1,231	12,631	779
Average	16,475	16,008	3,152	12,919	23,134	3,830	2,503	1,381	15,035	1,079
St. Dev	3,218	5,173	279	705	2,222	1,255	1,105	589	7,653	888
Coefficient of Variation	20%	34%	9%	5%	10%	34%	49%	48%	61%	114%

 Table 4.5 Average Mass Delivered to Digester Each Month (Manure Pit), 2014-2020
	Feedstock (kg)								
Year	Filtrate	SLS Solids	Pineapple	Pulp	FOG	Waste Feed	Other	Cart Food	
2014	19,922		11,547	1,112	18,952	2,017	10,870	2,151	
2015	18,478	3,032	11,093	1,272	35,270	2,146	2,849	1,845	
2016	14,955	3,386	11,345	1,317	35,052		1,341	2,417	
2017	18,812	2,844		2,013	32,703		1,067	121	
2018	22,917	2,802		1,656	31,780	1,620	1,498	117	
2019	14,565	3,228		2,088	34,277		2,857	92	
2020	14,102	3,160		1,630	40,428		13,222	67	
Max	22,917	3,386	11,547	2,088	40,428	2,146	13,222	2,417	
Min	14,102	2,802	11,093	1,112	18,952	1,620	1,067	67	
Mean	17,424	3,069	11,326	1,546	31,917	1,914	3,059	364	
Average	17,679	3,076	11,328	1,584	32,637	1,928	4,815	973	
St. Dev	3,031	207	186	346	6,143	224	4,663	1,020	
Coefficient of Variation	17%	7%	2%	22%	19%	12%	152%	280%	

Table 4.6 Average Mass Delivered to Digester Each Month (Food Pit), 2014-2020



Figure 4.3 Radar chart for feedstock supplies receiveed by SCAD in 2020



Figure 4.4 Radar chart for feedstock supplies receiveed by SCAD in 2019



Figure 4.5 Radar chart for feedstock supplies receiveed by SCAD in 2018



Figure 4.6 Radar chart for feedstock supplies receiveed by SCAD in 2017



Figure 4.7 Radar chart for feedstock supplies receiveed by SCAD in 2016



Figure 4.8 Radar chart for feedstock supplies receiveed by SCAD in 2015



Figure 4.9 Radar chart for feedstock supplies receiveed by SCAD in 2014

4.1.3 Feedstock Distribution

This study also observes the distribution of feedstock received by SCAD for each feedstock from 2014-to 2020. Violin charts were used to check the data distribution of each feedstock year to year. The violin chart consisted of a colored area, a white box, and a straight line with top and bottom points. The colored area represents data points that the parameter has. The more colored the area, the larger the data points. A large part of the colored area shows where most data is located. The white box is in the middle of the colored area. The top part of the box represents the third quartile of data points, a line inside the box represents the median, while the bottom part represents the first quartile. The straight line with top and bottom points represents the range of data with the highest and lowest value, respectively. ANOVA (analysis of variance) Tukey multiple comparisons were used to determine the significance of data compared year to year.

4.1.3.1 Manure Pit

From manure pit, Filtrate manure pit, Dairy Gutter and Parlor are three feedstocks with the largest data points. Figure 4.10 shows the violin chart for Filtrate manure pit. In general, the charts

in 2014, 2015, 2019, and 2020 have distinctive shapes which means the data in these years were distributed differently. Meanwhile, the chart shapes in 2016-2018 show were quite similar. There was a significant range of filtrate circulating into manure pit in 2015 with median and highest data point were about 23,000 kg/day and 40,000 kg/day, respectively. The range got smaller and more consistent in 2016-2018 which the medians were roughly 13,000 kg/day. In 2019 and 2020, data distribution for filtrate supplied to manure pit became more consistent with the range of data points was approximately 5,000 kg/day to 15,000. Lower supply of filtrate shows that digester received fewer solid feedstocks; thus, it did not need much dilution using the filtrate.



Figure 4.10 Data distribution of Filtrate manure pit

Table 4.7 shows the ANOVA Tukey multiple comparison performed for Filtrate manure pit data distribution. It shows that there was a significant difference in the quantity of Filtrate manure pit received by the digester at the p<0.05 level from year to year [F (6, 49) = 3.729, p = 0.00392).

 Table 4.7 ANOVA Tukey Multiple Comparison results for Filtrate manure pit data

 distribution

	Df	Sum Sq	Mean Sq	F Value	Pr (> F)
Year	6	1.533e+09	258,810,814	3.729	0.00392
Residuals	49	3.401e+09	69,412,500		

The violin chart of Dairy Gutter (Figure 4.11) shows different shapes for each year. In 2014, the chart shows an extensive range of data points with most data points located slightly below 12,500 kg/day, while the highest data point was about 16,000 kg/day. There was an outlier in 2017, with data points located around 8,000 kg/day. That means there was a moment where the digester received lower supplies of this feedstock. Meanwhile, feedstock supplies were typically stable in 2018-2020 which the median between 12,500 and 15,000 kg/day.



Figure 4.11 Data distribution of Dairy Gutter

Table 4.8 shows the ANOVA Tukey multiple comparison performed for Daily Gutter data distribution. It shows that there was a significant difference in the quantity of Dairy Gutter received by the digester at the p<0.05 level from year to year [F (6, 77) = 3.554, p = 0.00368).

	Df	Sum Sq	Mean Sq	F Value	Pr (> F)
Year	6	42,372,311	7,062,052	3.554	0.00368
Residuals	77	152,983,389	1,986,797		

 Table 4.8 ANOVA Tukey Multiple Comparison results for Dairy Gutter data distribution

Figure 4.12 shows the data distribution for Parlor. The chart shape was significantly different from year to year. The chart in 2015 shows the most extensive range with the highest peak reaching 40,000 kg/day, becoming the highest data point of all years. Data median also shows

two different trends. Median in 2014, 2015, and 2018 were around 25,000 kg/day, while the rest have median around 20,000 kg/day.



Figure 4.12 Data distribution of Parlor

Table 4.9 shows the ANOVA Tukey multiple comparison performed for Parlor data distribution. It shows that there was a significant difference in the quantity of Parlor received by the digester at the p<0.05 level from year to year [F (6, 77) = 5.017, p = 0.000).

Table 4.9 ANOVA	Tukey Multi	ole Compar	ison results	for Parlo	r data distributio

	Df	Sum Sq	Mean Sq	F Value	Pr (> F)
Year	6	4.214e+08	70,235,648	5.016	0.000
Residuals	77	1.078e+09	14,000,917		

Figure 4.13 shows the data distribution for Total Manure Pit. The chart shows three different trends. Data distribution in 2014 and 2015 share a similar pattern where the curves have clear violin shapes. Data distribution in 2016, 2017, and 2018 share similar shapes where the curves look like a coke bottle. Data distribution in 2019 and 2020 look similar where the curves have shorter shapes, meaning that the data have more consistent ranges. Data median shows a constant slight decrease from 2014 to 2018. There was quite a significant decrease in data median

from 2018 to 2019, then a slight decrease in 2020. Data distribution in 2015 and 2016 shows the most extensive range, from 40,000 to 100,000 kg/day. Data distribution from 2019 to 2020 shows the most consistent data distribution, ranging from roughly 30,000 to 60,000 kg/day.



Figure 4.13 Data distribution for total manure pit

Table 4.10 shows the ANOVA Tukey multiple comparison performed for total manure pit data distribution. It shows that there was a significant difference in the quantity of total manure pit received by the digester at the p<0.05 level from year to year [F (6, 77) = 7.098, p = 0.000).

 Table 4.10 ANOVA Tukey Multiple Comparison results for Total Manure Pit data

 distribution

	Df	Sum Sq	Mean Sq	F Value	Pr (> F)
Year	6	8.058e+09	1.343e+09	7.098	0.000
Residuals	77	1.457e+10	1.892e+08		

Data for other feedstocks in manure pit are provided in Appendix C.1. In general, some of the feedstocks in the manure pit did not have large enough data points, therefore they show little or no colored area around the white box (SLS Solids, Beef and ANS Other). Meanwhile, other feedstocks did not show a complete chart for each year due to the digester did not receiving them in certain years (Digestate, Waste Feed Manure, Poultry, and Swine). From the manure pit, Pulp and FOG were feedstocks with the largest data points. Figure 4.14 shows data distribution for Pulp. The charts show different shapes each year with 2015 being the year with the most consistent data distribution, ranging from roughly 1,000 to 1,600 kg/day. The peak of pulp received by the digester happened in 2019 exceeding 3,000 kg/day. Data distribution in 2020 was typically consistent although the supply was disrupted by COVID-19 pandemic.



Figure 4.14 Data distribution for Pulp

Table 4.11 shows the ANOVA Tukey multiple comparison performed for Pulp data distribution. It shows that there was a significant difference in the quantity of Pulp received by the digester at the p<0.05 level from year to year [F (6, 67) = 8.78, p = 0.000).

	Df	Sum Sq	Mean Sq	F Value	Pr (> F)
Year	6	10,479,022	1,746,504	8.78	0.000
Residuals	67	13,326,784	198,907		

Figure 4.15 shows the data distribution for FOG. Data distribution was typically consistent from 2015 to 2019. The data median in 2014 was the lowest at around 20,000 kg/day, while data median for 2020 was the highest which was slightly above 40,000 kg/day. As previously shown in the radar chart, the year 2019 had an outlier for the data peak which was around 80,000 kg/day, and outlier compared to the rest of data points which ranged from about 25,000 kg/day to 40,000 kg/day.



Figure 4.15 Data distribution for FOG

Table 4.12 shows the ANOVA Tukey multiple comparison performed for FOG data distribution. It shows that there was a significant difference in the quantity of FOG received by the digester at the p<0.05 level from year to year [F (6, 77) = 8.095, p = 0.000).

Tuble fill fill to the full fully fill the comparison results for 1 00 auta distribution	Table 4.12 ANO	VA Tuke	y Multiple Co	omparison	results for	FOG data	distribution
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	Df	Sum Sq	Mean Sq	F Value	Pr (> F)
Year	6	3.215e+09	535,833,184	8.095	0.000
Residuals	77	5.097e+09	66,195,896		

Figure 4.16 shows the data distribution for Total Food Pit. The chart in 2020 shows a perfect upside-down violin, while 2014 looks like a violin shape. Data distribution in 2015, 2016, and 2018 share similar chart shapes. The highest data point happened in 2019 which was around

90,000 kg/day, while the lowest data point happened in 2017 which was around 20,000 kg/day, associated with the engine outage. The data range from 2014 to 2016 was typically consistent from slightly above 40,000 kg/day to slightly above 80,000 kg/day, meanwhile the data range from 2018 to 2020 was typically consistent with most data points were approximately between 30,000 kg/day and 80,000 kg/day.



Figure 4.16 Data distribution for Total Food Pit

Table 4.13 shows the ANOVA Tukey multiple comparison performed for Total Food Pit data distribution. It shows that there was a significant difference in the quantity of FOG received by the digester at the p<0.05 level from year to year [F (6, 77) = 3.641, p = 0.003).

Table 4.13 ANOVA Tukey Multiple Comparison results for Total Food Pit data distribution

	Df	Sum Sq	Mean Sq	F Value	Pr (>F)
Year	6	4.559e+09	759,802,790	3.641	0.003
Residuals	77	1.607e+10	208,668,601		

Data for other feedstocks in food pit are provided in Appendix C.2. In general, some of the feedstocks in the manure pit did not have large enough data points, therefore they show little or no colored area around the white box (Filtrate Food Pit, SLS Solids Food, Other, and Cart Food).

Meanwhile, other feedstocks did not show a complete curve for each year due to the digester did not receiving them in certain years (Pineapples and Waste Feed).

4.1.3.3 Total Feedstock (Manure Pit and Food Pit)

Figure 4.17 shows the data distribution for Total Feedstock. Data distribution charts in 2015 and 2018 look similar but in a different data range, while the rest of the charts look quite distinctive in shape. Data distribution in 2014 and 2017 were typically similar, ranged approximately from 95,000 kg/day to 155,000 kg/day. The highest data point happened in 2016, while the lowest data point happened in 2019. Data curves in 2016 and 2019 show the most extensive distribution but in a different range. Data distribution in 2018-2020 shows quite a similar range, which means total feedstocks received by the digester were considered consistent in the last three years.



Figure 4.17 Data distribution for Total Feedstock

Table 4.14 shows the ANOVA Tukey multiple comparison performed for Total Feedstock data distribution. It shows that there was a significant difference in the quantity of Total Feedstock received by the digester at the p<0.05 level from year to year [F (6, 77) = 5.736, p = 0.000).

	Df	Sum Sq	Mean Sq	F Value	Pr (> F)
Year	6	1.47e+10	2.451e+09	5.736	0.000
Residuals	77	3.29e+10	4.273e+08		

 Table 4.14 ANOVA Tukey Multiple Comparison results for Total Feedstock data

 distribution

4.1.4 Feedstock Characterization

Table 4.15 shows the characterization of feedstocks received by SCAD during 2014-2020. Based on the data available at the digester, there are only seven feedstocks that have laboratory results for characterization: Parlor Manure, Beef, Dairy Gutter, FOG, Food Other, Pineapple, and Pulp. Characteristics measured were total solids (TS), volatile solids (VS), pH, electron conductivity (EC), soluble COD (sCOD), total nitrogen (TN), total phosphorus (TP), and ammonia. The data on table represents average, standard deviation, and number of samples in the brackets.

According to several samples, TS, VS, pH, and EC were the parameters which were measured the most for Parlor Manure, FOG, and Food Other feedstocks. For Parlor Manure, data collected was 11 for TS, VS, and pH, while it had 10 data for EC. They were taken from 2015 to 2018. Characterization data of Parlor Manure were 63,844±20,998 mg/L, 52,742±18,966 mg/L, 7.01±0.30, and 13.72±1.37 mS/cm for TS, VS, pH, and EC, respectively. Data collected for FOG were 28 for TS and VS, 27 for pH, and 23 for EC. They were taken from 2014 to 2019. Characterization data of FOG were 120,191±172,277 mg/L, 105,384±142,739 mg/L, 5.50±1.51, and 12.17±38.04 mS/cm for TS, VS, pH, and EC, respectively. Standard deviation for FOG samples was large due to the wide range of laboratory results. For example, TS ranged from 1,520 to 689,323 mg/L, VS ranged from 532 to 542,778 mg/L, and EC ranged from 0.95 to 189.90

mS/cm. The sample variation is driven largely by the grease interceptor management and the material collection practices.

Meanwhile, data collected for Food Other were 24 for TS and VS, 23 for pH, and 22 for EC. They were taken from 2014 to 2019. Characterization data of Food Other were 219,447±286,460 mg/L, 193,795±265,043 mg/L, 5.41±1.45, and 8.52±6.03 mS/cm for TS, VS, pH, and EC, respectively. Similar to FOG, standard deviation of Food Other feedstock characterizations was large due to significant difference between each data point. For example, TS ranged from 24,072 to 950,355 mg/L, VS ranged from 18,045 to 933,805 mg/L, pH ranged from 3.58 to 7.88, and EC ranged from 0.31 to 18.53 mS/cm.

Other feedstocks and parameters were collected at a minimum data point. For example, the digester only has 1 data point for Dairy Gutter and Pulp, which was taken in 2016 and 2014, respectively. Beef has 3 data points for TS, VS, pH, and EC yet does not have any data for SCOD, TN, TP, and ammonia. These data points were collected in 2014, 2016, and 2018. Characterization data of Beef were 462,152±109,098 mg/L, 393,620±100,761 mg/L, 8.64±0.06, and 1.89±0.76 mS/cm for TS, VS, pH, and EC, respectively. Pineapple has 2 data points for TS, VS, pH, and EC; one data point for TN, TP, and ammonia, while it does not have any data for SCOD. These data points were taken in 2014 and 2016. Characterization data of Pineapple were 127,389±14,312 mg/L, 114,749±5,820 mg/L, 3.91±0.04, and 2.26±0.62 mS/cm for TS, VS, pH, and EC, respectively. The lack of workforce becomes the main reason for minimum characterization analysis for SCAD feedstocks. In general, feedstock received by SCAD has significant differences in terms of characterization results.

Feedstock	TS	VS	pН	EC
	(mg/L)	(mg/L)	1	(mS/cm)
Parlor Manure	63,844±20,998 (11)	52,742±18,966 (11)	7.01±0.30 (11)	13.72±1.37 (10)
Beef	462,152±109,098 (3)	393,620±100,761 (3)	8.64±0.06 (3)	1.89±0.76 (3)
Dairy Gutter	162,268 (1)	142,940 (1)	8.23 (1)	7.74 (1)
FOG	120,191±172,277 (28)	105,384±142,739 (28)	5.50±1.51 (27)	12.17±38.04 (23)
Food Other	219,447±286,460 (24)	193,795±265,043 (24)	5.41±1.45 (23)	8.52±6.03 (22)
Pineapple	127,389±14,312 (2)	114,749±5,820 (2)	3.91±0.04 (2)	2.26±0.62 (2)
Pulp	275,459 (1)	262,105 (1)	4.36 (1)	1.49 (1)

Table 4.15 Characterization of SCAD Feedstock ^a

^a data including average \pm standard deviation, and (number of sample)

Table 4.15

Faadataalı	SCOD	TN	TP	NH3
reeusiock	SCOD	(mg/L)	(mg/L)	(mg/L)
Parlor Manure	25,900 (1)	2,190±214 (2)	1,210(1)	932 (1)
Beef	-	-	-	-
Dairy Gutter	-	5,050 (1)	1,386 (1)	1,085 (1)
FOG	109,380±167,383 (4)	15,250 (1)	300(1)	253.75 (1)
Food Other	-	9,139±10,502 (4)	1,032±1,360 (7)	195±363 (7)
Pineapple	-	595 (1)	109 (1)	62.70(1)
Pulp	-	-	-	-

4.2 Biogas and Electricity Production

Table 4.16 below shows the total amount of biogas produced by the digester in the year 2014-2020, Methane (CH₄) and Hydrogen Sulfide (H₂S) average. The highest biogas production was in 2016, reaching 1,418,746 Standard Cubic Meter (SCM). In terms of CH₄ percentage, the years 2018 and 2020 had the highest average which was 66%. The lowest H₂S average was in 2014 at 360 ppm.

	Biogas Total	CH ₄	H ₂ S	Electricity
Year	Biogas Totai	Average	Average	production
	(SCM)	(%)	(ppm)	(kWh)
2014	846,232	63%	360	1,727,073
2015	1,103,695	61%	433	2,118,966
2016	1,418,746	62%	667	1,470,356
2017	1,326,335	65%	421	2,169,693
2018	1,348,024	66%	387	2,680,954
2019	1,280,438	64%	520	2,333,449
2020	1,340,179	66%	652	2,664,665

Table 4.16 Biogas production, methane and H₂S average year 2014-2020

The digester experienced an outage of the CHP engine twice. The first outage occurred from December 2015 to April 2016, when the engine had a rod bearing failure. Consequently, the engine was rebuilt and back to daily operation from May 2016 to November 2016. Furthermore, the engine experienced a second failure with a similar issue. Therefore, CHP was not operating from December 2016 to January 2017. To overcome this condition, the digester installed a new engine in February 2017; thus, the electricity production is running again on the fourth week of February 2017. Therefore, gas production in 2016, as shown in Figure 4.18, experienced a significant decrease compared to the other years. Additionally, biogas produced during the engine

failure solely went to flare to burn the methane and reduce GHG released to the atmosphere; and to the boiler to heat the slurry. Another disturbance happened in 2017 when the feed additive was added to the digester and affected biogas production. July had the lowest production compared to the other months.



Figure 4.18 Correlation of SCAD biogas and electricity over the year

4.3 SCAD Output Data Distribution

Data distribution analysis for SCAD output, including biogas production, CH4 content, H2S content, electricity production, and laboratory analysis of effluent (total solids, total nitrogen, total phosphorus, pH, and VFA). Like feedstock analysis, the violin chart was also used to check the data distribution of each parameter year to year.

Data distribution for biogas production is shown in Figure 4.19. Among all years, biogas production in 2018 has the most consistent chart with a closed data range between 3,500 to 4,000 m3/day, followed by 2020. Meanwhile, biogas production data distribution in 2015 has the largest

range between 700 to 4,700 m3/day. Biogas production in 2017 has an outlier due to feed additive added to the digester during summer. In terms of the median of data, biogas production in 2017, 2018, and 2020 had a similar median of around 3,600 m3/day, while 2019 had a slightly lower median than those three years, which was around 3,400 m3/day.



Figure 4.19 Data distribution for biogas production

Table 4.17 shows the ANOVA Tukey multiple comparison performed for biogas production data distribution. It shows that there was a significant difference in the quantity of biogas produced by the digester at the p<0.05 level from year to year [F (6, 77) = 7.678, p = 0.000).

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	Df	Sum Sq	Mean Sq	F Value	Pr (> F)
Year	6	20,642,104	3,440,351	7.678	0.000
Residuals	77	34,500,609	448,060		

Data distribution for CH₄ content is shown in Figure 4.20. In general, methane content in biogas produced by SCAD ranged from 58% to 69%. Methane content in 2018 has the most consistent distribution, ranging from 64% to 67%. Meanwhile, methane content in 2014 has the

widest distribution, from 60% to 68%. Overall, the methane content is in a good range which biogas typically contains 55-70% of CH₄ (Chynoweth et al., 2001; Somers et al., 2018).



Figure 4.20 Data distribution for CH₄ content

Table 4.18 shows the ANOVA Tukey multiple comparison performed for CH₄ content data distribution. It shows that there was a significant difference in the percentage of CH₄ content at the p<0.05 level from year to year [F (6, 77) = 10.79, p = 0.000).

	Df	Sum Sq	Mean Sq	F Value	Pr (> F)
Year	6	268.5	44.75	10.79	0.000
Residuals	77	319.3	4.15		

Table 4.18 ANOVA Tukey Multiple Comparison results for CH₄ content distribution

Figure 4.21 provides the data distribution for H₂S content. H₂S is a toxic gas which poses safety concerns for people and may result in additional equipment maintenance (Hosseini and Wahid, 2014). H₂S can cause corrosion to stainless steel and copper and nickel alloys. It also can produce H₂SO₃ from reaction with water vapor and form sulfur dioxide (SO₂) from its combustion, leading to metal pipe and engine corrosion and gas leaks. H₂S concentration in biogas ranges between 50 and 10,000 parts per million (ppm), depending on the feedstocks (X. Wang et al.,

2018). H₂S content in SCAD biogas ranging from 3 to 2,251 ppm, which is still in an acceptable range. H₂S content in 2016 has the most extensive range, followed by 2020 and 2019. Meanwhile, H₂S content in 2017 shows the most stable data distribution, ranging from 98 to 799 ppm.



Figure 4.21 Data distribution for H₂S content

Table 4.19 shows the ANOVA Tukey multiple comparison performed for H₂S content data distribution. It shows that there was not any significant difference in H₂S content (p value>0.05) from year to year [F (6, 77) = 0.979, p = 0.445).

Table 4.19 ANOVA Tukey Multiple Comparison results for H₂S content distribution

	Df	Sum Sq	Mean Sq	F Value	Pr (> F)
Year	6	1,130,830	188,472	0.979	0.445
Residuals	77	14,823,077	192,507		

Figure 4.22 provides the data distribution for electricity production per day. In terms of chart shape, data in 2014 and 2018 have quite similar shape but in a different data range. Data distribution for electricity was typically consistent in 2016-2020 with data median ranging between 6,500 to 7,600 kWh/day. Electricity production in 2015 had the highest peak but also the widest

range of all years, which the data ranges from 2,200 to 9,700 kWh/day. Meanwhile, violin curves in 2016 and 2017 show the time when SCAD did not produce electricity due to engine issues, showed by outliers happened in 2017.



Figure 4.22 Data distribution for electricity production

Table 4.20 shows the ANOVA Tukey multiple comparison performed for electricity production data distribution. It shows that there was a significant difference in electricity production at the p<0.05 level from year to year [F (6, 71) = 7.397, p = 0.000).

 Table 4.20 ANOVA Tukey Multiple Comparison results for electricity production

 distribution

	Df	Sum Sq	Mean Sq	F Value	Pr (> F)
Year	6	58,173,817	9,695,636	7.397	0.000
Residuals	71	93,063,092	1,310,748		

Data distribution for effluent TS is shown in Figure 4.23. In general, TS content ranges 40,000 to 85,000 mg/L. Data distribution in 2018 has the most consistent shape, ranging from 53,000 to 65,000 mg/L. Meanwhile, there was an outlier that happened in 2016 where TS content was 85,000 mg/L. TS content in effluent had an increase trend from 2014 to 2017, then became more consistent in 2018-2020 with range from 50,000 to 73,000 mg/L. Data median also increased

from 2014 to 2017 from 50,000 to 68,000 mg/L, then became more consistent in 2018 to 2020 which was around 60,000 mg/L.



Figure 4.23 Data distribution for effluent TS

Table 4.21 shows the ANOVA Tukey multiple comparison performed for effluent TS data distribution. It shows that there was a significant difference in effluent TS at the p<0.05 level from year to year [F (6, 64) = 10.62, p = 0.000).

Table 4.21 ANOVA Tukey Multiple Comparison results for effluent TS distribution

	Df	Sum Sq	Mean Sq	F Value	Pr (> F)
Year	6	2.907e+09	484,541,468	10.62	0.000
Residuals	64	2.919e+09	45,609,925		

Figure 4.24 shows the data distribution for effluent TN. TN data distribution ranges from 1,400 to 5,000 mg/L. Data distribution fluctuated year by year, but it shows a constant decrease from 2016 to 2018. There was an outlier that happened in 2020 where TN was close to 5,000 mg/L. Data median in 2018-2020 were consistent around 3,000 mg/L. TN content in manure was not changed during anaerobic digestion. The process transforms protein and urea nitrogen into inorganic nitrogen such as ammonia gas (NH₃) and ammonium ion (NH₄⁺) (Field et al., 1984).



Figure 4.24 Data distribution for effluent TN

Table 4.22 shows the ANOVA Tukey multiple comparison performed for effluent TN data distribution. It shows that there was a significant difference in effluent TN at the p<0.05 level from year to year [F (5, 44) = 4.072, p = 0.004).

Table 4.22 ANOVA Tukey Multiple Comparison results for effluent TN distribution

	Df	Sum Sq	Mean Sq	F Value	Pr (> F)
Year	5	9,543,738	1,908,748	4.072	0.004
Residuals	44	20,623,980	468,727		

Figure 4.25 shows the data distribution for effluent TP. Data available for TP content was the least compared to the rest of the parameters. TP effluent content ranges from 200 to 2,100 mg/L. There was a small data point in 2015 thus it did not show any charts. TP content was typically consistent in 2017-2019 with data median around 360 mg/L. TP content in manure was not changed during anaerobic digestion process. This process converts organic phosphorus (Org-P) to phosphate (PO4⁺) phosphorus. Most total P losses in digesters were related to solids accumulation in the reactor (Field et al., 1984).



Figure 4.25 Data distribution for effluent TP

Table 4.23 shows the ANOVA Tukey multiple comparison performed for effluent TP data distribution. It shows that there was a significant difference in effluent TP at the p<0.05 level from year to year [F (6, 44) = 9.616, p = 0.000).

Table 4.23 ANOVA Tukey Multiple Comparison results for effluent TP distribution

	Df	Sum Sq	Mean Sq	F Value	Pr (> F)
Year	6	4,015,385	669,231	9.616	0.000
Residuals	44	3,062,221	69,596		

Data distribution for effluent pH is shown in Figure 4.26. In general, SCAD effluent pH ranges from 7.4 to 8.1. Effluent pH in 2017 and 2018 are the most consistent in the range of 7.7 to 8, while pH in 2019 and 2020 are consistent in the range of 7.4 to 7.9. These values are in an acceptable range for pH which is between 6.4 and 8.2 for a healthy digester (MSU Anaerobic Digestion and Research Center, 2019).



Figure 4.26 Data distribution for effluent pH

Table 4.24 shows the ANOVA Tukey multiple comparison performed for effluent pH data distribution. It shows that there was a significant difference in effluent pH at the p<0.05 level from year to year [F (6, 64) = 2.58, p = 0.0266).

Table 4.24 ANOVA Tukey Multiple Comparison results for effluent pH distribution

	Df	Sum Sq	Mean Sq	F Value	Pr (> F)
Year	6	0.333	0.05550	2.58	0.0266
Residuals	64	1.377	0.02151		

Data distribution for effluent VFA is shown in Figure 4.27. VFA acts in the biopolymers of biofuels production such as methane and hydrogen (Lukitawesa et al., 2020; Magdalena et al., 2019). Several factors are crucial for VFA productivity, which are hydraulic retention time (HRT), organic loading rate (OLR), temperature, pH, and pretreatment (Wainaina et al., 2019). Data distribution of VFA in 2018-2020 was relatively in a consistent range, while VFA content in 2016 has the widest range and highest peak, ranging from 2,300 to 7,400 mg/L. Some outliers happened in 2014 which the values were around 3,200 and 2,900 mg/L, while majority of the data were around 1,000 mg/L.



Figure 4.27 Data distribution for effluent VFA

Table 4.25 shows the ANOVA Tukey multiple comparison performed for effluent VFA data distribution. It shows that there was a significant difference in effluent VFA at the p<0.05 level from year to year [F (6, 63) = 7.911, p = 0.000).

 Df
 Sum Sq
 Mean Sq
 F Value
 Pr (>F)

 Year
 6
 66,702,800
 11,117,133
 7.911
 0.000

 Residuals
 63
 88,531,222
 1,405,257
 1405,257

 Table 4.25 ANOVA Tukey Multiple Comparison results for effluent VFA distribution

The acceptable range for VFA is 50-300 mg/L (Schnaars, 2012). Based on the average result of SCAD effluent measurement, VFA content in SCAD effluent ranges from 817 to 3,785 mg/L. These values are higher than the reference. However, VFA/TA ratio is also important to know whether the digester performed at its best ambience. Higher alkalinity values show that the digester has a better capacity to resist pH changes which is associated to the buffering capacity to maintain the digester stability. Alkalinity values in an anaerobic digester range between 1,500 and 5,000 mg/L (Krakat et al., 2017; Schnaars, 2012).

Table 4.26 provides effluent results for VFA and Alkalinity of SCAD during its operation from 2014 to 2020. Alkalinity ranges from 7.583 to 12,779 mg/L, which was higher than the

reference value. After calculating the ratio of VFA and Alkalinity, SCAD had VFA/TA ratio between 0.09 and 0.30. The ideal ratio is between 0.3 and 0.4 in the digester and would be 0.2 to 0.3 in the post-digester (Lossie & Pütz, 2011). However, Schnaars (2012) puts a broader range for VFA/TA ratio between 0.1 and 0.35 to maintain a digester in a well-operated condition. Furthermore, it is always better to maintain the ratio to be below 0.35 for a proper digester operation. Therefore, SCAD performance, in general, is well-operated in terms of VFA/TA ratio. The ratio in 2016 was the only year where SCAD performed in its optimum condition based on both references, while the ratio in 2018 was a bit below standard.

Year	VFA (mg/L)	Alkalinity (mg/L)	VFA/TAC Ratio
2014	1,301	11,007	0.12
2015	1,580	10,477	0.15
2016	3,785	12,779	0.30
2017	1,834	10,413	0.18
2018	817	9,524	0.09
2019	1,187	8,880	0.13
2020	1,378	7,583	0.18

Table 4.26 SCAD effluent VFA and Alkalinity

Data distribution for the digester temperature is available in Figure 4.28. Temperature is an important operational parameter as it provides a suitable ambience for microbial growth. Microbial colonies inside the digester are sensitive to temperature changes. Earlier studies found that digesters should not experience any temperature changes more than 2°C within 24 hours, otherwise the microbial colonies will be highly disturbed (Meegoda et al., 2018; Schnaars, 2012). Based on the temperature setting, an anaerobic digester can be divided into three temperature conditions: psychrophilic (15–23°C), mesophilic (35–41°C), and thermophilic (52–58°C).

Based on data from the digester temperature, SCAD is operating in mesophilic temperature. Diverse studies presented in academics revealed that temperature in the mesophilic

range is considered as the ideal temperature to promote bacterial activity within the digester (Arosemena, 2021). As shown in the Figure 4.28, the digester temperature ranges from 32.5 to 42.5 °C. The highest temperature recorded was in 2015, while the lowest temperature was in 2020. Moreover, the temperature range in 2015 and 2020 was the most significant, while temperature range in 2019 was the most stable.



Figure 4.28 Data distribution for digester temperature

Table 4.27 shows the ANOVA Tukey multiple comparison performed for digester temperature data distribution. It shows that there was a significant difference in digester temperature at the p<0.05 level from year to year [F (6, 77) = 3.277, p = 0.006).

 Table 4.27 ANOVA Tukey Multiple Comparison results for digester temperature

	Df	Sum Sq	Mean Sq	F Value	Pr (> F)
Year	6	55.9	9.317	3.277	0.006
Residuals	77	218.9	2.843		

SCAD experienced a decrease in the digester temperature during winter months each year. Figure 4.29 shows the decrease in digester temperature occurring typically during the November to February timeframe. The digester should be maintained at temperature 35–41°C for mesophilic condition to optimize microbial performance during anaerobic digestion process. However, there were some moments when SCAD temperature went down to be lower than 35 °C, which were in winter 2015, 2019, and 2020. It shows that the digester struggled to maintain the digester temperature in the last two years of this study.

There are two lines added in the figure which represent the highest and lowest temperature in East Lansing to check the gap between air and digester temperatures. It shows that despite the digester experiencing an inconsistent operational temperature lately, it was still above the highest air temperature all the time. Moreover, it emphasizes that more power was needed to heat up the digester during winter since the temperature significantly dropped after September each year. Therefore, heating system is critical for SCAD operational to maintain the mesophilic condition for anaerobic digestion process.



Figure 4.29 Digester temperature average in 2014-2020

4.4 SCAD Laboratory Analysis

This section compares SCAD laboratory analysis for influent and effluent of the digester. Influent samples were obtained from the mix tank before the mixture was pumped inside the digester. Effluent laboratory analysis was done more consistently than influent. Influent analysis was done only until 2018. The main reasons were due to the time and workforce constraints. Therefore, the digester decided to conduct laboratory analysis for filtrate, effluent, and solids only which were done once a month. Another reason to conclude the mix tank sampling was because the digester received feedstocks that had not been changed tremendously from the previous years.

Furthermore, other SCAD laboratory results will be provided in Appendix E.

4.4.1 Solids

Table 4.28 shows the comparison between total solids of influent and effluent. TS reduction ranged from 35% to 48%, with the highest reduction being in 2018. In average, SCAD TS for influent and effluent were 105,405 mg/L and 60,744 mg/L, respectively. Moreover, TS destroyed in average was 45,132 mg/L or 43%.

V		TS	5		TS Destroyed	Reduction	
Year		(mg/	'L)		(ma/I)	0/	
	Influent	n	Effluent	n	(mg/L)	70	
2014	93,359±12,280	12	49,085±5,709	11	44,274	47	
2015	107,615±4,216	4	57,082±8,282	10	50,533	47	
2016	104,673±20,115	7	65,928±7,110	12	38,745	37	
2017	107,881±17,283	3	70,359±5,929	9	37,522	35	
2018	113,498	1	58,913±3,944	12	54,585	48	
2019	-		59,702±6,768	10	-	-	
2020	-		64,138±6,668	7	-	-	
Max	113,498		70,359		54,585	48	
Min	93,359		49,085		37,522	35	
Mean	105,186		60,398		44,654	42	
Average	105,405		60,744		45,132	43	
St. Dev	6,666		6,377		6,603	6	

Table 4.28 Comparison of total solids between influent and effluent ^a

^a average \pm standard deviation

Table 4.29 shows the comparison between volatile solids of influent and effluent. VS reduction ranged from 40% to 54%, with the highest reduction being in 2018. In average, SCAD VS for influent and effluent were 92,803 mg/L and 48,209 mg/L, respectively. Moreover, VS destroyed on average was 44,754 mg/L or 48%.

V		V	8	VS Destroyed		
rear		(mg	$(\mathbf{m} \mathbf{a} / \mathbf{I})$	0/		
	Influent	n	Effluent	n	(mg/L)	70
2014	80,623±10,397	12	38,018±4,767	11	42,605	53
2015	93,101±3,727	4	43,803±7,684	10	49,298	53
2016	90,718±17,274	7	52,485±6,158	12	37,873	42
2017	96,555±16,237	3	58,056±4,682	9	38,499	40
2018	103,020	1	47,524±3,339	12	55,496	54
2019	-		47,354±6,291	10	-	-
2020	-		51,626±5,843	7	-	-
Max	103,020		58,056		55,496	54
Min	80,623		38,018		37,873	40
Mean	92,504		48,029		44,266	48
Average	92,803		48,209		44,754	48
St. Dev	7,366		5,989		6,740	6

Table 4.29 Comparison of volatile solids between influent and effluent ^a

^a average \pm standard deviation

4.4.2 pH and Electronic Conductivity

Table 4.30 shows the comparison between influent and effluent. Influent had a more acidic pH with a range from 5.18 to 6.44. Anaerobic digestion process increased the mixture pH thus the effluent had pH between 7.70 and 7.90. Table 4.31 shows electron conductivity comparison between influent and effluent. Influent has a range of EC from 13.57 to 19.03, while effluent has a range of EC from 14.00 to 20.40.

Year		pH Change			
	Influent	n	Effluent	n	0
2014	6.44±0.32	12	7.90±0.1	11	1.46
2015	5.82 ± 0.84	4	7.86±0.14	10	2.04
2016	5.18±0.67	7	7.80±0.2	12	2.62
2017	5.74±0.51	3	7.80±0.1	9	2.06
2018	5.27	1	7.80±0.1	12	2.53
2019	-	-	7.70±0.1	10	-
2020	-	-	7.70±0.2	7	-
Max	6.44		7.90		2.62
Min	5.18		7.70		1.46
Mean	5.67		7.79		2.10
Average	5.69		7.79		2.14
St. Dev	0.45		0.07		0.42

 Table 4.30 Comparison of pH between influent and effluent ^a

^a average \pm standard deviation

Table 4.31 Comparison of electronic conductivity between influent and effluent ^a

	EC							
Year	(mS/cm)							
	Influent	n	Effluent	n				
2014	13.86±2.07	12	20.00±0.9	6				
2015	13.57±3.63	3	17.00 ± 1.74	10				
2016	15.55±2.01	7	20.40±3.1	9				
2017	19,03±5.86	3	18.00±3.3	9				
2018	9.52	1	17.00±0.8	12				
2019	-	-	16.80±2.3	10				
2020	-	-	14.00±3.3	6				
Max	19.03		20.40					
Min	9.52		14.00					
Mean	13.96		17.48					
Average	14.31		17.60					
St. Dev	3.45		2.00					

^a average \pm standard deviation

4.4.3 Chemicals

Table 4.32 shows the comparison of soluble COD (sCOD) results between influent and effluent. The digester did not do frequent tests for (sCOD) of influent since the results were relatively consistent in 2014. Based on data in 2014 and 2018 which the only years (sCOD) was conducted for influent, (sCOD) reduction in 2014 and 2018 were 29,556 mg/L

and 29,729 ml/L or 79.57% and 76.82%, respectively. It shows that the (sCOD) reduction was consistent.

	sCOD							
Year	(mg/L)							
	Influent	n	Effluent	n				
2014	37,144±7,692	8	7,588±763	2				
2015	-	-	-	-				
2016	-	-	15,418±3,812	10				
2017	-	-	19,075±5,289	4				
2018	38,700	1	8,971±5,855	11				
2019	-	-	6,816±1,525	7				
2020	-	-	-	-				
Max	38,700		19,075					
Min	37,144		6,816					
Mean	37,914		10,641					
Average	37,922		11,574					
St. Dev	778		4,824					

Table 4.32 Comparison of soluble COD between influent and effluent ^a

^a average \pm standard deviation

4.4.4 Fiber Results

Table 4.33 shows the laboratory results for SCAD fiber. In average, total solids and volatile

solids are 264,388 and 235,204, respectively with VS content in TS ranging from 87% to 90%.

NZ	TS	VS	TS	VS	VS	n
rear	(mg/L)	(mg/L)	%	%	% of TS	
2014	238,755	215,338	23.9	21.5	90%	4
2015	264,328	236,833	26.4	23.6	89%	9
2016	249,041	221,187	24.9	22.1	89%	12
2017	277,084	248,754	27.7	24.9	90%	9
2018	268,160	235,912	27.1	23.6	87%	12
2019	268,436	238,752	26.9	23.9	89%	10
2020	284,911	249,653	28.5	25.0	88%	7
Max	284,911	249,653	28.5	25.0	90%	
Min	238,755	215,338	23.9	21.5	87%	
Mean	263,973	234,896	26.4	23.5	89%	
Average	264,388	235,204	26.5	23.5	89%	
St. Dev	14,666	11,948	1.5	1.2	1%	

 Table 4.33 Fiber laboratory results

4.4.5 Other laboratory results

Several laboratory analyses for different parameters were done outside of ADREC. The tests were conducted by A&L Great Lakes Laboratory in Fort Wayne, Indiana. The tests included manure nutrition analysis which were moisture, solids, Total Kjeldahl Nitrogen (TKN), phosphorus, potassium, sulfur, calcium, magnesium, sodium, iron, aluminum, manganese, copper, and zinc. Analysis of filtrate was done in 2017 and 2018; effluent was done in 2018; and solids were analyzed in all year except in 2017. The complete results of A&L Great Lakes Laboratory are provided in Appendix F.

4.5 Discussions

4.5.1 Correlation between feedstock and biogas production

As mentioned in the beginning, SCAD was fed with 18 different feedstocks from 2014 to 2020. Figure 4.30 shows a correlation between food waste amount and biogas production. There were different trends occurring years. Using 2014 as the starting point, 2015 and 2020 showed similar trends where the increase in food waste from the previous year is followed by an increase in biogas production. This trend also happened in 2016 but in the opposite direction. A different trend occurred in 2017 and 2018, where the decrease in food waste affected an increase in biogas production. Meanwhile, in 2019, the trend was reversed: an increase in food waste affected by a decrease in biogas production. The correlation between food waste and biogas production is not clear yet. Therefore, further analysis was done.



Figure 4.30 Correlation between food waste and biogas production

Multi linear regression (MLR) was done to determine specific feedstock that has significant impact on biogas production. There are two steps in determining the most significant parameter. First is Stepwise Procedure which calculates the top list of feedstocks that significantly affects output among all feedstocks available. Second is the Final Model which calculates the most significant factor among the list of feedstocks from Stepwise Procedure.

Among all feedstock, five feedstocks were included as the result of Stepwise Procedure, which are Filtrate in manure pit, Dairy Gutter, Parlor, ANS Other, and Pineapples (PA). The formula generated was Equation 2, while coefficients for each feedstock are provided in Table 4.34.

Biogas production = Filtrate_manure_pit + Dairy_G + Parlor + ANS_Other + P_A (Equation 2)
Parameter	Coefficient
Intercept	2330.91251
Filtrate_manure_pit	-0.01688
Dairy_G	0.18163
Parlor	-0.04628
ANS_Other	0.08850
P_A	-0.02158

Table 4.34 Coefficient of each parameter

The final model was run to determine the key feedstocks in Equation 2 that influence gas production. As shown in Table 4.35, Filtrate in the manure pit, Dairy Gutter, and Parlor are feedstocks that have the most significant impact on biogas production, with a p-value<0.05. However, this statistical analysis result does not mean that the other feedstocks are not important, for example, FOG and other manures. Those are the baseline in building biogas production due to their vast amount in feedstock composition. However, Filtrate in the manure pit, Dairy Gutter, and Parlor provide a critical role in SCAD feedstock matrices in terms of biogas production variation during the operation years. Therefore, food waste and manure are important baselines for biogas production, and it will significantly change once those three feedstocks are added to the feedstock composition.

This MLR result shows Multiple R-squared as 47.15%, and Adjusted R-squared as 43.75%. It means that the model was able to explain 43% to 47% of the data variance to predict the biogas production.

1 01							
	Estimate	Std. Error	t value	Pr(> t)			
(Intercept)	2.331e+03	7.506e+02	3.105	0.002648**			
Filtrate_manure_pit	-1.688e-02	7.464e-03	-2.261	0.026549*			
Dairy_G	1.816e-01	4.507e-02	4.030	0.000129***			
Parlor	-4.627e-02	1.748e-02	-2.648	0.009804**			
ANS_Other	8.850e-02	5.779e-02	1.531	0.129752			
P_A	-2.158e-02	1.501e-02	-1.438	0.154510			
Residual standard error	611.3 on 78 degre	es of freedom					
Multiple R-squared	0.4714						
Adjusted R-squared	0.4375						
F-statistic	13.91 on 5 and 78 DF						
p-value	1.03e-09						
Signif Codes: 0 '***' 0 00	1 '**' 0 01 '*' 0 04	5 • • 0 1 • • 1					

Table 4.35 MLR results for feedstock impact on the biogas production

Signif. Codes: 0 **** 0.001 *** 0.01 ** 0.05 . 0.1 * 1

Table 4.36 presents the result of ANOVA Type II test to support the result of MLR test. Filtrate in manure pit, Dairy Gutter, and Parlor gave a significant impact on biogas production at

the level of p<0.05.

	_			<u> </u>
	Sum Sq	Df	F value	Pr(>F)
Filtrate_manure_pit	1,910,236	1	5.1119	0.0265489*
Dairy_G	6,069,859	1	16.2433	0.0001286***
Parlor	2,619,518	1	7.0100	0.0098044**
ANS_Other	876,194	1	2.3488	0.1297516
P_A	772,430	1	2.0671	0.1545103
Residuals	29,147,290	78		

Table 4.36 ANOVA Type II Results for feedstock correlation with the biogas production

After the first MLR analysis, the data set was treated by using modelDFFITS to take out the data with high leverage or outliers to check whether the R-squared values can be improved. The result shows that Filtrate in manure pit, Dairy Gutter, Parlor, and FOG become four feedstocks that give the most significant impact in biogas production with a p-value<0.05, as shown in Table 4.37. Furthermore, R-squared values improved to be 58.24% and 56.01% for Multiple R-squared and Adjusted R-squared values, respectively. It means that the model was able to explain 56% to 58% of the data variance to predict the biogas production.

This percentage is understandable since SCAD received a wide variety of feedstock with uncertain frequency. Several feedstocks did not come every month; some of them were even only available in certain years. This uncertainty affects the R-squared values. Moreover, feedstock supplies are beyond the digester's control. It depends on whether the supplier still produces the same food waste, for example. However, the study can still draw a conclusion since there are independent variables that are statistically significant to correlate the relationships between the variables.

Table 4.37 MLR results for feedstock impact on the biogas production (after data treatment)EstimateStd. Errort valuePr(>|t|)

	Estimate	Std. Error	t value	Pr(> t)			
(Intercept)	1.804e+03	6.544e+02	2.756	0.007333 **			
Filtrate_manure_pit	-2.207e-02	5.581e-03	-3.953	0.000173 ***			
Dairy_G	1.915e-01	3.994e-02	4.795	8.05e-06 ***			
Parlor	-4.410e-02	1.520e-02	-2.901	0.004880 **			
FOG	1.327e-02	6.526e-03	2.034	0.045470 *			
Residual standard error	509.4 on 75 degre	es of freedom					
Multiple R-squared	0.5824						
Adjusted R-squared	0.5601						
F-statistic	26.15 on 4 and 75 DF						
p-value	1.372e-13						

Signif. Codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

4.5.2 The impact of OLR, HRT, and digester pH and temperature on biogas production and

methane percentage

OLR, HRT, and digester pH and temperatures are key parameters in determining optimum biogas production and methane percentage. Regarding OLR and HRT, SCAD data is organized based on daily numbers, an average of 7 days and an average of 30 days. The formula for determining OLR and HRT is provided in Appendix G. This section only includes data from 2018-to 2020, as the digester performance in those years was considered settled.

Table 4.38 shows data related to OLR in 2018-2020. It shows that data on 30 days average have the smallest standard deviation, which means that the data is the most stable, though the

difference between daily, 7 days average, and 30 days average are not significantly different. In general, the average OLR increased from 2018 to 2020. OLR average values in 2018 are 3.67, 3.65, and 3.59 g VS/L-d for daily, 7 days average, and 30 days average, respectively. In 2019, the values were 3.75, 3.76, and 3.80 g VS/L-d for daily, 7 days average, and 30 days average, and 30 days average, respectively. The values in 2020 are 4.16, 4.14, and 4.16 g VS/L-d for daily, 7 days average, and 30 days average, respectively. Moreover, the table also shows that the standard deviation of OLR in 30 days average in 2018-2020 is very stable, which are 0.50 for 2018 and 2019, and 0.49 for 2020. Therefore, OLR in 30 days average is chosen to represent SCAD OLR. Figure 4.31 shows the OLR in 30 days on average in 2018-2020.

	OLR									
	g VS/L-d									
Month		2018			2019			2020		
WIOIIIII		7 Dovo	30		7	30		7	30	
	Daily	/ Days	Days	Daily	Days	Days	Daily	Days	Days	
		AVG	AVG		AVG	AVG		AVG	AVG	
January	3.37	3.38	3.28	4.41	4.70	4.89	4.36	4.26	4.01	
February	3.40	3.40	3.36	2.69	2.67	3.40	4.56	4.67	4.58	
March	3.67	3.69	3.54	4.10	3.88	3.30	3.94	3.89	4.07	
April	3.41	3.49	3.62	4.11	4.08	4.06	3.84	3.82	3.98	
May	2.85	2.84	3.13	4.33	4.30	4.26	4.09	4.22	4.13	
June	3.72	3.61	3.18	3.97	4.03	4.22	3.30	3.23	3.59	
July	2.97	3.09	3.49	3.63	3.72	3.69	3.88	3.67	3.34	
August	3.23	3.11	2.95	3.44	3.37	3.48	4.40	4.32	4.05	
September	3.75	3.76	3.44	4.11	4.05	3.95	5.04	5.06	4.82	
October	4.10	3.97	3.99	3.86	3.81	3.96	5.01	5.07	5.14	
November	4.53	4.48	4.37	2.86	3.06	3.09	4.12	4.11	4.47	
December	5.00	5.00	4.71	3.52	3.40	3.30	3.34	3.43	3.80	
Average	3.67	3.65	3.59	3.75	3.76	3.80	4.16	4.14	4.16	
St. Dev	0.60	0.58	0.50	0.52	0.53	0.50	0.53	0.56	0.49	

 Table 4.38 Comparison of OLR in 2018-2020: daily, 7 days average, and 30 average



Figure 4.31 OLR in 30 days average 2018-2020

Table 4.39 shows data related to HRT in 2018-2020. It shows that data in 7 days average and 30 days average are relative stable, while daily data has a notable change from 2018 to 2019. HRT in 30 days average has the smallest standard deviation in 2018 and 2019, while HRT in 7 days average has the smallest standard deviation in 2020. HRT average values in 2018 are 21.7, 26.5, and 26.0 days for daily, 7 days average, and 30 days average, respectively. In 2019, the values are 31.2, 25.6, and 23.8 days for daily, 7 days average, and 30 days average, respectively. In 2019, the values in 2020 are 27.3, 21.8, and 20.3 days for daily, 7 days average, and 30 days average, and 30 days average, in 2018-2020 is slightly more stable than 7 days average, which is 2.7 for 2018, 3.5 for 2019, and 4.8 for 2020. Therefore, OLR in 30 days average is chosen to represent SCAD OLR. Figure 4.32 shows the HRT in 30 days average in 2018-2020.

					HRT					
	Day									
Month		2018			2019			2020		
WIOIIII		7	30		7	30		7	30	
	Daily	Days	Days	Daily	Days	Days	Daily	Days	Days	
		AVG	AVG		AVG	AVG		AVG	AVG	
January	24	24	24	29	21	21	20	20	20	
February	28	25	25	71	42	29	19	19	19	
March	23	23	23	26	25	30	30	24	22	
April	21	26	24	27	26	25	28	24	23	
May	20	28	25	18	19	21	20	22	22	
June	24	29	28	20	22	19	27	28	25	
July	20	28	29	28	25	24	28	21	24	
August	25	34	30	35	26	25	18	16	7	
September	18	27	31	20	24	20	25	19	19	
October	20	28	26	35	21	20	18	16	16	
November	21	24	25	40	31	27	27	26	21	
December	18	22	22	26	25	25	67	26	26	
Average	21.7	26.5	25.9	31.2	25.69	23.8	27.3	21.8	20.3	
St. Dev	2.8	3.1	2.7	13.6	5.9	3.5	12.6	3.8	4.8	

 Table 4.39 Comparison of HRT in 2018-2020: daily, 7 days average, and 30 average



Figure 4.32 HRT in 30 days average 2018-2020

MLR analysis was done to determine the operational parameter that has significant impact on the biogas production (Table 4.40) and the methane percentage (Table 4.41). The results show that none of the parameters tested have a significant impact on both two dependent variables. R-squared values for the biogas production analysis was 6.05% and -6.08% for Multiple R-squared and Adjusted R-squared, respectively. Meanwhile, R-squared values for the methane percentage analysis was 9.93% and -1.69% for Multiple R-squared and Adjusted R-squared, respectively. It means that the model was barely able to explain any of the data variance to predict the biogas production and the methane percentage.

It might show an indication that in the commercial scale, the operational parameters have some impacts on the dependent variables. However, any changes in one parameter will not be significantly impactful since there are many parameters that influence the biogas production and the methane percentage. For example, previous study revealed that any temperature changes more than 2°C within 24 hours in the digester will negatively affect the microbial community in the digester (Meegoda et al., 2018; Schnaars, 2012). It certainly has any impact on the biogas production. Nevertheless, the other operational parameters can still backup the digester performance; therefore, it avoids the digester from a significant disturbance.

A recent study conducted by Rossi et al. (2022) can be a good example on predicting the digester performance based on the operational parameters. The study was performed by considering 55 explanatory variables to predict the specific methane production (SMP) through the MLR model. The study eventually narrowed down the parameters to include only 9 out of 55. These parameters were included as the feedstock characteristics (total solids (TS), total volatile solids (TVS), C/N ratio, and lignin content), the operating parameters (volumetric flow of inlet (Q_{in}), HRT, and OLR), and the inhibitory compounds (N-NH⁺₄ and total VFA concentration). Through this process, the study was able to gain R² as 0.87 by including TVS, OLR, C/N ratio, and lignin as the predictors. Although this study might not be representative enough as the working

volume was only around 28 L capacity (SCAD working volume is roughly 74,000 L), this can be a reference for SCAD to better predict the impact of the operational parameters on the biogas production and the methane percentage for the future study.

	Estimate Std. Error t value Pr(>						
(Intercept)	3085.543	1696.503	1.819	0.0786 .			
OLR	-31.360 150.304 -0.209 0.8361						
HRT	-23.090	19.758	-1.169	0.2515			
Temperature	3.676	41.488	0.089	0.9300			
pH	135.762	195.569	0.694	0.4927			
Residual standard error	365.4 on 31 degre	es of freedom					
Multiple R-squared	0.06047						
Adjusted R-squared	-0.06076						
F-statistic	0.4988 on 4 and 31 DF						
p-value	0.7367						
Signif. codes: 0 '***' 0.001	I '**' 0.01 '*' 0.05	·.' 0.1 · ' 1					

 Table 4.40 MLR results for operational parameter impact on the biogas production

|--|

	Estimate	Std. Error	t value	Pr(> t)			
(Intercept)	63.23486	8.17748	7.733	1.01e-08 ***			
OLR	1.05067	0.72449	1.450	0.157			
HRT	0.10655	0.09524	1.119	0.272			
Temperature	0.07043	0.19998	0.352	0.727			
рН	-0.91353	0.94268	-0.969	0.340			
Residual standard error	1.761 on 31 degree	es of freedom					
Multiple R-squared	0.09931						
Adjusted R-squared	-0.01691						
F-statistic	0.8545 on 4 and 31 DF						
p-value	0.5019						
$C_{i} = \frac{1}{2} + \frac{1}{2$	· · * * · 0 01 · * · 0 05	(,01),1					

Signif. codes: 0 **** 0.001 *** 0.01 ** 0.05 ·. 0.1 * 1

4.5.3 Digester performance from year to year

Table 4.42 provides a summary of SCAD operation in terms of feedstock input and output production. In 2014, SCAD received 14,763 metric tons of manure feedstock and 8,533 metric tons of food feedstock, with a total of feedstock received 23,297 metric tons. This feedstock produced 846,232 SCM of biogas, 1,727,073 kWh of electricity, 11,273 metric tons of effluent, and 3,920 metric tons of wet fiber. In 2015, SCAD received 13,805 metric tons of manure

feedstock and 12,800 metric tons of food feedstock, with a total of 26,605 metric tons. This feedstock produced 1,103,695 SCM of biogas, 2,118,966 kWh of electricity, 14,100 metric tons of effluent, and 4,858 metric tons of wet fiber. In 2016, SCAD received 11,059 metric tons of manure feedstock and 11,726 metric tons of food feedstock, with a total of feedstock received was 22,785 metric tons. This feedstock produced 1,418,746 SCM of biogas, 1,470,356 kWh of electricity, 14,007 metric tons of effluent, and 3,583 metric tons of wet fiber. In 2017, SCAD received 11,109 metric tons of manure feedstock and 9,129 metric tons of food feedstock, with a total of feedstock received 20,238 metric tons. This feedstock produced 1,326,335 SCM of biogas, 2,169,693 kWh of electricity, 12,751 metric tons of effluent, and 2,779 metric tons of wet fiber. In 2018, SCAD received 10,859 metric tons of manure feedstock and 8,605 metric tons of food feedstock, with a total of feedstock received was 19,464 metric tons. This feedstock produced 1,348,024 SCM of biogas, 2,680,954 kWh of electricity, 14,857 metric tons of effluent, and 1,884 metric tons of wet fiber. In 2019, SCAD received 11,353 metric tons of manure feedstock and 10,539 metric tons of food feedstock, with the total feedstock receiving 21,893 metric tons. This feedstock produced 1,280,438 SCM of biogas, 2,333,449 kWh of electricity, 15,762 metric tons of effluent, and 2,758 metric tons of wet fiber. In 2020, SCAD received 10,332 metric tons of manure feedstock and 14,531 metric tons of food feedstock, with a total of feedstock received was 24,863 metric tons. This feedstock produced 1,340,180 SCM of biogas, 2,664,665 kWh of electricity, 17,745 metric tons of effluent, and 3,082 tons of wet fiber.

The range of input acquired by SCAD from 2014 to 2020 was 10,332 to 14,763 metric tons for manure feedstock, 8,533 to 14,531 metric tons for food feedstock, and 19,464 to 26,605 metric tons for a total of both manure and food feedstock. The range of output produced by SCAD from 2014 to 2020 was 846,232 to 1,418,746 SCM of biogas, 1,470,356 to 2,680,954 kWh of electricity,

11,273 to 17,745 metric tons of effluent, and 1,884 to 4,858 metric tons of wet fiber. During seven years of operation, SCAD has processed 83,281 metric tons of manure feedstock and 75,864 metric tons of food feedstock with a total of 159,145 metric tons of organic materials to yield 8,663,649 SCM of biogas, 15,165,156 kWh of electricity, 100,495 metric ton of effluent, and 22,864 metric ton of wet fiber.

The range of input acquired by SCAD during 2014 to 2020 was 10,332 to 14,763 metric tons for manure feedstock, 8,533 to 14,531 metric tons for food feedstock, and 19,464 to 26,605 metric tons for total of both manure and food feedstock. The range of output produced by SCAD during 2014 to 2020 was 846,232 to 1,418,746 SCM of biogas, 1,470,356 to 2,680,954 kWh of electricity, 11,273 to 17,745 metric ton of effluent, and 1,884 to 4,858 metric ton of wet fiber. During 7 years of operation, SCAD has processed 83,281 metric ton of manure feedstock and 75,864 metric ton of food feedstock with total of 159,145 metric ton of organic materials to yield 8,663,649 SCM of biogas, 15,165,156 kWh of electricity, 100,495 metric ton of effluent, and 22,864 metric ton of wet fiber.

In general, the operation of SCAD went through several trends. During 2014-2015, SCAD built good trends since its establishment in 2013 due to the output production increased from 2014 to 2015. SCAD operation was disrupted during 2016 and 2017 in terms of electricity production due to the CHP engine outage that happened in those years. Moreover, the SCAD operation in 2018-2020 was more consistent and settled. In summary, the operation of SCAD has been one of the waste management solutions at MSU by processing organic waste from farming and human consumption to provide renewable energy and fertilizer. Moreover, SCAD has provided emissions reduction, such as greenhouse gases and excess nutrients. To better understand the environmental

impact resulting by SCAD, a life cycle assessment was conducted, and the result will be provided in the next chapter.

		Input		Output			
Voor	Total Manure	Total Food	Total	Total	Total	Effluent	Wet Fiber
1 cai	Pit	Pit	Feedstock	Biogas	Electricity	Total	Total
	Metric Ton	Metric Ton	Metric Ton	SCM	kWh	Metric Ton	Metric Ton
2014	14,763	8,533	23,297	846,232	1,727,073	11,273	3,920
2015	13,805	12,800	26,605	1,103,695	2,118,966	14,100	4,858
2016	11,059	11,726	22,785	1,418,746	1,470,356	14,007	3,583
2017	11,109	9,129	20,238	1,326,335	2,169,693	12,751	2,779
2018	10,859	8,605	19,464	1,348,024	2,680,954	14,857	1,884
2019	11,353	10,539	21,893	1,280,438	2,333,449	15,762	2,758
2020	10,332	14,531	24,863	1,340,179	2,664,665	17,745	3,082
Total	83,281	75,864	159,145	8,663,649	15,165,156	100,495	22,864
Average	11,897	10,838	22,735	1,237,664	2,680,954	14,356	3,266
St. Deviation	1,683	2,294	2,499	198,468	1,470,356	2,081	958

Table 4.42 Summary of SCAD performance

CHAPTER 5. LIFE CYCLE ASSESSMENT

5.1. Introduction

Life Cycle Assessment (LCA) is a methodology that is developed to evaluate the environmental burdens of processes and products during the overall life cycle, which starts from raw materials handling and processing, manufacturing, transportation and distribution, consumption stage, recycling where needed, and final disposal. ISO 14040 defines this approach as a technique to determine the specific components of a product or process that create high environmental burdens and replace it with more sustainable and environmentally practices. LCA has been used by many institutions due to its integrated way of treating the framework, impact assessment, and data quality. Moreover, it enables the users to identify the potential environmental tradeoffs between stages by systematic analysis of the diverse impacts along the entire life cycle (Azapagic et al., 2006; Khasreen et al., 2009; Odey et al., 2021).

This assessment will compare the commercial MSU South Campus Anaerobic Digester with the conventional approach of organic waste management system: landfilling the food waste and storing the manure in a long-term storage. The comparison will be based on global warming potential (GWP) and water eutrophication potential (WEP). The inputs and outputs represent the major variables that contribute to environmental impacts for both scenarios.

5.2 Goal, Scope, and Functional Unit for Life Cycle Assessment

The LCA study was conducted based on the standard methodology provided by ISO 14040 series, Environmental Management Life-Cycle Assessment. The goal of this assessment was to evaluate the environmental impact of MSU SCAD which is an anaerobic co-digestion system using diverse types of manure and organic waste to produce electricity and agricultural coproducts. The objective was to determine how much the environmental impact of the co-digestion system

(Scenario 1) and to compare these impacts with the conventional reference method (Scenario 0) performance, which stored dairy manure in a long-term storage and landfilled food waste. This LCA only focused on the treatment of raw materials (Figure 5.1). Transportation and logistics of the waste to the treatment facility were not included in the assessment. The geographical scope occurred in the lower peninsula of Michigan. The temporal scope covers waste management for four years of SCAD operations between 2017-2020. The functional unit (FU) of this assessment was 10,913 metric tons of manure wastes and 10,701 metric tons of food wastes per year.





Figure 5.1 System boundaries of Scenario 1 and 0

5.3 Life Cycle Data Inventory

The life cycle inventory (LCI) provides all important information for life cycle impact assessment. In this study, LCI was divided into raw material and handling, anaerobic digestion and energy production, lagoon storage for manure, and food waste landfill. The raw material and handling section holds information regarding the material input into the system and processes before entering the digester. Additionally, this section also includes the emissions associated with manure storage if a digester was not in place. The anaerobic digestion and energy production section provides data regarding the process of anaerobic digestion and energy production from livestock manure and food waste. The third section provides data related to manure long-term storage in lagoon storage. The last section provides data related to the food waste landfill process.

Data quality was evaluated using the Weidema method. It includes six indicators of evaluation: acquisition method, independence of data supplier, representativeness, data age, geographical correlation, and technological correlation. The score ranges from one to five, where one is the best quality and five is the most uncertain. Table 5.1 presents how to apply the indicators based on the pedigree matrix.

Table 5.2 provides information about the inventory for this LCA study. This inventory is divided into raw materials, anaerobic digestion, and energy production, animal wastes lagoon storage and land application, and food wastes landfill inventory with landfill gas (LFG) combustion.

Indicator Score	1	2	3	4	5
Acquisition method	Measured data	Calculated data based on measurements	Calculated data partly based on assumptions	Qualified estimate (by expert)	Nonqualified estimate
Independence of data supplier	Verified data, information from public or other independent source	Verified information from enterprise with interest in the study	Independent source but based on nonverified information from industry	Nonverified information from industry	Nonverified information from the enterprise interested in the study
Representative ness	Representative data form enough samples of sites over an adequate period to even out normal fluctuations	Representative data from smaller number of sites but for adequate periods	Representative data from smaller number of sites, but from shorter periods	Data from adequate number of sites but shorter periods	Representativeness unknown or incomplete data from smaller number of sites and/or from shorter periods
Data Age	Less than 3 years	Less than 5 years	Less than 10 years	Less than 20 years	Age unknown or more than 20 years
Geographical correlation	Data from area under study	Average data from larger area in which the area under study is included	Data from area with similar production conditions	Data from area with slightly similar production conditions	Data from unknown area with very different production conditions
Technological correlation	Data from enterprises, processes, and materials under study	Data from processes and materials under study but from different enterprises	Data on related processes and materials under study but from different technology	Data on related processes or materials but same technology	Data on related processes or materials but from different technology

 Table 5.1 Data Quality Evaluation Using the Weidema Method (Weidema et al., 2004)

The first part of the inventory holds information and values related to manure waste, food waste, and its chemical compositions. Manure waste is considered as low energy material, while food waste is high energy material. Therefore, co-digestion of manure with food waste potentially increases biogas production in the anaerobic digestion process (Chen et al., 2015). Data for organic waste quantity were the average of feedstock supplies in 2017-2020, which is considered the stable period of SCAD operation. Based on the data average, SCAD was fed with 10,913 metric tons of manure waste and 10,701 metric tons of food waste per year. One thing to put into consideration is that CO_2 from manure wastes and food wastes is not counted in the calculation of greenhouse gas emissions because the CO_2 is considered of biogenic origin and therefore is assumed to be offset by CO_2 capture by regrowth of the plants.

The second part of the inventory provides information and values related to anaerobic digestion and energy production from this process. Based on SCAD operational data, biogas

composition consists of 65% (v/v) of CH₄ and 35% (v/v) of CO₂. The remaining consist of a small percentage of H₂S, NH₃, and H₂. Anaerobic digestion process in SCAD produced 1,323,757 m³/year of biogas, which further converted biogas into 2,462,190 kWh-e/year of electricity and 5,584,551 kWh-e/year of heat. The remaining material was effluent as much as 19,948 metric tons/year. This part also provides information associated with the chemical contents of the effluent, including TN, TP, and soluble COD. These values will contribute to water eutrophication potential calculation. Meanwhile, biogas combustion and land application of effluent will contribute to global warming potential calculation.

The third part of Table 5.2 provides information and values related to animal waste lagoon storage and land applications. They would include CH_4 and N_2O emissions if animal wastes were only stored in lagoon storage. Furthermore, this section also includes values related to water eutrophication potential from TN, TP, and COD of animal waste land application. The fourth section provides information and values related to food waste landfills with landfill gas combustion (LFG). Like the third section, this section includes CH_4 and N_2O emissions if food wastes were only landfilled, also values related to water eutrophication potential from TN, TP, and COD of food wastes landfill. The moisture content of the typical food waste is about 70% (EPA, 2018).

Raw materialsManure wastes10,913Metric ton/yearOperational data1, 1, 2, 2, 1, 1Total solids of manure wastes11.4%Operational data1, 1, 2, 2, 1, 1Volatile solids of manure wastes10.0%Operational data1, 1, 2, 2, 1, 1TN of manure wastes4,143mg/kgOperational data1, 1, 2, 2, 1, 1TP of manure wastes413mg/kgOperational data1, 1, 2, 2, 1, 1SCOD of manure wastes59,446mg/kgOperational data1, 1, 2, 2, 1, 1Food wastes10.01%Operational data1, 1, 2, 2, 1, 1Food wastes10.11%Operational data1, 1, 2, 2, 1, 1Total solids of food wastes10.1%Operational data1, 1, 2, 2, 1, 1Total solids of food wastes10.1%Operational data1, 1, 2, 2, 1, 1To food wastes9.3%Operational data1, 1, 2, 2, 1, 1TN of food wastes5,318mg/kgOperational data1, 1, 2, 2, 1, 1SCOD of food wastes17,525mg/kgOperational data1, 1, 2, 2, 1, 1CO2 content in biogas65% v/vOperational data1, 1, 2, 2, 1, 1CO2 content in biogas34% v/vOperational data1, 1, 2, 2, 1, 1CO2 content in biogas5,584,551kWh-e/yearOperational data1, 1, 2, 2, 1, 1To of effluent19,948Metric ton/yearOperational data1, 1, 2,		Value	Unit	Source	DQI
Manure wastes10,913Metric ton/yearOperational data1, 1, 2, 2, 1, 1Total solids of manure wastes11.4%Operational data1, 1, 2, 2, 1, 1Volatile solids of manure wastes10.0%Operational data1, 1, 2, 2, 1, 1TN of manure wastes4,143mg/kgOperational data1, 1, 2, 2, 1, 1TP of manure wastes413mg/kgOperational data1, 1, 2, 2, 1, 1SCOD of manure wastes59,446mg/kgOperational data1, 1, 2, 2, 1, 1Food wastes10.701Metric ton/yearOperational data1, 1, 2, 2, 1, 1Total solids of food wastes10.1%Operational data1, 1, 2, 2, 1, 1Volatile solids of food wastes9.3%Operational data1, 1, 2, 2, 1, 1TN of food wastes5,318mg/kgOperational data1, 1, 2, 2, 1, 1TN of food wastes17,525mg/kgOperational data1, 1, 2, 2, 1, 1TP of food wastes17,525mg/kgOperational data1, 1, 2, 2, 1, 1CO2 content in biogas65% v/vOperational data1, 1, 2, 2, 1, 1CO2 content in biogas34% v/vOperational data1, 1, 2, 2, 1, 1CO2 content in biogas34% v/vOperational data1, 1, 2, 2, 1, 1CO2 content in biogas34% v/vOperational data1, 1, 2, 2, 1, 1Heat production from biogas5,584,551kWh-e/yearOperational data1, 1, 2,			Raw materials		
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	TP of effluent	584	mg/kg	Operational data	1, 1, 2, 2, 1, 1
sCOD of effluent 7.894 mg/kg Onerational data 1.1.2.2.1.1	sCOD of effluent	7,894	mg/kg	Operational data	1, 1, 2, 2, 1, 1

Table 5.2 Inventory for the life cycle assessment ^a

Table 5.2 (cont'd)

N ₂ O emission from the effluent	0.005	g N ₂ O/g TN in the effluent	(RTI International, 2010)	2, 1, 1, 4, 1, 2
GWP of N2O	298	g CO ₂ -e/g N ₂ O	(RTI International, 2010)	2, 1, 1, 4, 1, 2
CH4 emission from effluent	3.08×10 ⁻⁴	Metric ton CO ₂ - e/metric ton TS in the effluent	(Turnbull & Komthunzi, 2004)	2, 1, 1, 4, 2, 2
Water eutrophication potential (WEP) of TN	0.9864	g N-e/kg TN in the effluent	(RTI International, 2010)	2, 1, 1, 4, 1, 2
Water eutrophication potential (WEP) of TP	7.29	g N-e/kg TP in the effluent	(RTI International, 2010)	2, 1, 1, 4, 1, 2
Water eutrophication potential (WEP) of COD	0.05	G N-e/kg COD in the effluent	(RTI International, 2010)	2, 1, 1, 4, 1, 2
An	imal wastes lago	oon storage and land	d application invent	ory
CH ₄ emission	0.127	Metric ton CH4/metric ton VS	(Owen & Silver, 2015)	2, 1, 1, 3, 2, 2
N ₂ O emission	0.005	g N ₂ O/g TN in the waste	(RTI International, 2010)	2, 1, 1, 4, 1, 2
Water eutrophication potential (WEP) of TN	0.9864	g N-e/kg TN in the waste	(RTI International, 2010)	2, 1, 1, 4, 1, 2
Water eutrophication potential (WEP) of TP	7.29	g N-e/kg TP in the waste	(RTI International, 2010)	2, 1, 1, 4, 1, 2
Water eutrophication potential (WEP) of COD	0.05	g N-e/kg COD in the waste	(RTI International, 2010)	2, 1, 1, 4, 1, 2

Table 5.2 (cont'd)

Food wastes landfill inventory with landfill gas (LFG) combustion				
CH ₄ emission,		Metric ton CO ₂ -	(Environmental	
food wastes	2.3	e/ton TS food	Protection	2, 1, 1, 1, 2, 2
landfill		waste	Agency, 2020) ^b	
N ₂ O emission	0.005	g N ₂ O/g TN in the waste	(RTI International, 2010)	2, 1, 1, 4, 1, 2
Water eutrophication potential (WEP) of TN	0.9864	g N-e/kg TN in the waste	(RTI International, 2010)	2, 1, 1, 4, 1, 2
Water eutrophication potential (WEP) of TP	7.29	g N-e/kg TP in the waste	(RTI International, 2010)	2, 1, 1, 4, 1, 2
Water eutrophication potential (WEP) of COD	0.05	G N-e/kg COD in the waste	(RTI International, 2010)	2, 1, 1, 4, 1, 2

a. CO_2 from manure wastes and food wastes is not counted in the calculation of greenhouse gas emissions because the CO_2 is considered of biogenic origin and therefore is assumed to be offset by CO_2 capture by regrowth of the plants.

b. The moisture content of the typical food wastes in the reference is set at 70%.

5.4 Data Quality Evaluation

This LCA study was supplied by legitimate sources, such as daily operational data from SCAD, research publications, and government annual reports. Table 5.3 provides the data quality evaluation for the life cycle inventory. Acquisition method, independence of data supplier, geographical correlation, and technological correlation were scored 1. Additionally, representativeness and data age were scored 2 and 3, respectively.

Majority of data were from SCAD operational data which are primary data for the study. Furthermore, the study acquired data from verified institutions including EPA and RTI International. For data age, there was one source from RTI International which is less than 20 years old. This data was used because the recent study related to this topic still refers to the data set in this publication. All data included in the study is geographically in the US.

Indicator	DQI Score	Discussion
Acquisition Method	1	Measured data
Independence of Data Supplier	1	Verified data, information from public or other independent source
Representativeness	2	Representative data from smaller number of sites but for adequate periods
Data Age	3	Less than 10 years
Geographical Correlation	1	Data from area under study
Technological Correlation	1	Data from enterprises, processes, and materials under study

Table 5.3 Data Quality Evaluation Summary for LCI

5.5 Mass and Energy Balance of the process at different months and years

Mass and energy balance analysis was carried out based on the operational data from 2017 to 2020. The energy inputs for the digestion operation include heat (W_{heat}, kWh-e/year) to maintain the digestion temperature as well as electricity (W_{electricity}, kWh-e/year) to power pumps, mixers, and other accessary equipment. The energy inputs were calculated using the following equations modified from a previous study (Bustamante and Liao, 2017):

$$W_{heat} = m \times C_p \times (T_R - T_0) \times (1 + 20\%) \times 0.0002778$$
(3)
$$W_{electricity} = m \times 0.00788$$
(4)

Where m is the amount of the wet weight of the feedstock per year (kg); C_p is the heat capacity of the wet FM (4.12 kJ/(kg K)); T_R is the reactor temperature (313 K); To is the temperature of the wet feedstock based on the average environmental temperature in East Lansing, MI (288 K); 20% is the percentage of the additional heat needed to maintain the digestion temperature; 0.0002778 is the conversion factor of KJ to kWh; and 0.00788 is the average electricity demand unit of the digester operation (kWh/kg wet FM).

Figure 5.2 provides mass balance scheme that happened in SCAD. In general, the input comes from the feedstock then be processed in the digester to produce biogas and effluent. Furthermore, the effluent is divided into wet solid digestate and filtrate. These numbers represent the average values of each parameter in 2017-2020.

From feedstock input, the amount of feedstock processed in the digester was 21,614 metric tons per year, which combined both manure pit and food pit. The amount of feedstock from manure pit and food pit were 10,913 metric ton and 10,701 metric ton per year, respectively. Furthermore, these feedstocks contain 10.6% (w/w) of total solids and 10.0% (w/w) of volatile solids. In terms of the nutrient contents, the amount of total nitrogen, total phosphorus, and soluble COD in the feedstock were 4,690 mg/kg, 633 mg/kg, and 38,700 mg/kg, respectively. These numbers were gained from a single measurement that happened in 2017, as the only measurement for feedstock during this period.

Anaerobic digestion process transformed the raw materials to produce biogas and effluent. Biogas production was 1,323,757 m³/year with CH₄, CO₂, and H₂S contents were 65% (v/v), 34% (v/v), and 495 ppm, respectively. Effluent yielded from anaerobic digestion process was 19,948 metric ton/year. The effluent contained 6.3% of total solids (w/w) and 5.1% of volatile solids (w/w). In terms of the nutrient contents, the amount of total nitrogen, total phosphorus, and soluble COD in the effluent were 3,246 mg/kg, 584 mg/kg, and 7,894 mg/kg, respectively.

Through the separation process, effluent became wet solid digestate and filtrate for further application. The amount of filtrate produced was 17,081 metric ton/year with total solids and volatile solids contents were 4.5% (w/w) and 3.0% (w/w), respectively. Filtrate also contained 3,293 mg/kg of total nitrogen and 509 mg/kg of total phosphorus. For wet solid digestate, the yield was 2,867 metric ton/year with total solids and volatile solids contents were 27.6% (w/w) and

24.4% (w/w), respectively. Digestate also contained 529 mg/kg of total nitrogen and 171 mg/kg total phosphorus.



Figure 5.2 Mass balance of the anaerobic digestion process *: Data are from a single measurement in 2017

Table 5.4 provides information about energy balance of anaerobic digestion process that happened in SCAD. The energy balance includes energy input, energy output, and net energy output coming from heat and electricity. SCAD consumed 742,090 kWh-e/year of heat to maintain the digester operational temperature and 170,320 kWh-e/year of electricity to power pumps, mixers, and other equipment on the site. Anaerobic digestion process yielded heat and electricity as much as 5,584,551 kWh-e/year and 2,462,190 kWh-e/year, respectively. Net energy output was gained from subtracting energy produced with energy consumed. The net energy output yielded by SCAD in the form of heat and electricity were 4,842,461 kWh-e/year and 2,291,870 kWh-e/year, respectively. Furthermore, from energy output values, the efficiency of electricity generation was calculated which was 29%.

Table 5.4 Energy balance of the anaerobic digestion process ^a

	Anaerobic digester
Energy input	
Heat input (W _{heat} , kWh-e/year) ^b	-742,090
Electricity input (Welectricity, kWh-e/year) ^c	-170,320
Energy output	
Energy output as heat (E _{heat} , kWh-e/year) ^d	5,584,551
Energy output as electricity (E _{electricity} , kWh-e/year) ^e	2,462,190
Net energy output	
Net heat output (kWh-e/year) ^f	4,842,461
Net electricity output (kWh-e/year) ^g	2,291,870

a. Negative numbers mean energy inputs, and positive numbers mean energy outputs.

b. Eq. 1 was used to calculate the heat input.

- c. Eq. 2 was used to calculate the electricity input.
- d. The annual biogas production of 1,323,757 m³ with 65% (v/v) of methane was used to calculate the energy content of the biogas. The low heating value of methane is 35.8 MJ/m³ methane. The thermal conversion efficiency of the CHP unit is 65%.
- e. The electricity output is the metered number of the digestion operation.
- f. The net heat output = E_{heat} W_{heat}
- g. The net electricity output = $E_{electricity}$ $W_{electricity}$

5.6 Impact Assessment

Two impact categories were chosen for Life Cycle Impact Assessment: Global Warming Potential (GWP) and Water Eutrophication Potential (WEP). The classification of each category is defined by the US Environmental Protection Agency (EPA). The LCIA provides an analysis of environmental impacts on both scenarios as a comparison to the emissions of waste management methods for animal wastes and food wastes. This could be done by calculating the environmental impacts yielded from processes or products associated with the proposed systems.

5.6.1 Global Warming Potential (GWP)

Global warming is defined as the raising of Earth's temperature due to GHG emissions globally, which mainly coming from human activities. The main GHGs are CO₂, CH₄, and N₂O.

Currently, the main energy supplies are still from fossil fuels such as oil and coal, which contribute roughly 65% of GHG emissions (EPA, 2019). Therefore, renewable energy is expected to address this concern. Anaerobic digestion has been promoted as a renewable energy system that potentially lower the global warming potential by reducing emissions from manure and food waste.

Global Warming Potential (GWP) is the amount of GHG released during the life cycle of a process. Carbon dioxide is commonly used as a reference gas to compare the impact of various greenhouse (Shine, 2009). For this LCA study, data were collected for CH₄, and N₂O emissions. They were normalized to 1 ton of CO₂ equivalent (CO₂-e) based on the following conversions: 1 kg CH₄ = 25 kg CO₂-e; and 1 kg N₂O = 298 kg CO₂-e (RTI International, 2010). CO₂ from manure wastes and food wastes is not counted in the calculation of greenhouse gas emissions because the CO₂ is considered of biogenic origin and therefore is assumed to be offset by CO₂ capture by regrowth of the plants.

Figure 5.3 provides information related to GWP contribution analysis for each scenario. According to the impact assessment, the AD system has an overall GWP of 1,842-ton CO₂-e/year, while the land applications system produces 6,190-ton CO₂-e/year. This result shows that the AD system can reduce GWP up to 70% lower than the landfills system.

Furthermore, biogas and AD effluent were the only parameters to be considered for calculating GWP. Electricity, heat utilization, digestate, and filtrate were not calculated as they are outside of the system boundaries. For the AD scenario, biogas combustion contributed 92% of the total GWP, while land application of effluent only contributed 8%. For the landfill scenario, biogas production was counted as CH₄ and N₂O emissions from manure land application with lagoon storage and food wastes land application with landfill gas combustion. Lagoon storage contributed

58% of the total GWP, followed by landfill gas combustion with 42% of contribution towards the total GWP.

The result of this study was supported by previous result from Chen et al. (2015), where a co-digestion of dairy manure and bakery wastes in the AD system was compared to the AD system with dairy manure only and landfilling the bakery waste. The study revealed that co-digestion of 7,147-ton dairy manure and 2,382-ton food waste had an overall GWP of 1.6×10^4 ton CO₂-e, while another scenario yielded an overall GWP of 2.7×10^4 ton CO₂-e, which means that AD scenario resulted in a 42% reduction in GWP. This could be evidence that the anaerobic co-digestion system has a capability to mitigate the global climate change rather than the conservative methods such as landfilling the food waste or digestion of manure only.



Figure 5.3 GWP contribution analysis of the anaerobic digestion process and land applications

5.6.2 Water Eutrophication Potential (WEP)

Eutrophication is a situation where a water body contains excessive nutrients that affect to a dense growth of plant life and death of water animals due to lack of oxygen. It is due to nutrients

runoff from the land, such as nitrogen and phosphorus, which then accumulate in the water. Consequently, it creates a "dead zone," which is an area with low oxygen content that suffocates marine life (Mueller & Helsel, 1996). Water eutrophication potential (WEP) is the impact resulting from excessive nutrient supplies on terrestrial and aquatic environments, particularly the most important substances such as nitrogen (N) and phosphorus (P). WEP can be presented as either nitrogen equivalents (N-eq.) or phosphate equivalents (PO4-eq.) (Guinee, 2002).

A kilogram of nitrogen equivalents (kg N-eq.) units was used to assess WEP in this study. TN, TP, and COD were the three parameters used for WEP assessment. All values related to TN, TP, and COD contents were measured as SCAD operational data, except for COD concentration of manure which was back calculated. WEP conversion values for TN, TP, and COD are 0.9864 g N-e/kg TN in the waste, 7.29 g N-e/kg TP in the waste, and 0.05 g N-e/kg COD in the waste, respectively (RTI International, 2010).

Figure 5.4 provides information related to WEP contribution analysis for each scenario. According to the impact assessment, the AD system has an overall WEP of 173 kg N-e/year, while the land applications system produces 232 kg N-e/year. This result shows that the AD system can reduce WEP up to 25% lower than the landfills system.

WEP from the AD system came from three parameters, which were TN, TP, and COD of AD effluent. TP became the most significant WEP contributor in the AD system, which was 54% of the total WEP, followed by TN, which contributed 41% of the total WEP. Meanwhile, COD only contributed 9% of the total WEP. In the landfills system, the chemical content was divided into manure waste and food waste. Both wastes have TN as the most significant contributor in WEP. Food waste TN became the highest contributor of WEP in landfills system, which was 27%, followed by Manure TN at 21%. Food waste TP, manure TP, and manure COD were closed

together, which were 17%, 16%, and 15% of WEP contribution, respectively. Food waste COD was the least contributor of WEP with 4% of contribution.

AD system does not contribute much to reducing the nutrient content in the organic materials. WEP in AD is more correlated to feedstock quality and the digester's HRT. It barely had any correlation to biogas productivity. Therefore, nutrient content in the feedstock did not significantly alter. Nevertheless, AD transformed the elemental nitrogen and phosphorus to become ammonia and phosphate, respectively. Compared to elemental nitrogen and phosphorus, ammonia and phosphate are more beneficial for soil conversions and plant adsorption (Arosemena, 2021; R. Chen et al., 2015; Field et al., 1984).



Figure 5.4 WEP contribution analysis of the anaerobic digestion process and land applications

5.7 Interpretation

5.7.1 Sensitivity Analysis

Sensitivity analysis is a method to measure any changes in a certain impact by simulating any changes in key parameters that influence the model. This method can report which parameters greatly affect any changes in each impact category. By doing a sensitivity analysis with LCA, a study can depict further actions to conduct to lower the environmental burdens of a product or a system.

5.7.1.1 Global Warming Potential (GWP)

The sensitivity analysis for the anaerobic digestion and the land applications systems were performed by modifying $\pm 25\%$ of each variable of interest while keeping the other variables constant for the base case. The parameters analyzed for the anaerobic digestion were biogas combustion and AD effluent, while the parameters analyzed for the land applications were NO₂ and CH₄ emissions from manure and food waste. The sensitivity analysis result for the anaerobic digestion system, as presented in Figure 5.5, shows that biogas combustion becomes the most sensitive parameter towards the impact category value, while effluent becomes the least sensitive parameter. The change in biogas combustion ($\pm 25\%$) resulted the highest change in GWP values at 29.44%. It is due to biogas combustion is the main process in converting methane into electricity. Meanwhile, the change in AD effluent only resulted 2.31% change in GWP values.

The sensitivity analysis result for the land application system, as presented in Figure 5.6, shows that CH₄ emission from manure becomes the most sensitive parameter towards the GWP value, followed by CH₄ emission from food waste. The change in CH₄ emissions from manure resulted 14% change in GWP value, while the change in CH₄ emission from food waste resulted 10% change in GWP value. The emissions of NO₂ from food waste and manure become the least sensitive parameters, which only affect 0.54% and 0.43% of change in GWP value, respectively.



Figure 5.5 Anaerobic digestion sensitivity analysis for GWP



Figure 5.6 Land applications sensitivity analysis for GWP

5.7.1.2 Water Eutrophication Potential (WEP)

The sensitivity analysis for the anaerobic digestion and the land applications systems were performed by modifying $\pm 25\%$ of each variable of interest while keeping the other variables constant for the base case. The parameters analyzed for the anaerobic digestion were TN, TP, and COD of the effluent. Meanwhile, the parameters analyzed for the land applications were TN, TP, and COD of both manure and food waste. The sensitivity analysis result for the anaerobic digestion system, as presented in Figure 5.7, shows that effluent TP becomes the most sensitive parameter towards the impact category value, while effluent COD becomes the least sensitive parameter. The change in effluent TP ($\pm 25\%$) resulted the highest change in GWP values, at 13.5%, followed by effluent TN at 10.2%. Meanwhile, the change in effluent COD only resulted 1.25% change in GWP values.

The sensitivity analysis result for the land application system, as presented in Figure 5.8, does not show significant difference between most of parameters. Food waste TN becomes the most sensitive parameter, followed by manure TN and food waste TP. These three parameters resulted 6.7%, 5.3%, and 4.1% change in WEP value, respectively. On the other hand, food waste COD becomes the least sensitive parameter with only 1.1% change in the WEP value. Manure TP and COD contribute similar impact to WEP value, which are 3.9%.



Figure 5.7 Anaerobic digestion sensitivity analysis for WEP



Figure 5.8 Land applications sensitivity analysis for WEP

5.7.2 Consistency and Completeness Check

The consistency check aims to demonstrate that assumptions, methods, and data used throughout the LCA process are aligned with the goal and scope of the study. It can show where the data consistency can be improved to compare systems to one another. The consistency check and explanations of inconsistency are explained within Table 5.5. The overall data adequately shows consistency to support the goal and scope of the study.

Category	Checklist and Inconsistencies
Data Source	Both scenarios have most data based on operational data and legitimate sources
Data Accuracy	Both scenarios are supplied with accurate data from measurement or calculation based on previous studies
Technological Representation	Both scenarios have data available for conducting the study
Temporal Representation	Both technologies are utilized up to date
Geographical Representation	Both technologies include data from the United States
System Boundary, Assumption and Model	Both systems serve as a waste management system

 Table 5.5 Checklist and Inconsistencies based on Data Quality

A completeness check ensures that the study has complete available sources for data interpretation. In case there are some gaps in the completeness of the data, it should be verified whether the incompleteness will affect the goal and scope of the study. Table 5.6 provides a completeness check for the AD system, while Table 5.7 provides a completeness check for the landfill system. In general, all data required for the study was completed. SCAD operational data contains most parameters needed for the study, even though there were several parameters that rely on a single measurement and one parameter needed a back-calculation. Manure storage and food waste landfill are two common methods for waste management in the US, therefore providing

legitimate sources for the impact assessment. There might be potential data gaps if the system boundary is extended since this study only focused on the waste treatment.

Life cycle stage	AD	Complete	Required Actions
Raw materials	Х	Yes	-
AD and energy production	Х	Yes	-
Output: Effluent	Х	Yes	-
X: data availa	ble	n.a.: not app	licable

Table 5.6 Completeness check for AD system

Table 5.7 Completeness check for landfills system

Life cycle stage	Landfill	Complete	Required Actions
Raw materials	Х	Yes	-
Manure storage	Х	Yes	-
Food waste landfill	Х	Yes	-
X: data availa	ble	n.a.: not app	licable

5.8 Technoeconomic Analysis

Technoeconomic analysis is conducted to evaluate whether renewable energy production can also attract investors based on its financial benefits of it. One of the significant issues in producing renewable energy on a large scale is the large investment for capital and/or operational costs, which makes renewable energy seems environmentally favorable but less economically competitive (Carneiro and Ferreira, 2012; Fersi et al., 2012). This section will analyze the cost and profit analysis of SCAD as a commercial digester, including capital expenditures (CapEX), operational expenditures (OpEX), and revenues. The payback period was calculated according to capex and total net profit. Table 5.8 provides information for the whole performance of technoeconomic analysis. The first section holds information about CapEX. Digester construction cost became the highest expense in this section, contributing 40.21% to the total CapEX. Feedstock receiving and combined heating and power (CHP) costs were the other major expenses with relatively similar values, contributing 20.29% and 21.71% to total CapEX, respectively. The total system CapEX was \$3,586,861.

The second section of Table 5.8 provides information about the OpEX, which is calculated per year. In this study, OpEX was taken from operational data in 2019-2020 as the recent years of the study. Among 12 items described in OpEX, labor cost became the highest expense which contributed roughly 41.5% to total OpEX. CHP engine service became the second-highest expense, contributing about 25% to total OpEX. DHT transport means that every gallon of food waste brought to the digester has to be exported to maintain nutrient balance on campus. It became the third-largest expense, contributing about 17.75% to total OpEX. The total system OpEX was \$298,156.

The third section belongs to revenue. Tipping fees and electricity were the two revenue sources of SCAD. Both commodities had similar values during 2019-2020, which were \$237,746 for electricity and \$217,854 for tipping fees. The total system revenue was \$455,600.

SCAD has experienced volatile revenue from electricity due to some changes in electricity pricing. The model was proposed to the board of trustees at the price of \$0.123/kWh. Prior to 2018, the price was between \$0.04 and \$0.06/kWh due to the standby charge implemented by Consumers Energy. Standby charge is a measurement of energy production in a 15-minute interval and during peak time from 09:00 AM to 06:00 PM. If the digester does not produce power during that 15 minute, the digester gets the lower rate, and it is for the entire month, not only for that 15 minute.

In 2018, SCAD was directly connected to the campus power supply, which resulted in a flat rate of \$0.1017/kWh.

Total net revenue was calculated by subtracting total revenue from total OPEX. The net revenue value of SCAD was \$157,444. The payback period was then calculated by considering 5-year average local inflation of 3.2% in the U.S. as the inflation rate and 20 years of depreciation period on CapEx. The annual depreciation rates from Modified Accelerated Cost Recovery System (MARCRS) are: 0.100, 0.188, 0.144, 0.115, 0.092, 0.074, 0.066, 0.066, 0.065, 0.065, and 0.033 (after 10 years). Based on cash flow calculation, the payback period will be in 21.5 years; then, the digester will start to gain profit afterward. This can be considered a quite promising payback time and economically competitive. An economic sensitivity analysis would be strongly recommended to be conducted to know which parameter is the most sensitive; thus, it can shorten the length of the payback period.

≜	<u> </u>	
Capital expenditure (CapEX)	Cost	Reference
Feedstock Receiving	\$727,927	Data
Digester	\$1,442,140	Data
СНР	\$778,651	Data
Bond	\$38,143	Data
Interconnection	\$300,000	Data
Site improvements & excavation	\$300,000	Data
Total CapEX	\$3,586,861	
OpEX (per year)		
AD Repairs	\$28,373	Data
ADMIN Fee	\$2,948	Data
Bio Analysis	\$2,827	Data
CHPS	\$74,226	Data
Labor	\$123,616	Data
Laundry	\$378	Data
Maintenance and Repair	\$6,482	Data
MISC	\$4,064	Data
Motor Pool / Vehicle	\$1,165	Data
Supplies	\$396	Data
Telephone	\$772	Data

Table 5.8. Economic performance of the digestion process ^a
Table 5.8 (cont'd)

Transport (DHT)	\$52,910	Data
Total OpEX (per year)	\$298,156	
Revenue (per year)		
Electricity	\$237,746	Data
Tipping	\$217,854	Data
Total revenue (per year)	\$455,600	
Total net revenue (per year) ^b	\$157,444	
Payback time (Years) ^c	21.5	

a. The OpEX and revenue are the operational data from 2019-2020.

b. The net revenue = Total revenue – Total OpEx

c. The 5-year average local inflation of 3.2% in the U.S. is used as the inflation rate. The depreciation period is set at 20 years. The depreciation is just on CapEx. The annual depreciation rates from MARCRS (Modified Accelerated Cost Recovery System) are: 0.100, 0.188, 0.144, 0.115, 0.092, 0.074, 0.066, 0.066, 0.065, 0.065, and 0.033 (after 10 years).



Figure 5.9 Total net cash flow and payback period of SCAD

There are three basic scenarios that can be considered to shorten the payback period. The first scenario would be fixing the electricity price; the second scenario would include digestate as part of revenue, while the third scenario would be the combination of the first and the second scenario.

As mentioned earlier, the proposed electricity price for SCAD to the board of trustees was \$0.123/kWh, while the current price for electricity is \$0.1017/kWh. That means there is a price gap of as much as \$0.0213/kWh. If the proposed price can be achieved, the revenue from electricity

would be \$307,713. Because of it, the total net revenue will be increased to \$227,412, and it will lower the payback period to be in 15 years.

The second scenario would be including digestate as part of the revenue. As per the current study, digestate is still part of OpEX, which is DHT transport. Digestate is a source of nutrients for the soil. If the digester can manage the digestate to be part of the revenue, then the DHT transport cost will be excluded from OpEX. Furthermore, the digestate price is assumed to be \$7.00 per metric ton. The average digestate production in 2019-2020 was about 18,471 metric tons. Therefore, the revenue from selling the digestate can be up to \$129,297. This additional revenue will increase the total revenue to \$584,897 and the total net revenue to \$339,652. It will also reduce the total OpEX to \$245,246. These changes can reduce the payback period to be in 10.5 years.

The third scenario would be the best scenario that SCAD can afford. With electricity price fixed and digested becoming part of the revenue, the total revenue will become \$654,865, total net revenue will be \$409,619, and the payback period will be in 9 years. Regardless of which scenario seems to be the most feasible, it shows that SCAD has a competitive economic advantage as a commercial-scale anaerobic digester.

5.8.1 Sensitivity Analysis

Sensitivity analysis was done to check which parameters significantly affect the payback period. Nine variables – feedstock receiving, digester, CHP, interconnections, site improvements, CHP maintenance, labor, electricity, and tipping fees – were taken into consideration for the economic sensitivity analysis of this study. The analysis was done by modifying each variable of interest by $\pm 25\%$ while keeping the other variables constant for the baseline scenario. The number

changed represents how much each variable affects the increasing or decreasing of the payback period compared to the baseline scenario.

The result (Figure 5.6) shows that electricity was the most sensitive among all variables, followed by tipping fees and labor costs as the second and third most sensitive variables. Increasing the revenue from electricity by 25% helps to decrease the payback period by 11.41 years, while the decrease in this parameter by 25% contributes to increasing the payback period by 26.06 years. The increasing of tipping fees by 25% reduces the payback period by 10.69 years while decreasing the tipping fees by 25% increases the payback period by 22.67 years. Meanwhile, the increase of labor cost by 25% increases the payback period by 17.61 years, while reducing the cost by 25% lowers the payback period by 9.44 years. Interconnections, site improvements, and feedstock receiving become the least sensitive among all parameters.



Figure. 5.10 Sensitivity analysis chart for SCAD payback period

CHAPTER 6. OVERALL CONCLUSION AND RECOMMENDATIONS

6.1 SCAD operational performance

The goal of this study is to evaluate the operational performance of Michigan State University's South Campus Anaerobic Digester (SCAD) as a commercial digester, as well as the environmental impact during its operation from 2014-to 2020. This study concluded that SCAD received a total of 18 different feedstocks throughout its operation thus far. Among all feedstocks, Dairy Gutter, Parlor, and FOG become the major feedstocks for the digester. Multilinear regression was conducted to determine feedstocks that have the most significant impact on biogas production. The filtrate in the manure pit, Dairy Gutter, Parlor, and FOG are feedstocks that have the most significant impact on biogas production, with a p-value<0.05. The Multiple R-squared was 58.24%, and the Adjusted R-squared was 56.01%, showing that the model was able to explain 56% to 58% of the data variance to predict the biogas production. It is due to SCAD receiving a wide variety of feedstock with an uncertain frequency which affects the R-squared values. However, the study can still draw a conclusion since there are independent variables that are statistically significant to correlate the relationships between the variables.

In general, the operation of SCAD went through several trends. From 2014 to 2015, SCAD built good trends since its establishment in 2013 due to the output production increased from 2014 to 2015. SCAD operation was disrupted from 2016 to 2017 in terms of electricity production due to the CHP engine outage that happened in those years. Moreover, the SCAD operation from 2018 to 2020 was more consistent and settled.

In summary, the operation of SCAD has been one of the waste management solutions at MSU by processing organic waste from farming and human consumption to provide renewable energy and fertilizer. During its operation years from 2014 to 2020, SCAD has processed 159,145

metric tons of feedstock that consist of 83,281 metric tons of manure wastes and 75,864 metric tons of food waste to produce 8,663,649 SCM of biogas, 15,165,156 kWh of electricity, 100,495 metric tons of effluent, and 22,864 metric tons of wet fiber.

6.2 Life Cycle Assessment and Technoeconomic Analysis

A life cycle assessment was conducted to compare the environmental impact of the AD system with a conventional system that combines lagoon storage for manure wastes and landfills for food wastes. This LCA study was supplied by legitimate sources, such as daily operational data from SCAD, research publications, and annual government reports. Acquisition method, independence of data supplier, geographical correlation, and technological correlation were scored 1. Additionally, representativeness and data age were scored 2 and 3, respectively.

The result showed that the AD system possesses fewer environmental burdens in both GWP and WEP compared to the conventional system. AD system has an overall GWP of 1,842-ton CO₂-e/year, while the land applications system produces 6,190-ton CO₂-e/year. This result shows that the AD system can reduce GWP up to 70% lower than the landfills system. AD system has an overall WEP of 157 kg N-e/year, while the land applications system produces 210 kg N-e/year. This result shows that the AD system has a WEP 26% lower than the landfills system.

Sensitivity analysis was conducted to determine parameters that give the most significant impact on the LCA study. From GWP, biogas combustion became the most sensitive parameter on the AD system. Meanwhile, CH₄ emissions from manure became the most sensitive parameter on the land application system. From WEP, effluent TP became the most sensitive parameter on the AD system. Meanwhile, food waste TN became the most sensitive parameter on the land application system. Technoeconomic analysis was conducted to understand the financial feasibility of SCAD as a commercial digester. The result showed that SCAD needs 21.5 years to accomplish its payback time, which is considered quite economically competitive. Three basic scenarios can be done to gain a better payback time for up to 9 years. Economic sensitivity analysis shows that the revenue from electricity is the most sensitive parameter to affect the payback period, followed by tipping fees and labor costs. Increasing the revenue from electricity can lower the payback period by 11.41 years. Meanwhile, interconnections, site improvements, and feedstock receiving are the least sensitive parameters.

6.3 Further recommendations

Future work for LCA study can include more impact parameters for the study, such as air acidification potential and smog potential. Since the current study has a system boundary that only includes organic waste treatment at the site, it would be interesting to know the dynamics in the LCA study once the system boundary is extended, such as including digestate land application or transportation to the treatment site.

As for technoeconomic analysis, digestate is not included in the revenue as per the current study. Including digestate as an organic fertilizer can be additional revenue that can shorten the payback time of the digester. Moreover, electricity prices are critical to maintaining SCAD revenue. Getting an ideal price as proposed will be another improvement for SCAD, although it depends on the policy implementation for renewable electricity pricing.

Data availability can be another aspect to improve, especially laboratory chemical analysis. The current study revealed that SCAD has fewer laboratory results for influent chemical analysis and several chemical analyses for effluent due to workforce constraints. If this challenge can be addressed and future work can provide more frequent chemical analysis for the digester influent and effluent, it will help to provide more accurate results for mass balance analysis and life cycle impact assessment. APPENDIX

A. MLR Codes

Statistical analysis - Multiple regression analysis
SCAD operation
Wei Liao, October 23, 2021
Fahmi Nov 23, 2021
Fahmi Dec 9, 2021
Fahmi Dec 14, 2021

Load libraries ----library(dplyr)
library(FSA)
library(psych)
library(car)
library(rcompanion)

Choose data file of "BiogasProduction_v3.txt" ----con <-file.choose(new = FALSE)
metadata <- read.table(con, header = T, row.names = 1, fill = TRUE)
head(metadata)</pre>

data.num

corr.test(data.num, use = "pairwise", method="pearson", adjust="none", alpha="0.05")

Stepwise procedure

model.null = lm(Biogas_production~1, data=data.num)
model.full = lm(Biogas_production ~ Digestate + Filtrate_manure_pit + SLS_Solids+Dairy_G +
Parlor + Beef + W_Feed_manure + Poultry + Swine + ANS_Other+
SLS_Solids_food + P_A + Pulp + FOG + W_Feed +
Other + Cart_Food, data=data.num)

step(model.null, scope = list(upper=model.full), direction="both", data=data.num)

Define the final model from the results of the stepwise procedure model.final = lm(Biogas_production ~ Filtrate_manure_pit + Dairy_G + Parlor + ANS_Other + P_A, data=data.num) summary(model.final)

Analysis of variance for individual feed components Anova(model.final, Type="II")

Simple plot of predicted values with 1-to-1 line data.num\$predy = predict(model.final)

abline(0,1, col="blue", lwd=2)

Checking assumptions of the model hist(residuals(model.final), col="darkgray")

plot(fitted(model.final),residuals(model.final))

RESULTS

Call: lm(formula = Biogas_production ~ Filtrate_manure_pit + Dairy_G + Parlor + ANS_Other + P_A, data = data.num)

Coefficients:

(Intercept) Fil	ltrate_manure_pit	Dairy_G
2330.91251	-0.01688	0.18163
Parlor	ANS_Other	P_A
-0.04628	0.08850	-0.02158

> # Define the final model from the results of the stepwise procedure

> model.final = lm(Biogas_production ~ Filtrate_manure_pit + Dairy_G + Parlor + ANS_Other + P_A,

+ data=data.num) > summary(model.final)

Call:

lm(formula = Biogas_production ~ Filtrate_manure_pit + Dairy_G + Parlor + ANS_Other + P_A, data = data.num)

Residuals:

Min 1Q Median 3Q Max -2064.80 -335.13 -26.08 292.74 1659.95 Coefficients:

 Estimate Std. Error t value Pr(>|t|)

 (Intercept)
 2.331e+03
 7.506e+02
 3.105
 0.002648 **

 Filtrate_manure_pit -1.688e-02
 7.464e-03
 -2.261
 0.026549 *

 Dairy_G
 1.816e-01
 4.507e-02
 4.030
 0.000129 ***

 Parlor
 -4.627e-02
 1.748e-02
 -2.648
 0.009804 **

 ANS_Other
 8.850e-02
 5.779e-02
 1.531
 0.129752

 P_A
 -2.158e-02
 1.501e-02
 -1.438
 0.154510

Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

Residual standard error: 611.3 on 78 degrees of freedom Multiple R-squared: 0.4714, Adjusted R-squared: 0.4375 F-statistic: 13.91 on 5 and 78 DF, p-value: 1.03e-09

> # Analysis of variance for individual feed components > Anova(model.final, Type="II") Anova Table (Type II tests)

 Response: Biogas_production

 Sum Sq Df F value
 Pr(>F)

 Filtrate_manure_pit
 1910236
 1
 5.1119
 0.0265489 *

 Dairy_G
 6069859
 1
 16.2433
 0.0001286 ***

 Parlor
 2619518
 1
 7.0100
 0.0098044 **

 ANS_Other
 876194
 1
 2.3448
 0.1297516

 P_A
 772430
 1
 2.0671
 0.1545103

 Residuals
 29147290
 78

 -- Signif. codes:
 0 '***'
 0.001 '**'
 0.01 '*'
 0.05 '.'
 0.1 '.'
 1



Figure A.1 Final model of MLR for biogas production



Figure A.2 Residual model of MLR for biogas production

B. Radar Chart Codes

Feed amount - Radar analysis## SCAD operation## Wei Liao, October 23, 2021## Fahmi, Noveberm 5, 2021 updated## Wei Liao, December 9, 2021 updated

Load libraries ----library(fmsb)

Choose data file of "FeedAmount_Radar.txt" ----con <-file.choose(new = FALSE)
metadata <- read.table(con, header = T, row.names = 1, fill = TRUE)
head(metadata)</pre>

```
### Set up the bound for the radargraph
data0<-metadata[,c(3,4,5,6,7,8,9,10,11,12,14,15,16,17,18,19,20,21)]
data0
```

maxmin <-data.frame(Digestate=c(max(data0), min(data0)), Filtrate_manure_pit=c(max(data0), min(data0)), SLS Solids=c(max(data0), min(data0)), Dairy_G=c(max(data0), min(data0)), Parlor=c(max(data0), min(data0)), Beef=c(max(data0), min(data0)), W_Feed_manure=c(max(data0), min(data0)), Poultry=c(max(data0), min(data0)), Swine=c(max(data0), min(data0)), ANS_Other=c(max(data0), min(data0)), #Total_manure_pit=c(max(data0), min(data0)), Filtrate food pit=c(max(data0), min(data0)), SLS Solids food=c(max(data0), min(data0)), P A=c(max(data0), min(data0)), Pulp=c(max(data0), min(data0)), FOG=c(max(data0), min(data0)), W_Feed=c(max(data0), min(data0)), Other=c(max(data0), min(data0)), Cart_Food=c(max(data0), min(data0)) #Total food pit=c(max(data0), min(data0)))

maxmin


```
<mark>## 2020 data</mark>
```

```
# select data
data1<-
metadata[which(metadata$Year=="2020"),c(3,4,5,6,7,8,9,10,11,12,14,15,16,17,18,19,20,21)]
data1
```

data1 <- rbind(maxmin, data1)</pre>

```
# Create the radar chart
# Set up the font
windowsFonts(A=windowsFont("Times New Roman")) #Import font
op <- par(family = "A", font =1)</pre>
```

Provide the names of columns

colnames(data1) <-c("Digestate", "Filtrate in manure pit", "SLS solids", "Dairy G", "Parlor", "Beef", "Feed wastes", "Poultry", "Swine", "Other animal wastes", "Filtrate in food pit", "SLS food solids", "P & A", "Food pulp", "FOG", "Food feed wastes", "Others", "Cart food wastes")

radarchart(data1, axistype=2, pty=32, plty=1, axislabcol="grey", na.itp=FALSE, title="", vlcex=1)

legend(x=1.5, y=1, legend = unique(metadata\$Month), title="Months of 2020", bty = "n", pch=20, col=unique(metadata\$Month), cex=1, pt.cex=2)

<mark>## 2019 data</mark>

select data data2<metadata[which(metadata\$Year=="2019"),c(3,4,5,6,7,8,9,10,11,12,14,15,16,17,18,19,20,21)] data2

data2 <- rbind(maxmin, data2)</pre>

Create the radar chart # Set up the font windowsFonts(A=windowsFont("Times New Roman")) #Import font op <- par(family = "A", font =1)</pre>

Provide the names of columns colnames(data2) <-c("Digestate", "Filtrate in manure pit", "SLS solids", "Dairy G", "Parlor", "Beef", "Feed wastes", "Poultry", "Swine", "Other animal wastes", "Filtrate in food pit", "SLS food solids", "P & A", "Food pulp", "FOG", "Food feed wastes", "Others", "Cart food wastes")

radarchart(data2, axistype=2, pty=32, plty=1, axislabcol="grey", na.itp=FALSE, title="", vlcex=1)

legend(x=1.5, y=1, legend = unique(metadata\$Month), title="Months of 2019", bty = "n", pch=20, col=unique(metadata\$Month), cex=1, pt.cex=2)

select data data3<metadata[which(metadata\$Year=="2018"),c(3,4,5,6,7,8,9,10,11,12,14,15,16,17,18,19,20,21)] data3 data3 <- rbind(maxmin, data3)</pre> # Create the radar chart # Set up the font windowsFonts(A=windowsFont("Times New Roman")) #Import font op <- par(family = "A", font = 1)# Provide the names of columns colnames(data3) <-c("Digestate", "Filtrate in manure pit", "SLS solids", "Dairy G", "Parlor", "Beef", "Feed wastes", "Poultry", "Swine", "Other animal wastes", "Filtrate in food pit", "SLS food solids", "P & A", "Food pulp", "FOG", "Food feed wastes", "Others", "Cart food wastes") radarchart(data3, axistype=2, pty=32, plty=1, axislabcol="grey", na.itp=FALSE, title="", vlcex=1) legend(x=1.5, y=1, legend = unique(metadata\$Month), title="Months of 2018", bty = "n", pch=20, col=unique(metadata\$Month), cex=1, pt.cex=2) ## 2017 data # select data data4<metadata[which(metadata\$Year=="2017"),c(3,4,5,6,7,8,9,10,11,12,14,15,16,17,18,19,20,21)] data4 data4 <- rbind(maxmin, data4) # Create the radar chart # Set up the font windowsFonts(A=windowsFont("Times New Roman")) #Import font op <- par(family = "A", font =1)# Provide the names of columns colnames(data4) <-c("Digestate", "Filtrate in manure pit", "SLS solids", "Dairy G", "Parlor", "Beef", "Feed wastes", "Poultry", "Swine", "Other animal wastes", "Filtrate in food pit", "SLS food solids", "P & A", "Food pulp", "FOG", "Food feed wastes", "Others", "Cart food wastes") radarchart(data4, axistype=2, pty=32, plty=1, axislabcol="grey", na.itp=FALSE, title="", vlcex=1)

legend(x=1.5, y=1, legend = unique(metadata\$Month), title="Months of 2017", bty = "n",

pch=20, col=unique(metadata\$Month), cex=1, pt.cex=2)

2016 data
select data
data5<metadata[which(metadata\$Year=="2016"),c(3,4,5,6,7,8,9,10,11,12,14,15,16,17,18,19,20,21)]
data5</pre>

data5 <- rbind(maxmin, data5)</pre>

Create the radar chart # Set up the font windowsFonts(A=windowsFont("Times New Roman")) #Import font op <- par(family = "A", font =1)</pre>

Provide the names of columns

colnames(data5) <-c("Digestate", "Filtrate in manure pit", "SLS solids", "Dairy G", "Parlor", "Beef", "Feed wastes", "Poultry", "Swine", "Other animal wastes", "Filtrate in food pit", "SLS food solids", "P & A", "Food pulp", "FOG", "Food feed wastes", "Others", "Cart food wastes")

radarchart(data5, axistype=2, pty=32, plty=1, axislabcol="grey", na.itp=FALSE, title="", vlcex=1)

legend(x=1.5, y=1, legend = unique(metadata\$Month), title="Months of 2016", bty = "n", pch=20, col=unique(metadata\$Month), cex=1, pt.cex=2)

select data data6<metadata[which(metadata\$Year=="2015"),c(3,4,5,6,7,8,9,10,11,12,14,15,16,17,18,19,20,21)] data6

data6 <- rbind(maxmin, data6)</pre>

Create the radar chart # Set up the font windowsFonts(A=windowsFont("Times New Roman")) #Import font op <- par(family = "A", font =1)</pre>

```
# Provide the names of columns
colnames(data6) <-c("Digestate", "Filtrate in manure pit", "SLS solids", "Dairy G",
"Parlor", "Beef", "Feed wastes", "Poultry", "Swine", "Other animal wastes",
"Filtrate in food pit", "SLS food solids", "P & A",
```

"Food pulp", "FOG", "Food feed wastes", "Others", "Cart food wastes")

radarchart(data6, axistype=2, pty=32, plty=1, axislabcol="grey", na.itp=FALSE, title="", vlcex=1)

legend(x=1.5, y=1, legend = unique(metadata\$Month), title="Months of 2015", bty = "n", pch=20, col=unique(metadata\$Month), cex=1, pt.cex=2)

select data data7<metadata[which(metadata\$Year=="2014"),c(3,4,5,6,7,8,9,10,11,12,14,15,16,17,18,19,20,21)] data7

data7 <- rbind(maxmin, data7)</pre>

Create the radar chart # Set up the font windowsFonts(A=windowsFont("Times New Roman")) #Import font op <- par(family = "A", font =1)</pre>

Provide the names of columns colnames(data7) <-c("Digestate", "Filtrate in manure pit", "SLS solids", "Dairy G", "Parlor", "Beef", "Feed wastes", "Poultry", "Swine", "Other animal wastes", "Filtrate in food pit", "SLS food solids", "P & A", "Food pulp", "FOG", "Food feed wastes", "Others", "Cart food wastes")

radarchart(data7, axistype=2, pty=32, plty=1, axislabcol="grey", na.itp=FALSE, title="", vlcex=1)

legend(x=1.5, y=1, legend = unique(metadata\$Month), title="Months of 2014", bty = "n", pch=20, col=unique(metadata\$Month), cex=1, pt.cex=2)

#######

C. Violin Chart and ANOVA Tukey Multiple Comparison Codes for Feedstock

FULL CODES

Feed amount - Violin analysis
SCAD operation
Wei Liao, December 9, 2021
Fahmi Dwilaksono, December 12, 2021 Updated
Fahmi, Jan 4,2022

```
# Load libraries -----
 library (MASS)
 library(ggplot2)
 library(grid)
 library(gridExtra)
 library(ggpubr)
 library(plyr)
 library(inferr)
# Plot bar chart with standard deviation -----
 #data : a data frame
 #varname : the name of a column containing the variable to be summarized
 #groupnames : vector of column names to be used as
 #grouping variables
 data_summary <- function(data, varname, groupnames){
  require(plyr)
  summary_func <- function(x, col){</pre>
   c(mean = mean(x[[col]], na.rm=TRUE),
     sd = sd(x[[col]], na.rm=TRUE))
  }
  data_sum<-ddply(data, groupnames, .fun=summary_func,
            varname)
  data_sum <- rename(data_sum, c("mean" = varname))</pre>
  return(data sum)
 }
```

```
# Choose data file FeedAmount_Violin.txt -----
con <-file.choose(new = FALSE)
metadata <- read.table(con, header = T, row.names = 1, fill = TRUE)
head(metadata)</pre>
```

Define factors for metadata ----metadata\$Feed_type <- factor(metadata\$Feed_type)
metadata\$Year<- factor(metadata\$Year)
metadata\$Month <- factor(metadata\$Month)</pre>

<mark>#Anova</mark>

#1. Digstate

```
##Data selection
data1<-metadata[which(metadata$Feed_type=="Digestate"),]
data1</pre>
```

```
## Anova
fit1 <- aov(Daily_Feed~Year, data1)</pre>
```

summary(fit1) Tukey1 <- TukeyHSD(fit1, conf.level=0.95) #Tukey multiple comparison Tukey1 #Output Tukey results

Mean and standard deviation

```
box_1_data <- data_summary(data1, varname="Daily_Feed",
groupnames=c("Year"))
```

box_1_data

```
#2. Filtrate manure pit
```

```
##Data selection
data2<-metadata[which(metadata$Feed_type=="Filtrate_manure_pit"),]
data2</pre>
```

```
## Anova
fit2 <- aov(Daily_Feed~Year, data2)
summary(fit2)
Tukey2 <- TukeyHSD(fit2, conf.level=0.95) #Tukey multiple comparison
Tukey2 #Output Tukey results
```

axis.text.y=element_text(size=20, family="A"), axis.title.y = element_text(size = 20, family="A"), axis.title.x=element_text(size=20, family="A"), legend.position = "top") box_2 box_2_1 <- box_2 + geom_boxplot(width=0.2) # Add median and quartile box 2_1

box_2_data

#3. SLS Solids

##Data selection
data3<-metadata[which(metadata\$Feed_type=="SLS_Solids"),]
data3</pre>

Anova fit3 <- aov(Daily_Feed~Year, data3) summary(fit3) Tukey3 <- TukeyHSD(fit3, conf.level=0.95) #Tukey multiple comparison Tukey3 #Output Tukey results

Plot windowsFonts(A=windowsFont("Times New Roman")) #Import font $box_3 <- ggplot(data3, aes(x=Year, y=Daily_Feed)) +$ geom_violin(trim=TRUE, fill="green") + xlab("Year")+ ylab("Daily feed (kg/day)") + labs(title = "SLS Solids", subtitle=NULL) + theme classic() +theme(title=element_text(size=20, family="A"), axis.text.x = element_text(size=20, family="A"), axis.text.y=element text(size=20, family="A"), axis.title.y = element_text(size = 20, family="A"), axis.title.x=element text(size=20, family="A"), legend.position = "top") box 3 box_3_1 <- box_3 + geom_boxplot(width=0.2) # Add median and quartile box 3 1 ## Mean and standard deviation box_3_data <- data_summary(data3, varname="Daily_Feed", groupnames=c("Year"))

box_3_data

<mark>#4. Dairy G</mark>

```
##Data selection
data4<-metadata[which(metadata$Feed_type=="Dairy_G"),]
data4
## Anova
fit4 <- aov(Daily Feed~Year, data4)
summary(fit4)
Tukey4 <- TukeyHSD(fit4, conf.level=0.95) #Tukey multiple comparison
Tukey4 #Output Tukey results
## Plot
windowsFonts(A=windowsFont("Times New Roman")) #Import font
box_4 <- ggplot(data4, aes(x=Year, y=Daily_Feed)) +
 geom violin(trim=TRUE, fill="green") +
 xlab("Year")+
 ylab("Daily feed (kg/day)") + labs(title = "Dairy G", subtitle=NULL) +
 theme classic() +
 theme(title=element_text(size=20, family="A"),
    axis.text.x = element text(size=20, family="A"),
    axis.text.y=element_text(size=20, family="A"),
    axis.title.y = element text(size = 20, family="A"),
    axis.title.x=element_text(size=20, family="A"), legend.position = "top")
box_4
box 4 1 <- box 4 + geom boxplot(width=0.2) # Add median and quartile
box_4_1
```

box_4_data

<mark>#5. Parlor</mark>

```
##Data selection
data5<-metadata[which(metadata$Feed_type=="Parlor"),]
data5</pre>
```

```
## Anova
fit5 <- aov(Daily_Feed~Year, data5)
summary(fit5)
Tukey5 <- TukeyHSD(fit5, conf.level=0.95) #Tukey multiple comparison
Tukey5 #Output Tukey results
```

Plot

```
windowsFonts(A=windowsFont("Times New Roman")) #Import font
box_5 <- ggplot(data5, aes(x=Year, y=Daily_Feed)) +
geom_violin(trim=TRUE, fill="green") +
xlab("Year")+
ylab("Daily feed (kg/day)") + labs(title = "Parlor", subtitle=NULL) +
theme_classic() +
theme(title=element_text(size=20, family="A"),
    axis.text.x = element_text(size=20, family="A"),
    axis.text.y=element_text(size=20, family="A"),
    axis.title.y = element_text(size=20, family="A"),
    axis.title.y = element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"),
    box_5
    box_5_1 <- box_5 + geom_boxplot(width=0.2) # Add median and quartile
    box_5_1
```

box_5_data

<mark>#6. Beef</mark>

```
##Data selection
data6<-metadata[which(metadata$Feed_type=="Beef"),]
data6</pre>
```

```
## Anova
fit6 <- aov(Daily_Feed~Year, data6)
summary(fit6)
Tukey6 <- TukeyHSD(fit6, conf.level=0.95) #Tukey multiple comparison
Tukey6 #Output Tukey results
```

```
## Plot
windowsFonts(A=windowsFont("Times New Roman")) #Import font
box_6 <- ggplot(data6, aes(x=Year, y=Daily_Feed)) +
geom_violin(trim=TRUE, fill="green") +
xlab("Year")+
ylab("Daily feed (kg/day)") + labs(title = "Beef", subtitle=NULL) +
theme_classic() +
theme(title=element_text(size=20, family="A"),
    axis.text.x = element_text(size=20, family="A"),
    axis.text.y=element_text(size=20, family="A"),
    axis.title.y = element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"), legend.position = "top")
box_6
box_6_1 <- box_6 + geom_boxplot(width=0.2) # Add median and quartile</pre>
```

box_6_1

box_6_data

#7. Waste Feed_Manure

##Data selection
data7<-metadata[which(metadata\$Feed_type=="W_Feed_manure"),]
data7</pre>

```
## Anova
fit7 <- aov(Daily_Feed~Year, data7)
summary(fit7)
Tukey7 <- TukeyHSD(fit7, conf.level=0.95) #Tukey multiple comparison
Tukey7 #Output Tukey results
```

Plot

windowsFonts(A=windowsFont("Times New Roman")) #Import font box_7 <- ggplot(data7, aes(x=Year, y=Daily_Feed)) + geom_violin(trim=TRUE, fill="green") + xlab("Year")+ ylab("Daily feed (kg/day)") + labs(title = "W Feed Manure", subtitle=NULL) + theme_classic() + theme(title=element_text(size=20, family="A"), axis.text.x = element_text(size=20, family="A"), axis.text.y=element_text(size=20, family="A"), axis.title.y = element_text(size=20, family="A"), axis.title.x=element_text(size=20, family="A"), axis.title.x=element_text(size=20, family="A"), legend.position = "top") box_7 box_7_1 <- box_7 + geom_boxplot(width=0.2) # Add median and quartile box_7_1

box_7_data

<mark>#8. Poultry</mark>

```
##Data selection
data8<-metadata[which(metadata$Feed_type=="Poultry"),]
data8</pre>
```

Anova fit8 <- aov(Daily_Feed~Year, data8) summary(fit8) Tukey8 <- TukeyHSD(fit8, conf.level=0.95) #Tukey multiple comparison Tukey8 #Output Tukey results

```
## Plot
windowsFonts(A=windowsFont("Times New Roman")) #Import font
box_8 <- ggplot(data8, aes(x=Year, y=Daily_Feed)) +
geom_violin(trim=TRUE, fill="green") +
xlab("Year")+
ylab("Daily feed (kg/day)") + labs(title = "Poultry", subtitle=NULL) +
theme_classic() +
theme(title=element_text(size=20, family="A"),
    axis.text.x = element_text(size=20, family="A"),
    axis.text.y=element_text(size=20, family="A"),
    axis.title.y = element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"),
    box_8
box_8_1 <- box_8 + geom_boxplot(width=0.2) # Add median and quartile
box 8_1
```

Mean and standard deviation box_8_data <- data_summary(data8, varname="Daily_Feed", groupnames=c("Year"))

box_8_data

<mark>#9. Swine</mark>

```
##Data selection
data9<-metadata[which(metadata$Feed_type=="Swine"),]
data9</pre>
```

```
## Anova
fit9 <- aov(Daily_Feed~Year, data9)
summary(fit9)
Tukey9 <- TukeyHSD(fit9, conf.level=0.95) #Tukey multiple comparison
Tukey9 #Output Tukey results
```

```
## Plot
windowsFonts(A=windowsFont("Times New Roman")) #Import font
box_9 <- ggplot(data9, aes(x=Year, y=Daily_Feed)) +
geom_violin(trim=TRUE, fill="green") +
xlab("Year")+
ylab("Daily feed (kg/day)") + labs(title = "Swine", subtitle=NULL) +</pre>
```

```
theme_classic() +
theme(title=element_text(size=20, family="A"),
    axis.text.x = element_text(size=20, family="A"),
    axis.text.y=element_text(size=20, family="A"),
    axis.title.y = element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"), legend.position = "top")
box_9
box_9_1 <- box_9 + geom_boxplot(width=0.2) # Add median and quartile
box_9_1</pre>
```

```
## Mean and standard deviation
```

box_9_data <- data_summary(data9, varname="Daily_Feed", groupnames=c("Year"))

box_9_data

#10. ANS Other

##Data selection
data10<-metadata[which(metadata\$Feed_type=="ANS_Other"),]
data10</pre>

```
## Anova
fit10 <- aov(Daily_Feed~Year, data10)
summary(fit10)
Tukey10 <- TukeyHSD(fit10, conf.level=0.95) #Tukey multiple comparison
Tukey10 #Output Tukey results
```

```
## Plot
windowsFonts(A=windowsFont("Times New Roman")) #Import font
box_10 <- ggplot(data10, aes(x=Year, y=Daily_Feed)) +
geom_violin(trim=TRUE, fill="green") +
xlab("Year")+
ylab("Daily feed (kg/day)") + labs(title = "ANS Other", subtitle=NULL) +
theme_classic() +
theme(title=element_text(size=20, family="A"),
    axis.text.x = element_text(size=20, family="A"),
    axis.text.y=element_text(size=20, family="A"),
    axis.title.y = element_text(size=20, family="A"),
    axis.title.y = element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"),
    box_10
box_10_1 <- box_10 + geom_boxplot(width=0.2) # Add median and quartile
box_10_1
```

box_10_data

#11. Total Manure Pit

##Data selection
data11<-metadata[which(metadata\$Feed_type=="Total_manure_pit"),]
data11</pre>

Anova fit11 <- aov(Daily_Feed~Year, data11) summary(fit11) Tukey11 <- TukeyHSD(fit11, conf.level=0.95) #Tukey multiple comparison Tukey11 #Output Tukey results

Plot

windowsFonts(A=windowsFont("Times New Roman")) #Import font box_11 <- ggplot(data11, aes(x=Year, y=Daily_Feed)) + geom_violin(trim=TRUE, fill="green") + xlab("Year")+ ylab("Daily feed (kg/day)") + labs(title = "Total Manure Pit", subtitle=NULL) + theme_classic() + theme(title=element_text(size=20, family="A"), axis.text.x = element_text(size=20, family="A"), axis.text.y=element_text(size=20, family="A"), axis.title.y = element_text(size=20, family="A"), axis.title.x=element_text(size=20, family="A"), axis.title.x=element_text(size=20, family="A"), axis.title.x=element_text(size=20, family="A"), box_11 box_11_1 <- box_11 + geom_boxplot(width=0.2) # Add median and quartile box_11_1

Mean and standard deviation

box_11_data <- data_summary(data11, varname="Daily_Feed", groupnames=c("Year"))

box_11_data

#12. Filtrate Food Pit

##Data selection
data12<-metadata[which(metadata\$Feed_type=="Filtrate_food_pit"),]
data12</pre>

Anova fit12 <- aov(Daily_Feed~Year, data12) summary(fit12) Tukey12 <- TukeyHSD(fit12, conf.level=0.95) #Tukey multiple comparison Tukey12 #Output Tukey results ## Plot windowsFonts(A=windowsFont("Times New Roman")) #Import font box_12 <- ggplot(data12, aes(x=Year, y=Daily_Feed)) + geom violin(trim=TRUE, fill="green") + xlab("Year")+ ylab("Daily feed (kg/day)") + labs(title = "Filtrate Food Pit", subtitle=NULL) + theme classic() +theme(title=element text(size=20, family="A"), axis.text.x = element_text(size=20, family="A"), axis.text.y=element text(size=20, family="A"), axis.title.y = element_text(size = 20, family="A"), axis.title.x=element text(size=20, family="A"), legend.position = "top") box 12 box_12_1 <- box_12 + geom_boxplot(width=0.2) # Add median and quartile box 12 1 ## Mean and standard deviation

box_12_data <- data_summary(data12, varname="Daily_Feed", groupnames=c("Year"))

box_12_data

#13. SLS Solids Food

##Data selection
data13<-metadata[which(metadata\$Feed_type=="SLS_Solids_food"),]
data13</pre>

```
## Anova
fit13 <- aov(Daily_Feed~Year, data13)
summary(fit13)
Tukey13 <- TukeyHSD(fit13, conf.level=0.95) #Tukey multiple comparison
Tukey13 #Output Tukey results
```

```
## Plot
windowsFonts(A=windowsFont("Times New Roman")) #Import font
box_13 <- ggplot(data13, aes(x=Year, y=Daily_Feed)) +
geom_violin(trim=TRUE, fill="green") +
xlab("Year")+
ylab("Daily feed (kg/day)") + labs(title = "SLS Solids Food", subtitle=NULL) +
theme_classic() +
theme(title=element_text(size=20, family="A"),
    axis.text.x = element_text(size=20, family="A"),
    axis.text.y=element_text(size=20, family="A"),
    axis.title.y = element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"), legend.position = "top")
```

box_13 box_13_1 <- box_13 + geom_boxplot(width=0.2) # Add median and quartile box_13_1

box_13_data

<mark>#14. P_A</mark>

```
##Data selection
data14<-metadata[which(metadata$Feed_type=="P_A"),]
data14</pre>
```

Anova fit14 <- aov(Daily_Feed~Year, data14) summary(fit14) Tukey14 <- TukeyHSD(fit14, conf.level=0.95) #Tukey multiple comparison Tukey14 #Output Tukey results

```
## Plot
windowsFonts(A=windowsFont("Times New Roman")) #Import font
box_14 <- ggplot(data14, aes(x=Year, y=Daily_Feed)) +
geom_violin(trim=TRUE, fill="green") +
xlab("Year")+
ylab("Daily feed (kg/day)") + labs(title = "Pinnapples", subtitle=NULL) +
theme_classic() +
theme(title=element_text(size=20, family="A"),
    axis.text.x = element_text(size=20, family="A"),
    axis.text.y=element_text(size=20, family="A"),
    axis.title.y = element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"),
    legend.position = "top")
box_14
box_14_1 <- box_14 + geom_boxplot(width=0.2) # Add median and quartile
box_14_1
```

<mark>#15. Pulp</mark>

##Data selection data15<-metadata[which(metadata\$Feed_type=="Pulp"),] data15 ## Anova fit15 <- aov(Daily_Feed~Year, data15) summary(fit15) Tukey15 <- TukeyHSD(fit15, conf.level=0.95) #Tukey multiple comparison Tukey15 #Output Tukey results ## Plot windowsFonts(A=windowsFont("Times New Roman")) #Import font box_15 <- ggplot(data15, aes(x=Year, y=Daily_Feed)) + geom violin(trim=TRUE, fill="green") + xlab("Year")+ ylab("Daily feed (kg/day)") + labs(title = "Pulp", subtitle=NULL) + theme classic() +theme(title=element_text(size=20, family="A"), axis.text.x = element text(size=20, family="A"), axis.text.y=element_text(size=20, family="A"), axis.title.y = element text(size = 20, family="A"), axis.title.x=element_text(size=20, family="A"), legend.position = "top") box_15 box 15 1 < box 15 + geom boxplot(width=0.2) # Add median and quartile box_15_1

box_15_data

<mark>#16. FOG</mark>

```
##Data selection
data16<-metadata[which(metadata$Feed_type=="FOG"),]
data16</pre>
```

Anova fit16 <- aov(Daily_Feed~Year, data16) summary(fit16) Tukey16 <- TukeyHSD(fit16, conf.level=0.95) #Tukey multiple comparison Tukey16 #Output Tukey results

Plot

windowsFonts(A=windowsFont("Times New Roman")) #Import font box_16 <- ggplot(data16, aes(x=Year, y=Daily_Feed)) + geom_violin(trim=TRUE, fill="green") + xlab("Year")+ ylab("Daily feed (kg/day)") + labs(title = "FOG", subtitle=NULL) + theme_classic() + theme(title=element_text(size=20, family="A"), axis.text.x = element_text(size=20, family="A"), axis.text.y=element_text(size=20, family="A"), axis.title.y = element_text(size=20, family="A"), axis.title.x=element_text(size=20, family="A"), axis.title.x=element_text(size=20, family="A"), axis.title.x=element_text(size=20, family="A"), box_16_1 <- box_16 + geom_boxplot(width=0.2) # Add median and quartile box_16_1

#17. Waste Feed

##Data selection
data17<-metadata[which(metadata\$Feed_type=="W_Feed"),]
data17</pre>

Anova fit17 <- aov(Daily_Feed~Year, data17) summary(fit17) Tukey17 <- TukeyHSD(fit17, conf.level=0.95) #Tukey multiple comparison Tukey17 #Output Tukey results

```
## Plot
windowsFonts(A=windowsFont("Times New Roman")) #Import font
box_17 <- ggplot(data17, aes(x=Year, y=Daily_Feed)) +
geom_violin(trim=TRUE, fill="green") +
xlab("Year")+
ylab("Daily feed (kg/day)") + labs(title = "Waste Feed", subtitle=NULL) +
theme_classic() +
theme(title=element_text(size=20, family="A"),
    axis.text.x = element_text(size=20, family="A"),
    axis.text.y=element_text(size=20, family="A"),
    axis.title.y = element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"), legend.position = "top")
box_17
box_17_1 <- box_17 + geom_boxplot(width=0.2) # Add median and quartile</pre>
```

box_17_1

#18. Other

```
##Data selection
data18<-metadata[which(metadata$Feed_type=="Other"),]
data18</pre>
```

```
## Anova
fit18 <- aov(Daily_Feed~Year, data18)
summary(fit18)
Tukey18 <- TukeyHSD(fit18, conf.level=0.95) #Tukey multiple comparison
Tukey18 #Output Tukey results
```

Plot

windowsFonts(A=windowsFont("Times New Roman")) #Import font box_18 <- ggplot(data18, aes(x=Year, y=Daily_Feed)) + geom_violin(trim=TRUE, fill="green") + xlab("Year")+ ylab("Daily feed (kg/day)") + labs(title = "Other", subtitle=NULL) + theme_classic() + theme(title=element_text(size=20, family="A"), axis.text.x = element_text(size=20, family="A"), axis.text.y=element_text(size=20, family="A"), axis.title.y = element_text(size=20, family="A"), axis.title.y = element_text(size=20, family="A"), axis.title.x=element_text(size=20, family="A"), box_18 box_18_1

box_18_data

<mark>#19. Cart Food</mark>

```
##Data selection
data19<-metadata[which(metadata$Feed_type=="Cart_Food"),]
data19</pre>
```

Anova fit19 <- aov(Daily_Feed~Year, data19) summary(fit19) Tukey19 <- TukeyHSD(fit19, conf.level=0.95) #Tukey multiple comparison Tukey19 #Output Tukey results ## Plot windowsFonts(A=windowsFont("Times New Roman")) #Import font box $19 \le \text{ggplot}(\text{data19}, \text{aes}(x=\text{Year}, y=\text{Daily Feed})) +$ geom_violin(trim=TRUE, fill="green") + xlab("Year")+ ylab("Daily feed (kg/day)") + labs(title = "Cart Food", subtitle=NULL) + theme classic() +theme(title=element_text(size=20, family="A"), axis.text.x = element_text(size=20, family="A"), axis.text.y=element text(size=20, family="A"), axis.title.y = element_text(size = 20, family="A"), axis.title.x=element text(size=20, family="A"), legend.position = "top") box 19 box_19_1 <- box_19 + geom_boxplot(width=0.2) # Add median and quartile box 19 1

Mean and standard deviation

box_19_data <- data_summary(data19, varname="Daily_Feed", groupnames=c("Year"))

box_19_data

#20. Total Food Pit

```
##Data selection
data20<-metadata[which(metadata$Feed_type=="Total_food_pit"),]
data20</pre>
```

```
## Anova
fit20 <- aov(Daily_Feed~Year, data20)
summary(fit20)
Tukey20 <- TukeyHSD(fit20, conf.level=0.95) #Tukey multiple comparison
Tukey20 #Output Tukey results
```

```
## Plot
windowsFonts(A=windowsFont("Times New Roman")) #Import font
box_20 <- ggplot(data20, aes(x=Year, y=Daily_Feed)) +
geom_violin(trim=TRUE, fill="green") +
xlab("Year")+
ylab("Daily feed (kg/day)") + labs(title = "Total Food Pit", subtitle=NULL) +
theme_classic() +
```

theme(title=element_text(size=20, family="A"),
 axis.text.x = element_text(size=20, family="A"),
 axis.text.y=element_text(size=20, family="A"),
 axis.title.y = element_text(size=20, family="A"),
 axis.title.x=element_text(size=20, family="A"), legend.position = "top")
box_20
box_20_1 <- box_20 + geom_boxplot(width=0.2) # Add median and quartile
box_20_1</pre>

Mean and standard deviation

box_20_data <- data_summary(data20, varname="Daily_Feed", groupnames=c("Year"))

box_20_data

#21. Total Feedstock

##Data selection
data21<-metadata[which(metadata\$Feed_type=="Total_feedstock"),]
data21</pre>

Anova fit21 <- aov(Daily_Feed~Year, data21) summary(fit21) Tukey21 <- TukeyHSD(fit21, conf.level=0.95) #Tukey multiple comparison Tukey21 #Output Tukey results

Plot windowsFonts(A=windowsFont("Times New Roman")) #Import font box_21 <- ggplot(data21, aes(x=Year, y=Daily_Feed)) + geom_violin(trim=TRUE, fill="green") + xlab("Year")+ ylab("Daily feed (kg/day)") + labs(title = "Total Feedstock", subtitle=NULL) + theme_classic() + theme(title=element_text(size=20, family="A"), axis.text.x = element_text(size=20, family="A"), axis.text.y=element_text(size=20, family="A"), axis.title.y = element_text(size=20, family="A"), axis.title.y = element_text(size=20, family="A"), axis.title.x=element_text(size=20, family="A"), axis.title.x=element_text(size=20, family="A"), box_21 box_21_1 <- box_21 + geom_boxplot(width=0.2) # Add median and quartile box_21_1

box_21_data

###

RESULTS

C.1 Violin Chart of Feedstock in Manure Pit C.1.1 Digestate > ## Anova > fit1 <- aov(Daily_Feed~Year, data1) > summary(fit1) Df Sum Sq Mean Sq F value Pr(>F) Year 4 145459560 36364890 0.906 0.476 Residuals 24 963661949 40152581 > Tukey1 <- TukeyHSD(fit1, conf.level=0.95) #Tukey multiple comparison > Tukey1 #Output Tukey results Tukey multiple comparisons of means 95% family-wise confidence level

Fit: aov(formula = Daily_Feed ~ Year, data = data1)

\$Year

diff lwr upr p adj 2016-2015 -1625.8753 -21303.490 18051.740 0.9991656 2017-2015 -5303.0678 -24800.975 14194.839 0.9276618 2018-2015 -4567.2934 -24730.849 15596.262 0.9615453 2020-2015 -9407.0445 -32270.367 13456.278 0.7445301 2017-2016 -3677.1925 -12067.756 4713.371 0.6989719 2018-2016 -2941.4181 -12780.226 6897.389 0.9011895 2020-2016 -7781.1692 -22374.479 6812.141 0.5291867 2018-2017 735.7743 -8738.506 10210.054 0.9993463 2020-2017 -4103.9767 -18454.048 10246.094 0.9145118 2020-2018 -4839.7510 -20081.966 10402.464 0.8802393





Year Daily_Feedsd1 201520755.92NA2 201619130.047853.0403 201715452.855927.5234 201816188.624250.2535 202011348.875349.909

C.1.2 Filtrate Manure Pit

>## Anova > fit2 <- aov(Daily_Feed~Year, data2) > summary(fit2) Df Sum Sq Mean Sq F value Pr(>F)6 1.553e+09 258810814 3.729 0.00392 ** Year Residuals 49 3.401e+09 69412500 ____ Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' '1 > Tukey2 <- TukeyHSD(fit2, conf.level=0.95) #Tukey multiple comparison > Tukey2 #Output Tukey results Tukey multiple comparisons of means 95% family-wise confidence level Fit: $aov(formula = Daily_Feed ~ Year, data = data2)$ \$Year diff lwr upr p adj 2015-2014 -3603.9396 -14059.61 6851.7330 0.9367379 2016-2014 -11798.0489 -22764.05 -832.0469 0.0274157 2017-2014 -10023.4382 -21713.24 1666.3592 0.1376290 2018-2014 -9645.7095 -24432.26 5140.8446 0.4245674 2019-2014 -14760.8341 -26941.32 -2580.3440 0.0085485

2020-2014 -12419.0466 -28950.92 4112.8235 0.2604355 2016-2015 -8194.1092 -19160.11 2771.8928 0.2661668 2017-2015 -6419.4986 -18109.30 5270.2988 0.6269482 2018-2015 -6041.7698 -20828.32 8744.7843 0.8681169 2019-2015 -11156.8944 -23337.38 1023.5956 0.0926584 2020-2015 -8815.1069 -25346.98 7716.7631 0.6582698 2017-2016 1774.6107 -10373.78 13923.0045 0.9993180 2018-2016 2152.3394 -12999.37 17304.0486 0.9994188 2019-2016 -2962.7852 -15584.05 9658.4844 0.9905911 2020-2016 -620.9977 -17480.26 16238.2679 0.9999998 2018-2017 377.7287 -15305.78 16061.2377 1.0000000 2019-2017 -4737.3958 -17992.38 8517.5885 0.9255114 2020-2017 -2395.6084 -19734.38 14943.1632 0.9995043 2019-2018 -5115.1246 -21167.71 10937.4587 0.9561045 2020-2018 -2773.3371 -22334.11 16787.4354 0.9994251 2020-2019 2341.7875 -15331.52 20015.1000 0.9996104 Year Daily Feed sd 1 2014 25025.88 7210.635 2 2015 21421.94 12445.617 3 2016 13227.83 5988.434 4 2017 15002.44 6870.595 5 2018 15380.17 10625.096 6 2019 10265.05 3988.524 7 2020 12606.83 4367.135

C.1.3 SLS Solids

\$ Y ear

diff lwr upr p adj 2016-2015 349.67407 -2280.370 2979.718 0.9983908 2017-2015 274.02501 -1561.700 2109.750 0.9972077 2018-2015 -168.21459 -2531.880 2195.451 0.9999236 2019-2015 244.98504 -1943.344 2433.314 0.9992972 2020-2015 727.06262 -2367.702 3821.827 0.9777973
2017-2016 -75.64907 -2739.198 2587.900 0.9999992 2018-2016 -517.88867 -3569.368 2533.590 0.9948985 2019-2016 -104.68903 -3022.461 2813.083 0.9999975 2020-2016 377.38854 -3269.827 4024.604 0.9995207 2018-2017 -442.23960 -2843.131 1958.651 0.9925485 2019-2017 -29.03997 -2257.525 2199.445 1.0000000 2020-2017 453.03761 -2670.251 3576.326 0.9975627 2019-2018 413.19963 -2266.945 3093.344 0.9967478 2020-2018 895.27721 -2564.775 4355.329 0.9663717 2020-2019 482.07758 -2860.650 3824.805 0.9976286





Year Daily_Feed sd 1 2015 2938.868 1252.5055 2 2016 3288.542 1312.8727 3 2017 3212.893 1673.5432 4 2018 2770.653 474.1477 5 2019 3183.853 949.3934 6 2020 3665.931 1288.7252

C.1.4 Dairy Gutter

Tukey multiple comparisons of means 95% family-wise confidence level

Fit: aov(formula = Daily_Feed ~ Year, data = data4)

\$Year

(liff lwi	r upr	p adj		
2015-2014	1192.7957	6 -549.44	637 2935.	0379 0.379	0328
2016-2014	2400.1064	7 657.86	433 4142.	3486 0.0014	4801
2017-2014	1858.3664	2 116.12	429 3600.	6086 0.0289	9095
2018-2014	1755.5144	4 13.27	230 3497.7	566 0.0470	523
2019-2014	1838.3327	8 96.09	064 3580.5	749 0.0318	566
2020-2014	1224.5850	0 -517.65	5714 2966.	8271 0.347	0729
2016-2015	1207.3107	1 -534.93	8143 2949.	5528 0.364	2678
2017-2015	665.57066	5 -1076.67	147 2407.	8128 0.907	8985
2018-2015	562.71868	3 -1179.52	2346 2304.	9608 0.957	3103
2019-2015	645.53702	2 -1096.70	0512 2387.	7792 0.9194	4903
2020-2015	31.78924	-1710.45	290 1774.0	0314 1.0000	0000
2017-2016	-541.7400	5 -2283.98	8218 1200.	5021 0.964	4623
2018-2016	-644.59203	3 -2386.83	3417 1097.	6501 0.920	0129
2019-2016	-561.77369	9 -2304.0	1583 1180.	4684 0.957	6524
2020-2016	-1175.5214	7 -2917.7	6361 566.	7207 0.396	9598
2018-2017	-102.8519	9 -1845.09	9412 1639.	3901 0.999	9971
2019-2017	-20.03365	-1762.27	578 1722.2	2085 1.0000)000
2020-2017	-633.78142	2 -2376.02	2356 1108.	4607 0.925	8387
2019-2018	82.81834	-1659.42	380 1825.0)605 0.9999	992
2020-2018	-530.92944	4 -2273.17	7157 1211.	3127 0.967	7964
2020-2019	-613.74778	8 -2355.98	8991 1128.	4944 0.935	8943

Year Daily_Feed sd 1 2014 11554.80 2239.1947 2 2015 12747.60 1283.0903 3 2016 13954.91 1216.0525 4 2017 13413.17 1876.9954 5 2018 13310.32 837.6477 6 2019 13393.14 809.3594 7 2020 12779.39 942.6853

C.1.5 Parlor

Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1 > Tukey5 <- TukeyHSD(fit5, conf.level=0.95) #Tukey multiple comparison > Tukey5 #Output Tukey results

Tukey multiple comparisons of means

95% family-wise confidence level

Fit: $aov(formula = Daily_Feed ~ Year, data = data5)$

\$Year

diff lwr upr p adj 2015-2014 1580.2767 -3044.705 6205.2580 0.9443209 2016-2014 - 3837.9931 - 8462.974 786.9881 0.1693431 2017-2014 - 4892.8213 - 9517.803 - 267.8401 0.0311264 2018-2014 -1571.4695 -6196.451 3053.5118 0.9457526 2019-2014 - 3500.9365 - 8125.918 1124.0448 0.2613178 2020-2014 - 4230.8038 - 8855.785 394.1775 0.0953095 2016-2015 -5418.2698 -10043.251 -793.2886 0.0113920 2017-2015 -6473.0980 -11098.079 -1848.1168 0.0011735 2018-2015 -3151.7462 -7776.727 1473.2350 0.3847257 2019-2015 - 5081.2132 - 9706.194 - 456.2319 0.0219480 2020-2015 -5811.0805 -10436.062 -1186.0993 0.0050681 2017-2016 -1054.8282 -5679.809 3570.1530 0.9927745 2018-2016 2266.5236 -2358.458 6891.5049 0.7534924 2019-2016 337.0567 -4287.925 4962.0379 0.9999897 2020-2016 -392.8107 -5017.792 4232.1706 0.9999746 2018-2017 3321.3518 -1303.629 7946.3331 0.3214669 2019-2017 1391.8848 -3233.096 6016.8661 0.9697196 2020-2017 662.0175 -3962.964 5286.9988 0.9994637 2019-2018 - 1929.4670 - 6554.448 2695.5143 0.8663878 2020-2018 - 2659.3343 - 7284.316 1965.6469 0.5912253 2020-2019 -729.8673 -5354.849 3895.1139 0.9990631

Year Daily_Feed sd 1 2014 25669.45 3318.805 2 2015 27249.73 5947.718 3 2016 21831.46 3686.228 4 2017 20776.63 3026.019 5 2018 24097.98 4252.452 6 2019 22168.52 2156.050 7 2020 21438.65 2477.829

C.1.6 Beef

- >## Anova
- > fit6 <- aov(Daily_Feed~Year, data6)</pre>
- > summary(fit6)

Df Sum Sq Mean Sq F value Pr(>F)6 100115286 16685881 5.879 8.31e-05 *** Year Residuals 56 158938879 2838194 ___ Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1 > Tukey6 <- TukeyHSD(fit6, conf.level=0.95) #Tukey multiple comparison > Tukey6 #Output Tukey results Tukey multiple comparisons of means 95% family-wise confidence level Fit: $aov(formula = Daily_Feed ~ Year, data = data6)$ \$Year diff lwr upr p adj 2015-2014 -1583.1873 -5017.726 1851.35126 0.7943863 2016-2014 - 2459.4665 - 4826.559 - 92.37421 0.0367349 2017-2014 - 3991.3828 - 6263.117 - 1719.64909 0.0000309 2018-2014 - 3503.7920 - 5819.357 - 1188.22661 0.0004279 2019-2014 - 2945.2863 - 5217.020 - 673.55254 0.0037558 2020-2014 -3327.0973 -6042.338 -611.85619 0.0073262 2016-2015 -876.2793 -4267.614 2515.05581 0.9850225 2017-2015 -2408.1955 -5733.673 917.28212 0.3045062 2018-2015 - 1920.6047 - 5276.178 1434.96833 0.5861362 2019-2015 -1362.0990 -4687.577 1963.37867 0.8700467 2020-2015 -1743.9100 -5386.788 1898.96820 0.7644217 2017-2016 -1531.9163 -3737.789 673.95606 0.3534888 2018-2016 -1044.3255 -3295.312 1206.66137 0.7894796 2019-2016 -485.8197 -2691.692 1720.05261 0.9935552 2020-2016 -867.6308 -3528.013 1792.75134 0.9525220 2018-2017 487.5908 -1662.895 2638.07671 0.9924675 2019-2017 1046.0965 -1057.120 3149.31328 0.7311977 2020-2017 664.2855 -1911.618 3240.18939 0.9851726 2019-2018 558.5058 -1591.980 2708.99170 0.9846223 2020-2018 176.6947 -2437.947 2791.33618 0.9999928 2020-2019 -381.8111 -2957.715 2194.09284 0.9992895



Figure C.3 Feedstock distribution for Beef

Year Daily_Feed sd 1 2014 6404.719 3618.6704 2 2015 4821.532 2269.4265 3 2016 3945.252 1348.7279 4 2017 2413.336 770.9569 5 2018 2900.927 1055.3299 6 2019 3459.433 929.5379 7 2020 3077.622 256.3979

C.1.7 Waste Feed Manure

> ## Anova > fit7 <- aov(Daily_Feed~Year, data7) > summary(fit7) Df Sum Sq Mean Sq F value Pr(>F) Year 5 23261394 4652279 2.906 0.0472 * Residuals 16 25618186 1601137 ---Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1 > Tukey7 <- TukeyHSD(fit7, conf.level=0.95) #Tukey multiple comparison > Tukey7 #Output Tukey results Tukey multiple comparisons of means 95% family-wise confidence level Fit: aov(formula = Daily_Feed ~ Year, data = data7)

\$Year

diff lwr upr p adj 2015-2014 -3023.9467 -6745.8869 697.9935 0.1491613 2016-2014 -648.2586 -4370.1988 3073.6816 0.9922582 2017-2014 -2638.3935 -5615.9456 339.1587 0.0985172 2019-2014 -2239.7768 -5217.3290 737.7753 0.2056048 2020-2014 -916.8304 -3894.3825 2060.7218 0.9138622 2016-2015 2375.6881 -1701.4931 6452.8693 0.4489565 2017-2015 385.5532 -3025.6613 3796.7677 0.9989874 2019-2015 784.1698 -2627.0447 4195.3844 0.9735442 2020-2015 2107.1163 -1304.0983 5518.3308 0.3888309 2017-2016 -1990.1349 -5401.3494 1421.0796 0.4476599 2019-2016 -1591.5183 -5002.7328 1819.6963 0.6671881 2020-2016 -268.5718 -3679.7864 3142.6427 0.9998256 2019-2017 398.6166 -2180.0192 2977.2525 0.9955334 2020-2019 1322.9464 -1255.6894 3901.5822 0.5784162



Figure C.4 Feedstock distribution for Waste Feed Manure

Year Daily_Feed sd 1 2014 4101.228 2328.1781 2 2015 1077.281 208.4797 3 2016 3452.969 152.3505 4 2017 1462.834 425.8712 5 2019 1861.451 597.4175 6 2020 3184.397 1771.8344

C.1.8 Poultry

- > Tukey8 <- TukeyHSD(fit8, conf.level=0.95) #Tukey multiple comparison
- > Tukey8 #Output Tukey results

Tukey multiple comparisons of means

95% family-wise confidence level

Fit: aov(formula = Daily_Feed ~ Year, data = data8)

\$Year

```
diff lwr upr p adj
2017-2016 507.16122 27.02892 987.2935 0.0380139
2019-2016 1377.78570 831.13103 1924.4404 0.0000800
2020-2016 1392.98103 773.13290 2012.8292 0.0002110
2019-2017 870.62448 347.92265 1393.3263 0.0021976
2020-2017 885.81982 286.98963 1484.6500 0.0050653
2020-2019 15.19533 -638.18197 668.5726 0.9998615
```



Figure C.5 Feedstock distribution for Poultry

Year Daily_Feed sd 1 2016 572.6599 315.0070 2 2017 1079.8211 225.7625 3 2019 1950.4456 0.0000 4 2020 1965.6409 213.9322

C.1.9 Swine

Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1 > Tukey9 <- TukeyHSD(fit9, conf.level=0.95) #Tukey multiple comparison > Tukey9 #Output Tukey results Tukey multiple comparisons of means

95% family-wise confidence level

Fit: aov(formula = Daily_Feed ~ Year, data = data9)

\$Year

diff lwr upr p adj 2015-2014 -8172.480 -23726.627 7381.666 0.3207542 2017-2014 13141.467 -8855.417 35138.352 0.2408659 2018-2014 4667.462 -13292.920 22627.843 0.7773253 2017-2015 21313.948 1233.631 41394.264 0.0400869 2018-2015 12839.942 -2714.204 28394.089 0.0967451 2018-2017 -8474.006 -30470.890 13522.879 0.5388782



Figure C.6 Feedstock distribution for Swine

Year Daily_Feed sd 1 2014 12745.482 7762.501 2 2015 4573.001 3664.656 3 2017 25886.949 NA 4 2018 17412.943 4232.458

C.1.10 ANS Other

Residuals 32 48650416 1520325

Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' '1

> Tukey10 <- TukeyHSD(fit10, conf.level=0.95) #Tukey multiple comparison

> Tukey10 #Output Tukey results

Tukey multiple comparisons of means 95% family-wise confidence level

Fit: $aov(formula = Daily_Feed ~ Year, data = data10)$

\$Year

diff lwr upr p adj 2015-2014 589.27271 -1784.0243 2962.570 0.9852771 2016-2014 286.21655 -2087.0805 2659.514 0.9997286 2017-2014 768.33625 -1660.8107 3197.483 0.9516286 2018-2014 534.89837 -2821.4505 3891.247 0.9986656 2019-2014 1416.00083 -1940.3480 4772.350 0.8347201 2020-2014 2887.73677 514.4398 5261.034 0.0092548 2016-2015 -303.05615 -2240.8451 1634.733 0.9988001 2017-2015 179.06354 -1826.7385 2184.866 0.9999531 2018-2015 -54.37434 -3118.2876 3009.539 1.0000000 2019-2015 826.72812 -2237.1851 3890.641 0.9775607 2020-2015 2298.46406 360.6752 4236.253 0.0118835 2017-2016 482.11970 -1523.6824 2487.922 0.9875485 2018-2016 248.68181 -2815.2315 3312.595 0.9999732 2019-2016 1129.78427 -1934.1290 4193.698 0.9040714 2020-2016 2601.52022 663.7313 4539.309 0.0032201 2018-2017 -233.43788 -3340.8131 2873.937 0.9999831 2019-2017 647.66458 -2459.7106 3755.040 0.9941396 2020-2017 2119.40052 113.5985 4125.203 0.0328451 2019-2018 881.10246 -2994.4753 4756.680 0.9906926 2020-2018 2352.83840 -711.0749 5416.752 0.2259142 2020-2019 1471.73594 -1592.1773 4535.649 0.7370542



Figure C.7 Feedstock distribution for ANS Other

Year Daily_Feed sd 1 2014 157.9634 99.61785 2 2015 747.2361 709.21161 3 2016 444.1800 562.77269 4 2017 926.2997 1015.13921 5 2018 692.8618 736.09366 6 2019 1573.9642 147.53947 7 2020 3045.7002 2272.07072

C.1.11 Total Manure Pit

\$Year

diff lwr upr p adj 2015-2014 -2271.4375 -19272.71 14729.840 0.9996395 2016-2014 -4954.4342 -21955.71 12046.843 0.9741916 2017-2014 -4340.4218 -21341.70 12660.855 0.9868731 2018-2014 -13978.4959 -30979.77 3022.781 0.1776912 2019-2014 -23422.4703 -40423.75 -6421.193 0.0014787 2020-2014 - 26187.7939 - 43189.07 - 9186.517 0.0002514 2016-2015 -2682.9967 -19684.27 14318.281 0.9990631 2017-2015 -2068.9843 -19070.26 14932.293 0.9997901 2018-2015 -11707.0583 -28708.34 5294.219 0.3719687 2019-2015 -21151.0328 -38152.31 -4149.756 0.0057191 2020-2015 -23916.3564 -40917.63 -6915.079 0.0010877 2017-2016 614.0124 -16387.26 17615.290 0.9999998 2018-2016 -9024.0616 -26025.34 7977.216 0.6782326 2019-2016 -18468.0361 -35469.31 -1466.759 0.0244380 2020-2016 -21233.3597 -38234.64 -4232.083 0.0054555 2018-2017 -9638.0740 -26639.35 7363.203 0.6073685 2019-2017 - 19082.0485 - 36083.33 - 2080.771 0.0178006 2020-2017 -21847.3721 -38848.65 -4846.095 0.0038195 2019-2018 -9443.9744 -26445.25 7557.303 0.6300296 2020-2018 -12209.2981 -29210.58 4791.979 0.3214553 2020-2019 -2765.3236 -19766.60 14235.954 0.9988876

Year Daily_Feed sd 1 2014 71283.72 11368.291 2 2015 69012.28 17246.236 3 2016 66329.29 18413.578 4 2017 66943.30 15397.431 5 2018 57305.23 14024.063 6 2019 47861.25 8318.581 7 2020 45095.93 7459.882

C.2 Violin Chart of Feedstock in Food Pit C.2.1 Filtrate Food Pit ## Anova > fit12 <- aov(Daily_Feed~Year, data12) > summary(fit12) Df Sum Sq Mean Sq F value Pr(>F) Year 6 5.047e+08 84113595 2.001 0.0802 . Residuals 58 2.438e+09 42039631 ---Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1 > Tukey12 <- TukeyHSD(fit12, conf.level=0.95) #Tukey multiple comparison > Tukey12 #Output Tukey results Tukey multiple comparisons of means 95% family-wise confidence level

Fit: aov(formula = Daily_Feed ~ Year, data = data12)

d	iff lwr	upr p	o adj	
2015-2014	-178.2101	-8444.244	8087.824	1.0000000
2016-2014	-3729.6602	-11814.00	1 4354.680	0.7947930
2017-2014	158.3603	-8880.207	9196.928	0000000.1
2018-2014	4296.0265	-4742.541	13334.594	0.7714169
2019-2014	-4123.0379	-14024.29	2 5778.217	0.8618493
2020-2014	-4589.5006	-13628.06	8 4449.067	0.7132479
2016-2015	-3551.4501	-11817.48	4 4714.584	0.8435061
2017-2015	336.5704	-8864.867	9538.008 ().9999998
2018-2015	4474.2366	-4727.201	13675.674	0.7523395
2019-2015	-3944.8278	-13994.98	2 6105.326	0.8918034
2020-2015	-4411.2904	-13612.72	8 4790.147	0.7643244
2017-2016	3888.0205	-5150.547	12926.588	0.8427671
2018-2016	8025.6868	-1012.881	17064.254	0.1135614
2019-2016	-393.3777	-10294.632	2 9507.877	0.9999997
2020-2016	-859.8403	-9898.408	8178.727	0.9999465
2018-2017	4137.6662	-5763.588	3 14038.921	0.8598617
2019-2017	-4281.3982	-14975.97	5 6413.179	0.8824651
2020-2017	-4747.8609	-14649.11	5 5153.394	0.7641371
2019-2018	-8419.0644	-19113.64	2 2275.513	0.2154937
2020-2018	-8885.5271	-18786.78	2 1015.727	0.1063931
2020-2019	-466.4627	-11161.040	0 10228.115	0.9999995





Year Daily_Feed sd 1 2014 18804.22 6936.435 2 2015 18626.01 6325.763 3 2016 15074.56 5643.998 4 2017 18962.58 6082.968 5 2018 23100.25 9488.308 6 2019 14681.19 4057.150

7 2020 14214.72 5168.211

C.2.2 SLS Solids Food

>## Anova

> fit13 <- aov(Daily_Feed~Year, data13)

> summary(fit13)

Df Sum Sq Mean Sq F value Pr(>F)

Year 5 2340050 468010 0.472 0.795

Residuals 38 37707433 992301

> Tukey13 <- TukeyHSD(fit13, conf.level=0.95) #Tukey multiple comparison

> Tukey13 #Output Tukey results

Tukey multiple comparisons of means 95% family-wise confidence level

Fit: aov(formula = Daily_Feed ~ Year, data = data13)

\$Year

diff lwr upr p adj 2016-2015 357.09030 -2005.4356 2719.6162 0.9974016 2017-2015 -189.04707 -2525.1752 2147.0811 0.9998748 2018-2015 -231.78551 -2546.5787 2083.0077 0.9996439 2019-2015 197.65271 -2084.7641 2480.0696 0.9998253 2020-2015 129.42492 -2598.5850 2857.4348 0.9999911 2017-2016 - 546.13737 - 1998.2315 905.9568 0.8665252 2018-2016 - 588.87581 - 2006.3913 828.6397 0.8112956 2019-2016 -159.43759 -1523.4425 1204.5674 0.9992436 2020-2016 -227.66538 -2250.8117 1795.4809 0.9993710 2018-2017 -42.73845 -1415.8067 1330.3298 0.9999989 2019-2017 386.69978 -931.0542 1704.4538 0.9489365 2020-2017 318.47198 -1673.7848 2310.7287 0.9966128 2019-2018 429.43823 -850.1118 1708.9883 0.9127921 2020-2018 361.21043 -1605.9855 2328.4064 0.9935241 2020-2019 -68.22780 -1997.2221 1860.7665 0.99999979



Figure C.9 Feedstock distribution for SLS Solids Food

Year Da	sd	
1 2015	3056.303	516.3881
2 2016	3413.393	1016.5215
3 2017	2867.256	930.6221
4 2018	2824.517	1268.9409
5 2019	3253.956	888.4573
6 2020	3185.728	228.3480

C.2.3 Pineapple (PA)

```
>## Anova
```

```
> fit14 <- aov(Daily_Feed~Year, data14)</pre>
```

> summary(fit14)

Df Sum Sq Mean Sq F value Pr(>F)

Year 2 1362345 681173 0.14 0.87

Residuals 26 126273793 4856684

> Tukey14 <- TukeyHSD(fit14, conf.level=0.95) #Tukey multiple comparison

> Tukey14 #Output Tukey results

Tukey multiple comparisons of means

95% family-wise confidence level

Fit: aov(formula = Daily_Feed ~ Year, data = data14)

\$Year

diff lwr upr p adj 2015-2014 -476.3094 -2711.951 1759.332 0.8576849 2016-2014 -221.6931 -3136.613 2693.227 0.9805173 2016-2015 254.6163 -2660.303 3169.536 0.9743880





Year Daily_Feed sd 1 2014 11657.20 2206.999 2 2015 11180.89 2370.419 3 2016 11435.51 1649.753

C.2.4 Pulp

Anova > fit15 <- aov(Daily_Feed~Year, data15) > summary(fit15) Df Sum Sq Mean Sq F value Pr(>F) Year 6 10479022 1746504 8.78 4.63e-07 *** Residuals 67 13326784 198907 ___ Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1 > Tukey15 <- TukeyHSD(fit15, conf.level=0.95) #Tukey multiple comparison > Tukey15 #Output Tukey results Tukey multiple comparisons of means 95% family-wise confidence level Fit: $aov(formula = Daily_Feed ~ Year, data = data15)$ \$Year diff lwr upr p adj 2015-2014 203.13362 -350.31903 756.5863 0.9211267 2016-2014 249.21100 -304.24164 802.6637 0.8161929 2017-2014 950.72196 384.83062 1516.6133 0.0000582 2018-2014 590.38779 36.93514 1143.8404 0.0290603 2019-2014 1026.10070 472.64805 1579.5534 0.0000077 2020-2014 564.94884 -310.13664 1440.0343 0.4477754

2016-2015 46.07739 -507.37526 599.5300 0.9999766 2017-2015 747.58834 181.69700 1313.4797 0.0027771 2018-2015 387.25417 -166.19848 940.7068 0.3493295 2019-2015 822.96709 269.51444 1376.4197 0.0004964 2020-2015 361.81522 -513.27025 1236.9007 0.8687476 2017-2016 701.51095 135.61961 1267.4023 0.0061482 2018-2016 341.17678 -212.27587 894.6294 0.5044607 2019-2016 776.88970 223.43705 1330.3423 0.0011967 2020-2016 315.73783 -559.34764 1190.8233 0.9269321 2018-2017 -360.33417 -926.22551 205.5572 0.4647631 2019-2017 75.37874 -490.51260 641.2701 0.9996339 2020-2017 -385.77312 -1268.77808 497.2318 0.8363818 2019-2018 435.71292 -117.73973 989.1656 0.2177536 2020-2018 -25.43895 -900.52442 849.6465 1.0000000 2020-2019 -461.15187 -1336.23734 413.9336 0.6815731

Year Daily_Feed sd 1 2014 1078.566 332.9565 2 2015 1281.700 167.5954 3 2016 1327.777 440.1656 4 2017 2029.288 466.0263 5 2018 1668.954 574.3428 6 2019 2104.667 558.1254 7 2020 1643.515 469.1668

<u>C.2.5 FOG</u>

95% family-wise confidence level

Fit: $aov(formula = Daily_Feed ~ Year, data = data16)$

\$Year

diff lwr upr p adj 2015-2014 16422.8276 6366.315 26479.340 0.0000869 2016-2014 16203.2890 6146.776 26259.802 0.0001120 2017-2014 13835.6144 3779.102 23892.127 0.0015085 2018-2014 12791.7480 2735.235 22848.261 0.0043489 2019-2014 15421.9768 5365.464 25478.490 0.0002714 2020-2014 21622.0881 11565.575 31678.601 0.0000001 2016-2015 -219.5385 -10276.051 9836.974 1.0000000 2017-2015 -2587.2132 -12643.726 7469.300 0.9863374 2018-2015 - 3631.0796 - 13687.592 6425.433 0.9282912 2019-2015 -1000.8507 -11057.364 9055.662 0.9999353 2020-2015 5199.2605 -4857.252 15255.773 0.7044122 2017-2016 -2367.6746 -12424.188 7688.838 0.9914359 2018-2016 - 3411.5410 - 13468.054 6644.972 0.9461552 2019-2016 -781.3122 -10837.825 9275.201 0.9999850 2020-2016 5418.7990 -4637.714 15475.312 0.6626833 2018-2017 -1043.8664 -11100.379 9012.646 0.9999171 2019-2017 1586.3624 -8470.150 11642.875 0.9990653 2020-2017 7786.4737 -2270.039 17842.987 0.2369671 2019-2018 2630.2288 -7426.284 12686.742 0.9851162 2020-2018 8830.3401 -1226.173 18886.853 0.1232539 2020-2019 6200.1113 -3856.402 16256.624 0.5083279

Year Daily_Feed sd 1 2014 19128.24 2959.616 2 2015 35551.07 8151.082 3 2016 35331.53 8097.009 4 2017 32963.85 7527.845 5 2018 31919.99 4847.275 6 2019 34550.22 14652.833 7 2020 40750.33 5266.926

C.2.6 Waste Feed

Fit: $aov(formula = Daily_Feed ~ Year, data = data17)$

\$Year

diff lwr upr p adj 2015-2014 129.8785 -5409.711 5669.468 0.9895938 2018-2014 -400.3705 -5939.960 5139.219 0.9092011

2018-2015 -530.2490 -7314.833 6254.335 0.8954420





Year Daily_Feed sd 1 2014 2033.302 814.3993 2 2015 2163.180 NA 3 2018 1632.931 NA

C.2.7 Other

Anova > fit18 <- aov(Daily_Feed~Year, data18)</pre> > summary(fit18) Df Sum Sq Mean Sq F value Pr(>F)6 1.515e+09 252446057 17.69 1.34e-11 *** Year Residuals 59 8.421e+08 14272946 ____ Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1 > Tukey18 <- TukeyHSD(fit18, conf.level=0.95) #Tukey multiple comparison > Tukey18 #Output Tukey results Tukey multiple comparisons of means 95% family-wise confidence level Fit: $aov(formula = Daily_Feed ~ Year, data = data18)$ \$Year diff lwr upr p adj 2015-2014 -10337.815271 -17781.398 -2894.233 0.0014831 2016-2014 -11858.180056 -19301.762 -4414.598 0.0001748 2017-2014 -12134.719979 -20288.756 -3980.684 0.0005340 2018-2014 -11700.216642 -19143.799 -4256.634 0.0002197 2019-2014 -10330.305803 -18018.005 -2642.607 0.0023257

2020-2014	118.311914 -7325.270 7561.894 1.0000000
2016-2015	-1520.364785 -6228.100 3187.370 0.9551763
2017-2015	-1796.904708 -7562.679 3968.869 0.9622032
2018-2015	-1362.401371 -6070.136 3345.333 0.9737142
2019-2015	7.509468 -5077.425 5092.444 1.0000000
2020-2015	$10456.127185 5748.392 \ 15163.862 \ 0.0000001$
2017-2016	-276.539923 -6042.314 5489.234 0.9999991
2018-2016	157.963414 -4549.771 4865.698 0.9999999
2019-2016	1527.874253 -3557.060 6612.809 0.9683570
2020-2016	11976.491970 7268.757 16684.227 0.0000000
2018-2017	434.503337 -5331.271 6200.277 0.9999866
2019-2017	1804.414176 -4273.245 7882.074 0.9701660
2020-2017	12253.031893 6487.258 18018.806 0.0000004
2019-2018	1369.910839 -3715.024 6454.846 0.9816758
2020-2018	11818.528556 7110.794 16526.263 0.0000000
2020-2019	10448.617717 5363.683 15533.552 0.0000009





Year Da	ally_Feed	sd
1 2014	13209.657	10452.4314
2 2015	2871.842	5042.0168
3 2016	1351.477	853.5117
4 2017	1074.937	216.1485
5 2018	1509.441	996.2938
6 2019	2879.352	1944.4467
7 2020	13327.969	5174.6180

C.2.8 Cart Food

Fit: aov(formula = Daily_Feed ~ Year, data = data19) \$Year

diff lwr upr p adj 2015-2014 -388.577147 -1269.6485 492.4942 0.8331802 2016-2014 187.560292 -693.5110 1068.6316 0.9950230 2017-2014 -2126.439296 -3007.5106 -1245.3680 0.0000000 2018-2014 -2130.332627 -3011.4039 -1249.2613 0.0000000 2019-2014 -2155.620381 -3036.6917 -1274.5491 0.0000000 2020-2014 -2180.945935 -3062.0172 -1299.8746 0.0000000 2016-2015 576.137439 -304.9339 1457.2088 0.4357217 2017-2015 -1737.862149 -2618.9335 -856.7908 0.0000014 2018-2015 -1741.755481 -2622.8268 -860.6842 0.0000013 2019-2015 -1767.043235 -2648.1145 -885.9719 0.0000009 2020-2015 -1792.368788 -2673.4401 -911.2975 0.0000006 2017-2016 -2313.999588 -3195.0709 -1432.9283 0.0000000 2018-2016 -2317.892919 -3198.9642 -1436.8216 0.0000000 2019-2016 -2343.180673 -3224.2520 -1462.1094 0.0000000 2020-2016 -2368.506227 -3249.5775 -1487.4349 0.0000000 2018-2017 -3.893331 -884.9646 877.1780 1.0000000 2019-2017 -29.181085 -910.2524 851.8902 0.9999999 2020-2017 -54.506639 -935.5780 826.5647 0.99999961 2019-2018 -25.287754 -906.3591 855.7836 1.0000000 2020-2018 -50.613307 -931.6846 830.4580 0.9999975 2020-2019 -25.325553 -906.3969 855.7458 1.0000000



Figure C.13 Feedstock distribution for Cart Food

Year Daily_Feed sd 1 2014 2248.64454 949.06204 2 2015 1860.06739 1201.71795 3 2016 2436.20483 1097.05343 4 2017 122.20524 57.59513 5 2018 118.31191 59.13605 6 2019 93.02416 32.70010 7 2020 67.69861 22.86615

C.2.9 Total Food Pit

\$Year

diff lwr upr p adj 2015-2014 13781.6369 -4073.381 31636.6548 0.2402617 2016-2014 5834.7051 -12020.313 23689.7229 0.9548619 2017-2014 -6451.4390 -24306.457 11403.5788 0.9280614 2018-2014 -3507.1355 -21362.153 14347.8823 0.9967990 2019-2014 -7225.5316 -25080.549 10629.4863 0.8821913 2020-2014 8102.4005 -9752.617 25957.4184 0.8138326 2016-2015 -7946.9318 -25801.950 9908.0860 0.8271496 2017-2015 -20233.0759 -38088.094 -2378.0581 0.0161612 2018-2015 -17288.7725 -35143.790 566.2454 0.0640479 2019-2015 -21007.1685 -38862.186 -3152.1506 0.0108739 2020-2015 -5679.2364 -23534.254 12175.7814 0.9603379 2017-2016 -12286.1441 -30141.162 5568.8737 0.3728430 2018-2016 -9341.8406 -27196.858 8513.1772 0.6927108 2019-2016 -13060.2367 -30915.255 4794.7812 0.2999843 2020-2016 2267.6954 -15587.322 20122.7133 0.9997311 2018-2017 2944.3035 -14910.714 20799.3213 0.9987976 2019-2017 -774.0925 -18629.110 17080.9253 0.9999995 2020-2017 14553.8395 -3301.178 32408.8574 0.1857086 2019-2018 -3718.3960 -21573.414 14136.6218 0.9955849 2020-2018 11609.5360 -6245.482 29464.5539 0.4427454 2020-2019 15327.9321 -2527.086 33182.9499 0.1406406

Year Daily_Feed sd 1 2014 56727.27 13338.33 2 2015 70508.91 11747.97 3 2016 62561.98 15610.50 4 2017 50275.83 14544.59 5 2018 53220.14 14366.86 6 2019 49501.74 19066.94 7 2020 64829.67 10934.65

C.2.10 Total Feedstock

Anova > fit21 <- aov(Daily_Feed~Year, data21) > summary(fit21) Df Sum Sq Mean Sq F value Pr(>F) Year 6 1.47e+10 2.451e+09 5.736 5.68e-05 *** Residuals 77 3.29e+10 4.273e+08 ---Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1 > Tukey21 <- TukeyHSD(fit21, conf.level=0.95) #Tukey multiple comparison > Tukey21 #Output Tukey results Tukey multiple comparisons of means 95% family-wise confidence level Fit: aov(formula = Daily_Feed ~ Year, data = data21) \$Year

diff lwr upr p adj

2015-2014 11511.2200 -14038.03 37060.475 0.8188421 2016-2014 881.2536 -24668.00 26430.509 0.9999999 2017-2014 -10790.8025 -36340.06 14758.452 0.8595263 2018-2014 -17484.5730 -43033.83 8064.682 0.3795427 2019-2014 - 30647.0569 - 56196.31 - 5097.802 0.0087674 2020-2014 - 18084.4862 - 43633.74 7464.769 0.3386491 2016-2015 -10629.9663 -36179.22 14919.289 0.8678570 2017-2015 -22302.0224 -47851.28 3247.232 0.1276646 2018-2015 - 28995.7930 - 54545.05 - 3446.538 0.0159145 2019-2015 -42158.2768 -67707.53 -16609.022 0.0000713 2020-2015 - 29595.7062 - 55144.96 - 4046.451 0.0128581 2017-2016 -11672.0561 -37221.31 13877.199 0.8090433 2018-2016 -18365.8267 -43915.08 7183.428 0.3203061 2019-2016 - 31528.3105 - 57077.57 - 5979.056 0.0063055 2020-2016 -18965.7399 -44514.99 6583.515 0.2831181 2018-2017 -6693.7706 -32243.03 18855.484 0.9849832 2019-2017 - 19856.2544 - 45405.51 5693.000 0.2330349 2020-2017 -7293.6838 -32842.94 18255.571 0.9767234 2019-2018 -13162.4838 -38711.74 12386.771 0.7078275 2020-2018 -599.9132 -26149.17 24949.342 1.0000000 2020-2019 12562.5706 -12986.68 38111.826 0.7505992

Year Daily_Feed sd 1 2014 128010.01 16625.39 2 2015 139521.23 22888.56 3 2016 128891.27 25993.66 4 2017 117219.21 16957.01 5 2018 110525.44 21605.39 6 2019 97362.96 23159.83 7 2020 109925.53 14971.79

D. Violin Chart and ANOVA Tukey Multiple Comparison Codes for Digester Output Parameters

FULL CODES

Biogas - Violin analysis ## SCAD operation ## Wei Liao, December 9, 2021 ## Fahmi Dwilaksono, December 12, 2021 # Load libraries ----library (MASS) library(ggplot2) library(ggplot2) library(grid) library(gridExtra) library(ggpubr) library(plyr) library(inferr)

```
# Plot bar chart with standard deviation -----
 #data : a data frame
 #varname : the name of a column containing the variable to be summarized
 #groupnames : vector of column names to be used as
 #grouping variables
 data_summary <- function(data, varname, groupnames){
  require(plyr)
  summary_func <- function(x, col){</pre>
   c(mean = mean(x[[col]], na.rm=TRUE),
     sd = sd(x[[col]], na.rm=TRUE))
  }
  data_sum<-ddply(data, groupnames, .fun=summary_func,
            varname)
  data_sum <- rename(data_sum, c("mean" = varname))</pre>
  return(data sum)
 }
# Choose data file Biogas_Violin.txt -----
con <- file.choose(new = FALSE)
metadata <- read.table(con, header = T, row.names = 1, fill = TRUE)
head(metadata)
```

```
# Define factors for metadata -----
metadata$Year<- factor(metadata$Year)
metadata$Month <- factor(metadata$Month)</pre>
```

```
#Anova
#1. Biogas produciton
```

```
## Anova
fit1 <- aov(Biogas_production~Year, metadata)
summary(fit1)
Tukey1 <- TukeyHSD(fit1, conf.level=0.95) #Tukey multiple comparison
Tukey1 #Output Tukey results</pre>
```

```
## Plot
windowsFonts(A=windowsFont("Times New Roman")) #Import font
box_1 <- ggplot(metadata, aes(x=Year, y=Biogas_production)) +
geom_violin(trim=TRUE, fill="green") +
xlab("Year")+
ylab("Biogas production (m3/day)") + labs(title = "Biogas production", subtitle=NULL) +
theme_classic() +
theme(title=element_text(size=20, family="A"),
axis.text.x = element_text(size=20, family="A"),
axis.text.y=element_text(size=20, family="A"),
```

axis.title.y = element_text(size = 20, family="A"), axis.title.x=element_text(size=20, family="A"), legend.position = "top") box_1 box_1_1 <- box_1 + geom_boxplot(width=0.2) # Add median and quartile box_1_1

Mean and standard deviation

box_1_data <- data_summary(metadata, varname="Biogas_production", groupnames=c("Year"))

box_1_data

#2. Methane content

Anova

fit2 <- aov(Methane_content~Year, metadata) summary(fit2) Tukey2 <- TukeyHSD(fit2, conf.level=0.95) #Tukey multiple comparison Tukey2 #Output Tukey results

Plot

windowsFonts(A=windowsFont("Times New Roman")) #Import font box_2 <- ggplot(metadata, aes(x=Year, y=Methane_content)) + geom_violin(trim=TRUE, fill="green") + xlab("Year")+ ylab("Methane content (%)") + labs(title = "Methane content", subtitle=NULL) + theme_classic() + theme(title=element_text(size=20, family="A"), axis.text.x = element_text(size=20, family="A"), axis.text.y=element_text(size=20, family="A"), axis.title.y = element_text(size=20, family="A"), axis.title.x=element_text(size=20, family="A"), axis.title.x=element_text(size=20, family="A"), legend.position = "top") box_2 box_2_1 <- box_2 + geom_boxplot(width=0.2) # Add median and quartile box_2_1

box_2_data

#3. H2S

```
## Anova
fit3 <- aov(H2S_content~Year, metadata)
summary(fit3)
Tukey3 <- TukeyHSD(fit3, conf.level=0.95) #Tukey multiple comparison
```

Tukey3 #Output Tukey results

```
## Plot
windowsFonts(A=windowsFont("Times New Roman")) #Import font
box_3 <- ggplot(metadata, aes(x=Year, y=H2S_content)) +
geom_violin(trim=TRUE, fill="green") +
xlab("Year")+
ylab("H2S content (ppmv)") + labs(title = "H2S content", subtitle=NULL) +
theme_classic() +
theme(title=element_text(size=20, family="A"),
    axis.text.x = element_text(size=20, family="A"),
    axis.text.y=element_text(size=20, family="A"),
    axis.title.y = element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"), legend.position = "top")
box_3
box_3_1 <- box_3 + geom_boxplot(width=0.2) # Add median and quartile
box_3_1
```

Mean and standard deviation

box_3_data <- data_summary(metadata, varname="H2S_content", groupnames=c("Year"))

box_3_data

#4. Electricity

```
## Anova
fit4 <- aov(Electricity~Year, metadata)
summary(fit4)
Tukey4 <- TukeyHSD(fit4, conf.level=0.95) #Tukey multiple comparison
Tukey4 #Output Tukey results
```

```
## Plot
windowsFonts(A=windowsFont("Times New Roman")) #Import font
box_4 <- ggplot(metadata, aes(x=Year, y=Electricity)) +
geom_violin(trim=TRUE, fill="green") +
xlab("Year")+
ylab("Electricity (kWh)") + labs(title = "Electricity", subtitle=NULL) +
theme_classic() +
theme(title=element_text(size=20, family="A"),
    axis.text.x = element_text(size=20, family="A"),
    axis.text.y=element_text(size=20, family="A"),
    axis.title.y = element_text(size=20, family="A"),
    axis.title.y = element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"),
    box_4
box_4_1 <- box_4 + geom_boxplot(width=0.2) # Add median and quartile
box_4_1
```

box_4_data

#5. Effluent_TS

Anova fit5 <- aov(Effluent_TS~Year, metadata) summary(fit5) Tukey5 <- TukeyHSD(fit5, conf.level=0.95) #Tukey multiple comparison Tukey5 #Output Tukey results

Plot

windowsFonts(A=windowsFont("Times New Roman")) #Import font box_5 <- ggplot(metadata, aes(x=Year, y=Effluent_TS)) + geom_violin(trim=TRUE, fill="green") + xlab("Year")+ ylab("Effluent_TS (mg/L)") + labs(title = "Effluent_TS", subtitle=NULL) + theme_classic() + theme(title=element_text(size=20, family="A"), axis.text.x = element_text(size=20, family="A"), axis.text.y=element_text(size=20, family="A"), axis.title.y = element_text(size=20, family="A"), axis.title.y = element_text(size=20, family="A"), axis.title.x=element_text(size=20, family="A"), box_5 box_5_1 <- box_5 + geom_boxplot(width=0.2) # Add median and quartile box_5_1

Mean and standard deviation

```
box_5_data <- data_summary(metadata, varname="Effluent_TS",
groupnames=c("Year"))
```

box_5_data

#6. Effluent_TN

Anova fit6 <- aov(Effluent_TN~Year, metadata) summary(fit6) Tukey6 <- TukeyHSD(fit6, conf.level=0.95) #Tukey multiple comparison Tukey6 #Output Tukey results

```
## Plot
windowsFonts(A=windowsFont("Times New Roman")) #Import font
box_6 <- ggplot(metadata, aes(x=Year, y=Effluent_TN)) +</pre>
```

geom_violin(trim=TRUE, fill="green") +
xlab("Year")+
ylab("Effluent_TN (mg/L)") + labs(title = "Effluent_TN", subtitle=NULL) +
theme_classic() +
theme(title=element_text(size=20, family="A"),
 axis.text.x = element_text(size=20, family="A"),
 axis.text.y=element_text(size=20, family="A"),
 axis.title.y = element_text(size=20, family="A"),
 axis.title.x=element_text(size=20, family="A"),
 legend.position = "top")
box_6
box_6_1 <- box_6 + geom_boxplot(width=0.2) # Add median and quartile
box_6_1</pre>

Mean and standard deviation

box_6_data <- data_summary(metadata, varname="Effluent_TN", groupnames=c("Year"))

box_6_data

#7. Effluent_TP

Anova fit7 <- aov(Effluent_TP~Year, metadata) summary(fit7) Tukey7 <- TukeyHSD(fit7, conf.level=0.95) #Tukey multiple comparison Tukey7 #Output Tukey results

```
## Plot
windowsFonts(A=windowsFont("Times New Roman")) #Import font
box_7 <- ggplot(metadata, aes(x=Year, y=Effluent_TP)) +
geom_violin(trim=TRUE, fill="green") +
xlab("Year")+
ylab("Effluent_TP (mg/L)") + labs(title = "Effluent_TP", subtitle=NULL) +
theme_classic() +
theme(title=element_text(size=20, family="A"),
    axis.text.x = element_text(size=20, family="A"),
    axis.text.y=element_text(size=20, family="A"),
    axis.title.y = element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"), legend.position = "top")
box_7
box_7_1 <- box_7 + geom_boxplot(width=0.2) # Add median and quartile
box_7_1
```

box_7_data

#8. Effluent_pH

Anova fit8 <- aov(Effluent_pH~Year, metadata) summary(fit8) Tukey8 <- TukeyHSD(fit8, conf.level=0.95) #Tukey multiple comparison Tukey8 #Output Tukey results

Plot windowsFonts(A=windowsFont("Times New Roman")) #Import font box_8 <- ggplot(metadata, aes(x=Year, y=Effluent_pH)) + geom_violin(trim=TRUE, fill="green") + xlab("Year")+ ylab("Effluent_pH") + labs(title = "Effluent_pH", subtitle=NULL) + theme_classic() + theme(title=element_text(size=20, family="A"), axis.text.x = element_text(size=20, family="A"), axis.text.y=element_text(size=20, family="A"), axis.title.y = element_text(size=20, family="A"), axis.title.y = element_text(size=20, family="A"), axis.title.x=element_text(size=20, family="A"), box_8 box_8_1 <- box_8 + geom_boxplot(width=0.2) # Add median and quartile box_8_1

#9. Effluent_VFA

```
## Anova
fit9 <- aov(Effluent_VFA~Year, metadata)
summary(fit9)
Tukey9 <- TukeyHSD(fit9, conf.level=0.95) #Tukey multiple comparison
Tukey9 #Output Tukey results
```

```
## Plot
windowsFonts(A=windowsFont("Times New Roman")) #Import font
box_9 <- ggplot(metadata, aes(x=Year, y=Effluent_VFA)) +
geom_violin(trim=TRUE, fill="green") +
xlab("Year")+
ylab("Effluent_VFA (mg/L)") + labs(title = "Effluent_VFA", subtitle=NULL) +
theme_classic() +
```

```
theme(title=element_text(size=20, family="A"),
    axis.text.x = element_text(size=20, family="A"),
    axis.text.y=element_text(size=20, family="A"),
    axis.title.y = element_text(size=20, family="A"),
    axis.title.x=element_text(size=20, family="A"), legend.position = "top")
box_9
box_9_1 <- box_9 + geom_boxplot(width=0.2) # Add median and quartile
box_9_1</pre>
```

Mean and standard deviation

box_9_data <- data_summary(metadata, varname="Effluent_TP", groupnames=c("Year"))

box_9_data

RESULTS

D.1 Biogas Production

Year 6 20642104 3440351 7.678 1.76e-06 *** Residuals 77 34500609 448060

---Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

> Tukey1 <- TukeyHSD(fit1, conf.level=0.95) #Tukey multiple comparison

> Tukey1 #Output Tukey results

Tukey multiple comparisons of means 95% family-wise confidence level

Fit: aov(formula = Biogas_production ~ Year, data = metadata)

	diff	lwr	upr	p adj	
2015-201	4 694	.00000	-133.36	997 152	21.3700 0.1599213
2016-201	4 1541	1.66667	714.29	669 23	69.0366 0.0000055
2017-201	4 1308	8.91667	481.54	669 21	36.2866 0.0001566
2018-201	4 1358	8.50000	531.13	003 21	85.8700 0.0000784
2019-201	4 1128	8.33333	300.96	336 19	55.7033 0.0017106
2020-201	4 1334	4.41667	507.04	669 21	61.7866 0.0001099
2016-201	5 847	.66667	20.296	69 167	5.0366 0.0410426
2017-201	5 614	.91667	-212.45	331 144	42.2866 0.2817500
2018-201	5 664	.50000	-162.86	997 149	91.8700 0.1998249
2019-201	5 434	.33333	-393.03	664 126	51.7033 0.6893826
2020-201	5 640	.41667	-186.95	331 146	67.7866 0.2372841
2017-201	6 -232	.75000	-1060.11	.997 5	94.6200 0.9783861
2018-201	6 - 183	.16667	-1010.53	664 64	44.2033 0.9938444
2019-201	6 -413	.33333	-1240.70	0331 4	14.0366 0.7365206
2020-201	6 - 207	.25000	-1034.61	.997 62	20.1200 0.9881153
2018-201	7 49.	58333 -	-777.786	64 876	5.9533 0.9999968
2019-201	7 -180	.58333	-1007.95	5331 64	46.7866 0.9942997

2020-2017 25.50000 -801.86997 852.8700 0.9999999 2019-2018 -230.16667 -1057.53664 597.2033 0.9795748 2020-2018 -24.08333 -851.45331 803.2866 1.000000 2020-2019 206.08333 -621.28664 1033.4533 0.9884622

Year Biogas_production sd

1 2014	2330.500 553.7258
2 2015	3024.500 1143.9877
3 2016	3872.167 794.9921
4 2017	3639.417 702.3386
5 2018	3689.000 199.2326
6 2019	3458.833 523.6508
7 2020	3664.917 286.1870

D.2 Methane Content

Df Sum Sq Mean Sq F value Pr(>F) Year 6 268.5 44.75 10.79 1.11e-08 *** Residuals 77 319.3 4.15 ---Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

> Tukey2 <- TukeyHSD(fit2, conf.level=0.95) #Tukey multiple comparison

> Tukey2 #Output Tukey results

Tukey multiple comparisons of means 95% family-wise confidence level

Fit: aov(formula = Methane_content ~ Year, data = metadata)

	diff	lwr	upr	p a	ıdj			
2015-2014	4 -2.750	00000	-5.26714	4655	-0.2328	8534	0.0231	534
2016-2014	4 -1.16	66667	-3.6838	1322	1.3504	1799	0.7984	379
2017-2014	4 1.666	66667	-0.85047	7989	4.1838	8132 (0.42024	415
2018-2014	4 2.250	00000	-0.26714	4655	4.7671	466	0.1105	162
2019-2014	4 0.666	56667	-1.85047	7989	3.1838	3132	0.9841	179
2020-2014	4 2.583	33333	0.06618	678	5.1004	799 (0.04045	567
2016-2013	5 1.583	33333	-0.93381	322	4.1004	799 (0.48372	268
2017-2013	5 4.416	66667	1.89952	2011	6.9338	132 (0.00002	206
2018-2013	5 5.000	00000	2.48285	345	7.5171	466 (0.00000)12
2019-2013	5 3.416	56667	0.89952	2011	5.9338	132 (0.00182	285
2020-2013	5 5.333	33333	2.81618	678	7.8504	799 (0.00000)02
2017-201	5 2.833	33333	0.31618	678	5.3504	799 (0.01729	959
2018-201	5 3.416	66667	0.89952	.011	5.9338	132 (0.00182	285
2019-201	5 1.833	33333	-0.68381	322	4.3504	799	0.30494	407
2020-201	5 3.750	00000	1.23285	345	6.2671	466 (0.00044	418
2018-2017	7 0.583	33333	-1.93381	322	3.1004	799	0.99212	290
2019-2017	7 -1.000	00000	-3.51714	4655	1.5171	466	0.8910	075

2020-2017 0.9166667 -1.60047989 3.4338132 0.9254758 2019-2018 -1.5833333 -4.10047989 0.9338132 0.4837268 2020-2018 0.3333333 -2.18381322 2.8504799 0.9996575 2020-2019 1.9166667 -0.60047989 4.4338132 0.2548637

Year Methane_content sd

1 2014	63.58333 2.9374799
2 2015	60.83333 2.3290003
3 2016	62.41667 2.2746961
4 2017	65.25000 1.5447860
5 2018	65.83333 0.9374369
6 2019	64.25000 0.9653073
7 2020	66.16667 2.3677121

D.3 H₂S Content

Df Sum Sq Mean Sq F value Pr(>F) Year 6 1130830 188472 0.979 0.445 Residuals 77 14823077 192507 > Tukey3 <- TukeyHSD(fit3, conf.level=0.95) #Tukey multiple comparison > Tukey3 #Output Tukey results

Tukey multiple comparisons of means 95% family-wise confidence level

Fit: aov(formula = H2S_content ~ Year, data = metadata)

	diff	lwr	upr	p	adj			
2015-201	4 71.	83333	-470.4	866	614.	1532	0.999	96570
2016-201	4 307	.66667	-234.0	5532	849	.9866	0.60	65461
2017-201	4 61.	16667	-481.1	532	603.	4866	0.999	98653
2018-201	4 27.	66667	-514.6	532	569.	9866	0.999	99988
2019-201	4 160	.83333	-381.4	4866	703	.1532	0.97	18399
2020-201	4 292	.00000	-250.3	3199	834	.3199	0.66	34742
2016-201	5 235	.83333	-306.4	4866	778	.1532	0.84	21465
2017-201	5 -10	.66667	-552.9	9866	531.	6532	1.000	00000
2018-201	5 -44	.16667	-586.4	866	498.	1532	0.999	99802
2019-201	5 89.	00000	-453.3	199	631.	3199	0.998	38299
2020-201	5 220	.16667	-322.	1532	762	.4866	0.88	06089
2017-201	6 -246	5.50000	-788.	8199	295	.8199	0.81	26780
2018-201	6 -280	0.00000	-822.	3199	262	.3199	0.70	57319
2019-201	6 -146	5.83333	-689.	1532	395	.4866	5 0.98	22128
2020-201	6 -15	.66667	-557.9	9866	526.	6532	1.000	00000
2018-201	7 -33	.50000	-575.8	8199	508.	8199	0.999	99962
2019-201	7 99.	66667	-442.6	532	641.	9866	0.997	77881
2020-201	7 230	.83333	-311.4	4866	773	.1532	0.85	50638
2019-201	8 133	.16667	-409.	1532	675	.4866	0.98	92975

2020-2018 264.33333 -277.9866 806.6532 0.7581408 2020-2019 131.16667 -411.1532 673.4866 0.9901197

Year H2S_content sd 1 2014 359.8333 273.7334 2 2015 431.6667 291.7634 3 2016 667.5000 614.5482 4 2017 421.0000 214.4625 5 2018 387.5000 290.1907 6 2019 520.6667 512.9892 7 2020 651.8333 645.3405

D.4 Electricity

> ## Anova > fit4 <- aov(Electricity~Year, metadata) > summary(fit4) Df Sum Sq Mean Sq F value Pr(>F) Year 6 58173817 9695636 7.397 3.61e-06 *** Residuals 71 93063092 1310748 ---Signif. codes: 0 `***` 0.001 `**` 0.01 `*` 0.05 `.` 0.1 ` ` 1 > Tukey4 <- TukeyHSD(fit4, conf.level=0.95) #Tukey multiple comparison > Tukey4 #Output Tukey results Tukey multiple comparisons of means

95% family-wise confidence level

Fit: aov(formula = Electricity ~ Year, data = metadata)

	diff	lwr	upr	p adj			
2015-201	4 1550	.33333	132.03	68 296	8.6299	0.02314	196
2016-201	4 2526	.72619	874.46	507 417	8.9917	0.00030)11
2017-201	4 2114	.03788	663.86	55 356	4.2102	0.00065	559
2018-201	4 2462	.00000	1043.70	035 388	30.2965	0.0000	284
2019-201	4 1706	.41667	288.12	201 312	4.7132	0.00858	360
2020-201	4 2637	.08333	1218.73	868 405	55.3799	0.0000	065
2016-201	5 976.	39286	-675.87	26 2628	8.6583	0.55698	49
2017-201	5 563.	70455	-886.46	78 2013	3.8769	0.89954	23
2018-201	5 911.	66667	-506.62	99 2329	9.9632	0.45479	83
2019-201	5 156.	08333 -	1262.21	32 157	4.3799	0.9998	810
2020-201	5 1086	.75000	-331.54	465 250	5.0465	0.2467	925
2017-201	6 -412.	68831 -	-2092.39	954 126	57.0188	0.9890	915
2018-201	6 -64.7	72619 -	1716.99	17 158	7.5393	0.99999	97
2019-201	6 -820.	30952 -	-2472.57	750 83	1.9559	0.74001	60
2020-201	6 110.	35714 -	1541.90)83 176	2.6226	0.99999	938
2018-201	7 347.	96212 -	1102.21	02 179	8.1345	0.9903	727

2019-2017 -407.62121 -1857.7936 1042.5511 0.9781490 2020-2017 523.04545 -927.1269 1973.2178 0.9277426 2019-2018 -755.58333 -2173.8799 662.7132 0.6722154 2020-2018 175.08333 -1243.2132 1593.3799 0.9997677 2020-2019 930.66667 -487.6299 2348.9632 0.4293029

Year Electricitysd1 20144880.417728.59982 20156430.7502041.87033 20167407.1431080.83144 20176994.4551037.21935 20187342.417692.69836 20196586.833952.57997 20207517.500870.4768

D.5 Effluent TS

95% family-wise confidence level

Fit: aov(formula = Effluent_TS ~ Year, data = metadata)

	diff lw	r upr	p adj		
2015-2014	4 7997.682	-984.79	256 1698	0.156 0.11	22739
2016-2014	16843.182	8261.7	5241 2542	24.611 0.0	000023
2017-2014	4 21274.293	12034.1	2035 305	14.466 0.	0000000
2018-2014	4 9827.932	1246.50	241 1840	9.361 0.0	147911
2019-2014	10598.882	1616.4	0744 1958	81.356 0.0	108218
2020-2014	4 15052.896	5113.1	8934 2499	92.603 0.0	003813
2016-2015	5 8845.500	43.053	39 17647	.947 0.04	81143
2017-2015	5 13276.611	3830.82	2238 2272	22.400 0.0	012007
2018-2015	5 1830.250	-6972.19	661 1063	2.697 0.9	954403
2019-2015	5 2601.200	-6592.65	563 1179	5.056 0.9	769386
2020-2015	5 7055.214	-3075.92	2119 1718	6.350 0.3	538760
2017-2016	5 4431.111	-4634.15	5279 1349	6.375 0.7	508227
2018-2016	5 -7015.250	-15408.0	5353 137	7.554 0.1	611800

2019-2016 -6244.300 -15046.74661 2558.147 0.3318569 2020-2016 -1790.286 -11567.60594 7987.035 0.9977393 2018-2017 -11446.361 -20511.62502 -2381.097 0.0049887 2019-2017 -10675.411 -20121.19984 -1229.622 0.0169013 2020-2017 -6221.397 -16581.69843 4138.905 0.5346432 2019-2018 770.950 -8031.49661 9573.397 0.9999681 2020-2018 5224.964 -4552.35594 15002.285 0.6658345 2020-2019 4454.014 -5677.12119 14585.150 0.8312859

Year Effluent_TSsd1 201449084.82 5988.0192 201557082.50 8729.6593 201665928.00 7426.2694 201770359.11 6288.4405 201858912.75 4119.3016 201959683.70 7100.3037 202064137.71 7202.142

D.6 Effluent TN

Fit: aov(formula = Effluent_TN ~ Year, data = metadata)

\$Year

diff lwr upr p adj 2016-2014 497.666667 -844.8832 1840.2166 0.8769966 2017-2014 9.777778 -1349.8754 1369.4309 1.0000000 2018-2014 -560.000000 -1876.4785 756.4785 0.8008652 2019-2014 -698.033333 -2040.5832 644.5166 0.6354059 2020-2014 -350.333333 -1792.4633 1091.7966 0.9779748 2017-2016 -487.888889 -1424.9651 449.1874 0.6340520 2018-2016 -1057.666667 -1930.9197 -184.4136 0.0095733 2019-2016 -1195.700000 -2107.7830 -283.6170 0.0040739 2020-2016 -848.000000 -1901.1828 205.1828 0.1790566 2018-2017 -569.777778 -1469.1038 329.5482 0.4232430 2019-2017 -707.811111 -1644.8874 229.2651 0.2364522 2020-2017 -360.111111 -1435.0113 714.7891 0.9161162 2019-2018 -138.033333 -1011.2864 735.2197 0.9969255 2020-2018 209.666667 -810.0732 1229.4065 0.9895575 2020-2019 347.700000 -705.4828 1400.8828 0.9207828

Year Effluent_TN sd 1 2014 3645.333 258.6220 2 2016 4143.000 653.1854 3 2017 3655.111 577.1333 4 2018 3085.333 581.5000 5 2019 2947.300 876.2909 6 2020 3295.000 819.1642

D.7 Effluent TP

> ## Anova > fit7 <- aov(Effluent_TP~Year, metadata) > summary(fit7) Df Sum Sq Mean Sq F value Pr(>F) Year 6 4015385 669231 9.616 9.39e-07 *** Residuals 44 3062221 69596 ---Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1 > Tukey7 <- TukeyHSD(fit7, conf.level=0.95) #Tukey multiple comparison > Tukey7 #Output Tukey results Tukey multiple comparisons of means

95% family-wise confidence level

Fit: aov(formula = Effluent_TP ~ Year, data = metadata)

\$Year

diff lwr upr p adj 2015-2014 -352.333333 -1293.2144 588.54775 0.9062776 2016-2014 -628.033333 -1164.4181 -91.64859 0.0125302 2017-2014 -792.000000 -1335.2179 -248.78205 0.0009006 2018-2014 -724.250000 -1250.2185 -198.28149 0.0019504 2019-2014 -794.733333 -1331.1181 -258.34859 0.0007161 2020-2014 -27.000000 -603.1696 549.16964 0.9999991 2016-2015 -275.700000 -1130.2977 578.89768 0.9521294 2017-2015 -439.666667 -1298.5697 419.23632 0.6946584 2018-2015 - 371.916667 - 1220.0154 476.18208 0.8221838 2019-2015 -442.400000 -1296.9977 412.19768 0.6836316 2020-2015 325.333333 -554.7803 1205.44699 0.9113842 2017-2016 -163.966667 -538.3538 210.42047 0.8230582 2018-2016 -96.216667 -445.1047 252.67138 0.9776756 2019-2016 - 166.700000 - 531.1017 197.70168 0.7919737
2020-2016601.033333180.25851021.808140.00119702018-201767.750000-291.5549427.054900.99702252019-2017-2.733333-377.1205371.653801.00000002020-2017765.000000335.54851194.451490.00003562019-2018-70.483333-419.3714278.404710.99565222020-2018697.250000289.83651104.663460.00007262020-2019767.733333346.95851188.508140.0000228

Year Effluent_TPsd1 20141168.333351.983972 2015816.0000NA3 2016540.3000129.169184 2017376.333378.576405 2018444.0833325.769156 2019373.600077.729167 20201141.3333571.92715

*group with fewer than 2 data points will be deleted. 2015 only has 1 data point so it was removed

D.8 Effluent pH

Fit: aov(formula = Effluent_pH ~ Year, data = metadata)

\$Year

```
diff lwr upr p adj
2015-2014 -0.031818182 -0.2269049 0.163268578 0.9988202
2016-2014 -0.056818182 -0.2431948 0.129558443 0.9666104
2017-2014 -0.048484848 -0.2491685 0.152198755 0.9897882
2018-2014 -0.065151515 -0.2515281 0.121225110 0.9361855
2019-2014 -0.191818182 -0.3869049 0.003268578 0.0569543
2020-2014 -0.197532468 -0.4134090 0.018344039 0.0944586
2016-2015 -0.025000000 -0.2161768 0.166176809 0.9996669
2017-2015 -0.016666667 -0.2218160 0.188482632 0.9999795
```

 $\begin{array}{l} 2018-2015 & -0.033333333 & -0.2245101 & 0.157843476 & 0.9982809 \\ 2019-2015 & -0.160000000 & -0.3596777 & 0.039677665 & 0.1996821 \\ 2020-2015 & -0.165714286 & -0.3857484 & 0.054319784 & 0.2635888 \\ 2017-2016 & 0.008333333 & -0.1885515 & 0.205218166 & 0.9999996 \\ 2018-2016 & -0.008333333 & -0.1906133 & 0.173946602 & 0.9999993 \\ 2019-2016 & -0.135000000 & -0.3261768 & 0.056176809 & 0.3372597 \\ 2020-2016 & -0.140714286 & -0.3530640 & 0.071635413 & 0.4143088 \\ 2018-2017 & -0.0166666667 & -0.2135515 & 0.180218166 & 0.9999739 \\ 2019-2017 & -0.143333333 & -0.3484826 & 0.061815965 & 0.3499429 \\ 2020-2017 & -0.149047619 & -0.3740589 & 0.075963618 & 0.4147777 \\ 2019-2018 & -0.1266666667 & -0.3178435 & 0.064510142 & 0.4144769 \\ 2020-2018 & -0.132380952 & -0.3447307 & 0.079968746 & 0.4893366 \\ 2020-2019 & -0.005714286 & -0.2257484 & 0.214319784 & 1.0000000 \\ \end{array}$

Year Effluent_pH sd 1 2014 7.881818 0.1250454 2 2015 7.850000 0.1269296 3 2016 7.825000 0.1959824 4 2017 7.833333 0.1000000 5 2018 7.816667 0.1267304 6 2019 7.690000 0.1523884 7 2020 7.684286 0.1767161

D.9 Effluent VFA

> ## Anova > fit9 <- aov(Effluent_VFA~Year, metadata) > summary(fit9) Df Sum Sq Mean Sq F value Pr(>F) Year 6 66702800 11117133 7.911 2.26e-06 *** Residuals 63 88531222 1405257 ---Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1 > Tukey9 <- TukeyHSD(fit9, conf.level=0.95) #Tukey multiple comparison > Tukey9 #Output Tukey results Tukey multiple comparisons of means 95% family-wise confidence level Fit: aov(formula = Effluent_VFA ~ Year, data = metadata)

\$Year

diff lwr upr p adj 2015-2014 278.85455 -1298.6290 1856.3381 0.9981357 2016-2014 2483.37121 976.3183 3990.4241 0.0000896 2017-2014 532.89899 -1089.8410 2155.6390 0.9521232 2018-2014 -484.87879 -1991.9317 1022.1741 0.9565537 2019-2014 -114.54545 -1692.0290 1462.9381 0.9999894

2020-2014	76.78788 -1755.5436 1909.1193 0.9999996
2016-2015	2204.51667 658.6492 3750.3841 0.0009793
2017-2015	254.04444 -1404.8055 1912.8943 0.9991716
2018-2015	-763.73333 -2309.6008 782.1341 0.7409955
2019-2015	-393.40000 -2008.0059 1221.2059 0.9892838
2020-2015	-202.06667 -2066.4530 1662.3197 0.9998878
2017-2016	-1950.47222 -3542.4951 -358.4493 0.0071468
2018-2016	$-2968.25000\ -4442.1768\ -1494.3232\ 0.0000013$
2019-2016	$-2597.91667\ -4143.7841\ -1052.0492\ 0.0000621$
2020-2016	-2406.58333 -4211.7676 -601.3990 0.0025172
2018-2017	-1017.77778 -2609.8007 574.2451 0.4579787
2019-2017	-647.44444 -2306.2943 1011.4055 0.8958544
2020-2017	-456.11111 -2358.9424 1446.7202 0.9901690
2019-2018	370.33333 -1175.5341 1916.2008 0.9901989
2020-2018	561.66667 -1243.5176 2366.8510 0.9630106
2020-2019	191.33333 -1673.0530 2055.7197 0.9999185

Year Effluent_VFA sd 1 2014 1301.5455 918.1749 2 2015 1580.4000 1312.9674 3 2016 3784.9167 1859.0623 4 2017 1834.4444 1103.4844 5 2018 816.6667 597.9409 6 2019 1187.0000 1039.4021 7 2020 1378.3333 796.3772

E. Other SCAD laboratory results

E.1 Total Suspended Solids (TSS) Table E 1 Summary of TSS result

Table E.1 Summary of 188 results			
Year	Filtrate (mg/L)	Effluent (mg/L)	
2014	22,972	33,604	
2015		39,000	
2016	23,862	35,556	
2017	26,633	33,378	
2018	33,172	32,819	
2019	26,995	33,328	
2020	49,236	48,833	
Max	49,236	48,833	
Min	22,972	32,819	
Mean	29,384	36,302	
Average	30,478	36,645	
St. Dev	9,001	5,356	

E.2 Volatile Suspended Solids (VSS) Table E 2 Summary of VSS results

Table E.2 Summary of VSS results			
Year	Filtrate (mg/L)	Effluent (mg/L)	
2014	18,556	27,247	
2015		33,333	
2016	19,795	30,828	
2017	20,567	28,311	
2018	22,901	27,587	
2019	22,858	29,650	
2020	34,542	42,389	
Max	34,542	42,389	
Min	18,556	27,247	
Mean	22,694	30,995	
Average	23,203	31,335	
St. Dev	5,306	4,916	

E.3 Ammonia Table E.3 Summary of ammonia results

Year	Filtrate	Effluent
	(mg/L)	(mg/L)
2014	1,942	1,850
2015	1,825	1,566
2016	2,368	2,493
2017	1,917	2,029
2018	1,829	1,845
2019	1,875	1,755
2020	1,609	1,547
Max	2,368	2,493
Min	1,609	1,547
Mean	1,898	1,847
Average	1,909	1,869
St. Dev	213	299

E.4 Alkalinity

Year	Filtrate (mg/L)	Effluent (mg/L)
2014	1,942	1,850
2015	1,825	1,566
2016	2,368	2,493
2017	1,917	2,029
2018	1,829	1,845
2019	1,875	1,755
2020	1,609	1,547
Max	2,368	2,493
Min	1,609	1,547
Mean	1,898	1,847
Average	1,909	1,869
St. Dev	213	299

 Table E.4 Summary of alkalinity results

E.5 Total Dissolved Solids (TDS)

Table E.5 Summary of TDS results

Voor	Filtrate	Effluent
rear	(mg/L)	(mg/L)
2014	17,667	17,653
2015		20,317
2016	17,535	30,988
2017	14,070	34,948
2018	11,299	26,094
2019	13,515	26,374
2020	6,379	16,871
Max	17,667	34,948
Min	6,379	16,871
Mean	12,725	23,942
Average	13,411	24,749
St. Dev	3,862	6,318

F. A&L Great Lakes Laboratory Results

F.1 Solids

 Table F.1 Summary of Solids results

Year	Filtrate (%)	Effluent (%)	Solids (%)
2014			24.74
2015			39.35
2016			26.23
2017	3.87		
2018	5.24	3.87	27.48
2019			28.25
2020			31.25
Max	5.24	3.87	39.35
Min	3.87	3.87	24.74
Mean	4.50	3.87	29.20
Average	4.56	3.87	29.55
St. Dev	0.69	-	4.81

F.2 Total Kjeldahl Nitrogen (TKN) Table F.2 Summary of TKN results

Year	Filtrate (%)	Effluent (%)	Solids (%)
2014			0.43
2015			0.72
2016			0.63
2017	0.29		
2018	0.311	0.311	0.45
2019			0.55
2020			0.59
Max	0.31	0.31	0.72
Min	0.29	0.31	0.43
Mean	0.30	0.31	0.55
Average	0.30	0.31	0.56
St. Dev	0.01	-	0.10

F.3 Phosphorus (P) Table F.3 Summary of Phosphorus results

Year	Filtrate (%)	Effluent (%)	Solids (%)
2014			0.14
2015			0.35
2016			0.23
2017	0.03		
2018	0.037	0.033	0.147
2019			0.19
2020			0.18
Max	0.04	0.03	0.35
Min	0.03	0.03	0.14
Mean	0.03	0.03	0.20
Average	0.03	0.03	0.21
St. Dev	0.00	-	0.07

F.4 Potassium (K) Table F.4 Summary of Potassium results

Year	Filtrate (%)	Effluent (%)	Solids (%)
2014			0.26
2015			0.65
2016			0.31
2017	0.14		
2018	0.144	0.159	0.143
2019			0.22
2020			0.19
Max	0.14	0.16	0.65
Min	0.14	0.16	0.14
Mean	0.14	0.16	0.26
Average	0.14	0.16	0.30
St. Dev	0.00	-	0.17

F.5 Moisture

Year	Filtrate (%)	Effluent (%)
2017	96.13	
2018	94.76	0.159
Max	96.13	0.16
Min	94.76	0.16
Mean	95.44	0.16
Average	95.45	0.16
St. Dev	0.68	-

 Table F.5 Summary of Moisture results

F.6 Sulphur (S)

Table F.6 Summary of Sulphur results

Year	Filtrate (%)	Effluent (%)
2017	0.03	
2018	0.03	0.02
Max	0.03	0.02
Min	0.03	0.02
Mean	0.03	0.02
Average	0.03	0.02
St. Dev	-	-

F.7 Magnesium (Mg) Table F.7 Summary of Magnesium results

Year	Filtrate (%)	Effluent (%)		
2017	0.05			
2018	0.05	0.07		
Max	0.05	0.07		
Min	0.05	0.07		
Mean	0.05	0.07		
Average	0.05	0.07		
St. Dev	-	-		

F.8 Calcium (Ca)

Year	Filtrate (%)	Effluent (%)
2017	0.11	
2018	0.12	0.13
Max	0.12	0.13
Min	0.11	0.13
Mean	0.11	0.13
Average	0.12	0.13
St. Dev	0.01	-

Table F.8 Summary of Calcium results

F.9 Natrium (Na)

Table F.9 Summary of Natrium results

Year	Filtrate (%)	Effluent (%)		
2017	0.07			
2018	0.08	0.09		
Max	0.08	0.09		
Min	0.07	0.09		
Mean	0.07	0.09		
Average	0.08	0.09		
St. Dev	0.01	-		

F.10 Aluminum (Al) Table F.10 Summary of Aluminum results

Year	Filtrate (ppm)	Effluent (ppm)		
2017	48.5			
2018	44	67		
Max	48.50	67.00		
Min	44.00	67.00		
Mean	46.20	67.00		
Average	46.25	67.00		
St. Dev	2.25	-		

Year	Filtrate (ppm)	Effluent (ppm)		
2017	4.35			
2018	4.3	4.9		
Max	4.35	4.90		
Min	4.30	4.90		
Mean	4.32	4.90		
Average	4.33	4.90		
St. Dev	0.02	-		

F.11 Copper (Cu) Table F.11 Summary of Copper results

F.12 Iron (Fe)

Table F.12 Summary of Iron results

Year	Filtrate (ppm)	Effluent (ppm)		
2017	90.5			
2018	106	162		
Max	106.00	162.00		
Min	90.50	162.00		
Mean	97.94	162.00		
Average	98.25	162.00		
St. Dev	7.75	-		

F.13 Manganese (Mn)

Table F.13 Summary of Manganese results

Year	Filtrate (ppm)	Effluent (ppm)
2017	9.65	
2018	11	12
Max	11.00	12.00
Min	9.65	12.00
Mean	10.30	12.00
Average	10.33	12.00
St. Dev	0.68	-

F.14 Zinc (Zn)

Year	Filtrate (ppm)	Effluent (ppm)		
2017	61			
2018	53	57		
Max	61.00	57.00		
Min	53.00	57.00		
Mean	56.86	57.00		
Average	57.00	57.00		
St. Dev	4.00	-		

 Table F.14 Summary of Zinc results

F.15 Nitrogen (N)

Table F.15 Summary of Nitrogen results

Year	Solids (%)
2017	0.28
2018	0.45
Max	0.45
Min	0.28
Mean	0.35
Average	0.37
St. Dev	0.09

F.16 Ammonia (NH₃)

Table F.16 Summary of Ammonia results

Year	Solids (%)
2017	0.18
2018	0.14
Max	0.18
Min	0.14
Mean	0.16
Average	0.16
St. Dev	0.02

G. Formula for OLR and HRT Calculation

G.1 OLR Formula

OLR = IF (Total VS g=0, "", (Total VS g / (Working Volume Liters*Days)))

G.2 HRT Formula

HRT = IF (Volume of Digestate Wasted=0, "", ((Working Volume Gallons/Volume of Digestate

Wasted)) * Days)

H. R-Code of MLR for feedstock impact on biogas production (after data treated)

RegressionData1 <- read.table("C:/BIOGAS-AHAY/MLR_ORIGINAL DATA -REGRESSION.csv", header = TRUE, sep=',') validationData <- read.table("C:/BIOGAS-AHAY/MLR_ORIGINAL DATA -VALIDATION.csv", header = TRUE, sep=',') #BL2020a <- read.table("C:/MSU/Biogas data/BiogasProduction_REMOVED.csv", header = TRUE, sep=',') names(RegressionData1) = c('row.name','Year','Mon','Digestate','FMP','SLSS','DG','Parlor','Beef','WFM','Poultry','Swine','A NS','TMP','FFP','SLSSF','PA','Pulp','FOG','WFEED','Other','CartF','TFP','TFS','BiogasP','CH4P',' MethaneC','H2SC','Elec','TS','TN','TP','pH','VFA') summary(RegressionData1)

RegressionData <- RegressionData1 #BL2020[138,] RegressionData\$Year = as.factor(RegressionData\$Year) RegressionData\$Mon = as.factor(RegressionData\$Mon)

row.name <- RegressionData[1] Year <- RegressionData[2] MoN <- RegressionData[3] Digestate <- RegressionData[3] FMP <- RegressionData[5] SLSS <- RegressionData[5] DG <- RegressionData[6] DG <- RegressionData[7] Parlor <- RegressionData[8] Beef <- RegressionData[9] WFM <- RegressionData[10] Poultry <- RegressionData[11] Swine <- RegressionData[12] ANS <- RegressionData[13] TMP <- RegressionData[14] FFP <- RegressionData[15] SLSSF <- RegressionData[16] PA <- RegressionData[17] Pulp <- RegressionData[18] FOG <- RegressionData[19] WFEED <- RegressionData[20] Other <- RegressionData[21] CartF <- RegressionData[22] TFP <- RegressionData[23] TFS <- RegressionData[24] BiogasP <- RegressionData[25] CH4 <- RegressionData[26] MethaneC <- RegressionData[27] H2SC <- RegressionData[28] Elec <- RegressionData[29] TS <- RegressionData[30] TN <- RegressionData[31] TP <- RegressionData[32] pH <- RegressionData[33] VFA <- RegressionData[34]

library(tidyverse) library(GGally)

prints first set of rows showing works as a visual check if data were read correctly head(RegressionData)

#shows the data type and sample of the data str(RegressionData)

library(e1071) library(outliers)

#histograms of weights
hist(RegressionData\$FMP, breaks= 12, freq = F, xlab = 'FMP', main="")
hist(sqrt(RegressionData\$FMP), breaks= 12, freq = F, xlab = 'SQRT(FMP)', main="")

ggpairs(columns = c('FMP','DG','Parlor','Beef','TMP','Pulp','FOG','Other','CartF'), data = RegressionData, upper = list(continuous = wrap('cor', size = 8))) ggpairs(columns = c('FFP','Pulp','FOG','Other','CartF','TFP','TFS'), data = RegressionData, upper = list(continuous = wrap('cor', size = 8)))

ggpairs(columns = c('BiogasP','CH4P','MethaneC','H2SC','Elec','TS','TN','TP','pH','VFA'), data = RegressionData, upper = list(continuous = wrap('cor', size = 8)))

ggpairs(columns = c('BiogasP','TMP','TFP','FMP','MethaneC','H2SC','TS','TN','TP','pH','VFA'), data = RegressionData, upper = list(continuous = wrap('cor', size = 8)))

ggpairs(columns = c('BiogasP', 'sqrtFMP'), data = RegressionData, upper = list(continuous = wrap('cor', size = 8)))

ggpairs(columns = c('BiogasP', 'TFP'), data = RegressionData, upper = list(continuous = wrap('cor', size = 8)),aes(color = Mon, alpha = 0.5))

ggpairs(columns = c('BiogasP', 'TMP'), data = RegressionData, upper = list(continuous = wrap('cor', size = 8)),aes(color = Mon, alpha = 0.5))

ggpairs(columns = c('BiogasP', 'TFS'), data = RegressionData, upper = list(continuous = wrap('cor', size = 8)),aes(color = Mon, alpha = 0.5))

ggpairs(columns = c('BiogasP', 'TFS'), data = RegressionData, upper = list(continuous = wrap('cor', size = 8)),aes(color = Year, alpha = 0.5))

ggpairs(columns = c('BiogasP', 'TMP'), data = RegressionData, upper = list(continuous = wrap('cor', size = 8)),aes(color = Year, alpha = 0.5))

ggpairs(columns = c('BiogasP', 'TFP'), data = RegressionData, upper = list(continuous = wrap('cor', size = 8)),aes(color = Year, alpha = 0.5))

#Comparisons between the different Months
boxplot(RegressionData\$BiogasP ~RegressionData\$Mon, ylab ="Biogas Production", xlab
="Mon")
boxplot(RegressionData\$BiogasP ~RegressionData\$Year, ylab ="Biogas Production", xlab
="Year")

#Analysis of variance

#Conduct H0 : Mean Biogas production is the same for Months # Ha: at least one Month is different aov_Months <- aov(RegressionData\$BiogasP~RegressionData\$Mon) summary(aov_Months) #attributes(aov_species) aov_Months\$coefficients

TukeyHSD(aov_Months) plot(TukeyHSD(aov_Months), las=1)

#Full model for standardized data full_model_std <- lm(BiogasP ~ FMP + DG + PA + Beef + FFP + Pulp + FOG + Other + CartF , data=RegressionData_std)

modelDFFITS <- dffits(multiple.regression)</pre>

DFITTSThreshold <- sqrt(4*nparameters/ndata) # not a statistical test but to provide a general judgment

which(abs(modelDFFITS)>DFITTSThreshold)
RegressionData[index(RegressionData)==62,]

RegressionData_noinfluence <- RegressionData[-c(49,50,71,72),]

full_model <- lm(BiogasP ~ FMP + DG + Digestate + SLSS + SLSSF + WFM + Poultry + Parlor + Beef + FFP + Swine + ANS + PA + WFEED + Pulp + FOG + Other + CartF, data=RegressionData_noinfluence)

Bothfit.p <- ols_step_both_p(full_model, pent = .01, prem = .05, details = TRUE) Bothfit.p multiple.regression <- lm(BiogasP ~ FMP + DG + Parlor + FOG, data=RegressionData_noinfluence) summary(multiple.regression)

RESULT

	FMP	DG	Parlor	Beef	TMP	Pulp	FOG	Other	CartF	
3e-05 -	\sim	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	F),
1e-05 - 0e+00 -	~	-0.147	0.392***	0.141	0.703***	-0.115	-0.328**	-0.120	0.439***	P
15000 -	1 (c) 1	\wedge	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	0
10000 -		$ \ \ \ \ \ \ \ \ \ \ \ \ \$	-0.184.	-0.015	0.057	0.185.	0.322**	0.008	-0.101	ရ
40000 - 35000 - 30000 -	: .: . :		\wedge	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Par
25000 - 20000 - 15000 -	24		\sim	0.004	0.474***	-0.116	-0.256*	0.134	0.410***	for
10000 -		÷		\mathbf{n}	Corr:	Corr:	Corr:	Corr:	Corr:	Be
5000 - 0 -	1.2			×	0.163	0.179	-0.374***	-0.166	0.228*	ē
1e+05 - 8e+04 -						Corr:	Corr:	Corr:	Corr:	TH
6e+04 - 4e+04 -	64	1	19 m.	1	· _	-0.050	-0.272*	-0.204.	0.449***	P
3000 - 2000 -	ا فتعا		1. 1. 1. 1. 1. 1. 1. 1. 1. 1. 1. 1. 1. 1. 1		S. 1. 14	\wedge	Corr:	Corr:	Corr:	PL
1000 - 0 -	S. C		298 g			\sim \setminus	-0.062	-0.539***	-0.176	þ
80000 - 60000 -	•		•	•	•		\wedge	Corr:	Corr:	F
40000 - 20000 -	1	. X		1 4	24 . A.	Sec.		0.199.	-0.211.	õ
25000 - 20000 - 15000 -	· · · ·		* .	· · · ·		ı ·	1.0	1	Corr:	Off
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4000 - 3000 -			:	. · · · ·			.j., "	۴. ۱	\backslash	Ca
1000 -			· · · · ·	<u>.</u>	1. 	· · · · · · · · · · · ·	·	2. 		÷.
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Figure H.1 Correlation plot between the parameters part 1 (Note: FMP = filtrate manure pit, DG = Dairy Gutter, TMP = Total Manure Pit, CartF = Cart Food)



Figure H.2 Correlation plot between the parameters part 2 (Note: FFP = Filtrate Food Pit, TFP = Total Food Pit, TFS = Total Feedstock)

	BiogasP	CH4P	MethaneC	H2SC	Elec	TS	TN	TP	pH	VEA	
6e-04 - 4e-04 -	\wedge	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Biog
2e-04 -	\sim \	0.984***	0.078	-0.054	0.443***	0.402***	0.101	-0.156	-0.052	0.101	Jasp
3000 - 2000 -		\wedge	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	£
1000 -	a service and a service of the servi	\searrow	0.248*	-0.093	0.482***	0.406***	-0.027	-0.116	-0.061	0.067	I4P
67.5 - 65.0 -	7. đ.	, <u>a</u> ≩.	\wedge	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Meth
62.5 - 60.0 - • 57.5 -	···		\sim (-0.249*	0.295**	0.022	-0.480***	0.101	-0.056	-0.195	aneC
2000 - 1500 -	::.	· · · ·		\wedge	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	H2
1000 - 500 - 0 -			. .	~	-0.322**	0.322**	0.295*	0.048	0.045	0.278*	SC
10000 - 7500 -	in the second	من المراجعة	Senne:	1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1	\wedge	Corr:	Corr:	Corr:	Corr:	Corr:	
2500 -•					\checkmark	0.004	-0.370**	0.005	-0.076	-0.189	ec
80000 - 70000 -•	· 24			Ser.		\wedge	Corr:	Corr:	Corr:	Corr:	E.
60000 - 50000 - 40000 -		1.198		2		/	0.214	-0.049	-0.096	0.318**	<i>co</i>
5000 - 4000 -		1.1	1.1	Sec	1.30		\wedge	Corr:	Corr:	Corr:	Ξ
3000 - 2000 -					8.4		\nearrow	0.240.	0.119	0.419**	2
2000 - 1500 -							••••	Λ	Corr:	Corr:	-
1000 - 500 -			31 1 in 11	in.		and the second	· · · · · ·	The	0.132	-0.033	q
8.00 - • 7.75 - 7.50 -	• •	• • • • • • • • • • • • • • • • • • • •	•••••		•••••••••••••••••••••••••••••••••••••••	·· ·	• • • • • • • •		(Corr:	p
7.25 - 7.00 -	-							-		-0.026	1
6000 - 4000 -			• •	÷., .			•			\wedge	<pre><f< pre=""></f<></pre>
2000 - 0 - 1					i			ð 🐂 👘	: !		

Figure H.3 Correlation plot between the operational parameters (Note: BiogasP = Biogas Production, CH4P = Methane production, MethaneC = Methane Concentration, H2SC = hydrogen sulfide concentration, Elec = Electricity)

	BiogasP	TMP	TFP	FMP	MethaneC	H2SC	TS	TN	TP	pH	VFA	
6e-04 - 4e-04 -	\wedge	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Bioga
2e-04 - 0e+00	/ \	0.372***	-0.058	0 478***	0 078	-0 054	ገ 4በ2***	0 101	-0 156	-0.052	0 101	lsb
1e+05 - 8e+04 -	1.5.		Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	TN.
6e+04 - 4e+04 -			0 101	0 703***	-0 237*	-0 047	-0 085	0 313*	0 1 1 9	0 167	0 141	9
80000 -*	10	1. 64.	\wedge	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Ħ
40000 - 20000 -	31	200	$/ \setminus$	0 159	-0 210	0 231*	-0 023	0 2 1 4	0 153	0.032	0 237*	ų.
40000 - 30000 -			·	\sim	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	FA
10000 -		11	1.76	~	-0.324**	-0 088	-0.376**	0.318*	0 242	0 125	-0.039	P
67.5 - 65.0 -	Sec.	£77	··	1.12.11	\wedge	Corr:	Corr:	Corr:	Corr:	Corr:	Corr:	Meth
62.5 - 60.0 - •	···			· · · ·	\sim \langle	-0 249*	0 022	0 480**	0 101	-0 056	-0 195	aneC
2000 - 1500 -		•	• :	: •		\wedge	Corr:	Corr:	Corr:	Corr:	Corr:	H2
1000 - • 500 -				33.5.		-	0 322**	0 295*	0 048	0 045	0 278*	SC
80000 - 70000 -•	· *	Victory of	ANY .	1.		25.00	$ \land$	Corr:	Corr:	Corr:	Corr:	-
60000 - 50000 - 40000 -				1.		4	/	0 2 1 4	-0 049	-0.096	0.318**	r
5000 - 4000 -			130	102.	1.41.e	ine .		\wedge	Corr:	Corr:	Corr:	-
3000 - 2000 -							100	\nearrow	0 240	0 1 1 9	0 419**	r
2000 - 1500 -			• •						Λ	Corr:	Corr:	-
1000 - 500 -	يتقور ا	mine .			Sec. 14.14	time .		· · · · · · · ·	ha	0 132	-0.033	σ
8.00 - • 7.75 -				••••	•••••		·· ·	••••••		[Corr:	
7.50 - 7.25 - 7.00 -											-0 026	Ĩ
6000-			:	· . ·		÷., .					\wedge	<
2000 -			with .	1 shinis		14		: Altor	3	:!		¢,A
10	0@00@000#000	4e+04e+08e+04e-2	COORDOCHDOCOD		10.550.062.565.067.5	0 5001000502000	000000000000000000000000000000000000000	2000300040005000	500100015002000	7.007.257.507.758.0	00 200040005000	

Figure H.4 Correlation plot between the total feedstock and the operational parameters



Figure H.5 Correlation plot between the total manure pit and the biogas production



Figure H.6 Correlation plot between the total food pit and the biogas production



Figure H.7 Correlation plot between the total feedstock and the biogas production



Figure H.8 The data distribution of the biogas production from 2014 to 2020



Figure H.9 Linear model plot of biogas production (residuals vs fitted)

I. R-Code of MLR for operational parameters impact on biogas production

Load libraries ----library(dplyr)
library(FSA)
library(psych)
library(car)
library(rcompanion)

Choose data file of "MLR_operational_parameters.txt" ----con <-file.choose(new = FALSE) metadata <- read.table(con, header = T, row.names = 1, fill = TRUE) head(metadata)

corr.test(data.num, use = "pairwise", method="pearson", adjust="none", alpha="0.05")

```
library(psych)
```

```
pairs.panels(data.num,
    method = "pearson", # correlation method
    hist.col = "#00AFBB",
    density = TRUE, # show density plots
    ellipses = TRUE # show correlation ellipses
)
```

```
summary(data.num$Methane)
plot(data.num$Methane)
hist(data.num$Methane)
```

```
model1<-lm(Methane~OLR+HRT+Temperature+pH,data=data.num)
summary(model1)</pre>
```

```
model2<-lm(Biogas~OLR+HRT+Temperature+pH,data=data.num)
summary(model2)</pre>
```





Figure I.1 Correlation plot between the operational parameters and the dependent variables (biogas production and methane concentration)



Figure I.2 Scatter plots of variables (Note: green = regression line, red = non-parametric mean, blue = non-parametric variance)

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